



April 3, 2014

Janet Hashimoto  
U.S. Environmental Protection Agency  
Region IX  
Monitoring and Assessment Office  
75 Hawthorne Street, 11<sup>th</sup> Floor  
San Francisco, CA 94105

David Smith  
U.S. Environmental Protection Agency  
Region IX  
Clean Water Standards and Permit Office  
75 Hawthorne Street, 11<sup>th</sup> Floor  
San Francisco, CA 94105

Bruce Wolfe, Executive Officer  
California Regional Water Quality Control Board  
San Francisco Bay Region  
1515 Clay Street, Suite 1400  
Oakland, CA 94612

**Re: 1997-2012 Southwest Ocean Outfall Regional Monitoring  
Program Summary Report**

Dear Ms. Hashimoto, Mr. Smith and Mr. Wolfe:

Enclosed please find a copy of the City and County of San Francisco Public Utilities Commission Southwest Ocean Outfall (SWOO) Regional Monitoring Program summary report for the period 1997 through 2012, as required in the Oceanside Water Pollution Control Plant NPDES Permit No. CA0037681. In lieu of preparing a separate annual report for the 2012 offshore and beach water quality monitoring data, they are included in the comprehensive analysis of the sixteen-year data set.

This report explores shoreline bacteria data and relationships to San Francisco's combined sewer system and rainfall. The offshore data collected over this period provide a regional context for evaluating SWOO impacts and continue to demonstrate that conditions at the outfall and reference sites are essentially the same. Differences that did occur between outfall and references stations appear to be haphazard and transitory. None of the indicators measured (sediment quality including priority pollutants, biological community structure, bioaccumulation of priority pollutants in crab tissue) provide an indication of persistent or long-term impacts from the wastewater discharge.

Other important findings in this report include an overall increase in sediment fines (silt and clay) in the study area beginning in 2007 that corresponds to a reduction of suspended sediment loads within the San Francisco Estuary system and a contraction of the ebb-tide delta at the mouth of the Bay as documented by the U.S. Geological Survey. In addition, the fisheries section of the report presents support for our request to drop the trawling requirements from the offshore monitoring program. A summary of this request and its support have been included in the ROWD submitted with the permit renewal application.

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Mayor

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Commissioner

Harlan L. Kelly, Jr.  
General Manager



If you have any questions regarding information contained in this report, please contact Michael Kellogg, Supervising Biologist, Oceanside WPCP, at 415.242.2218.

Sincerely,

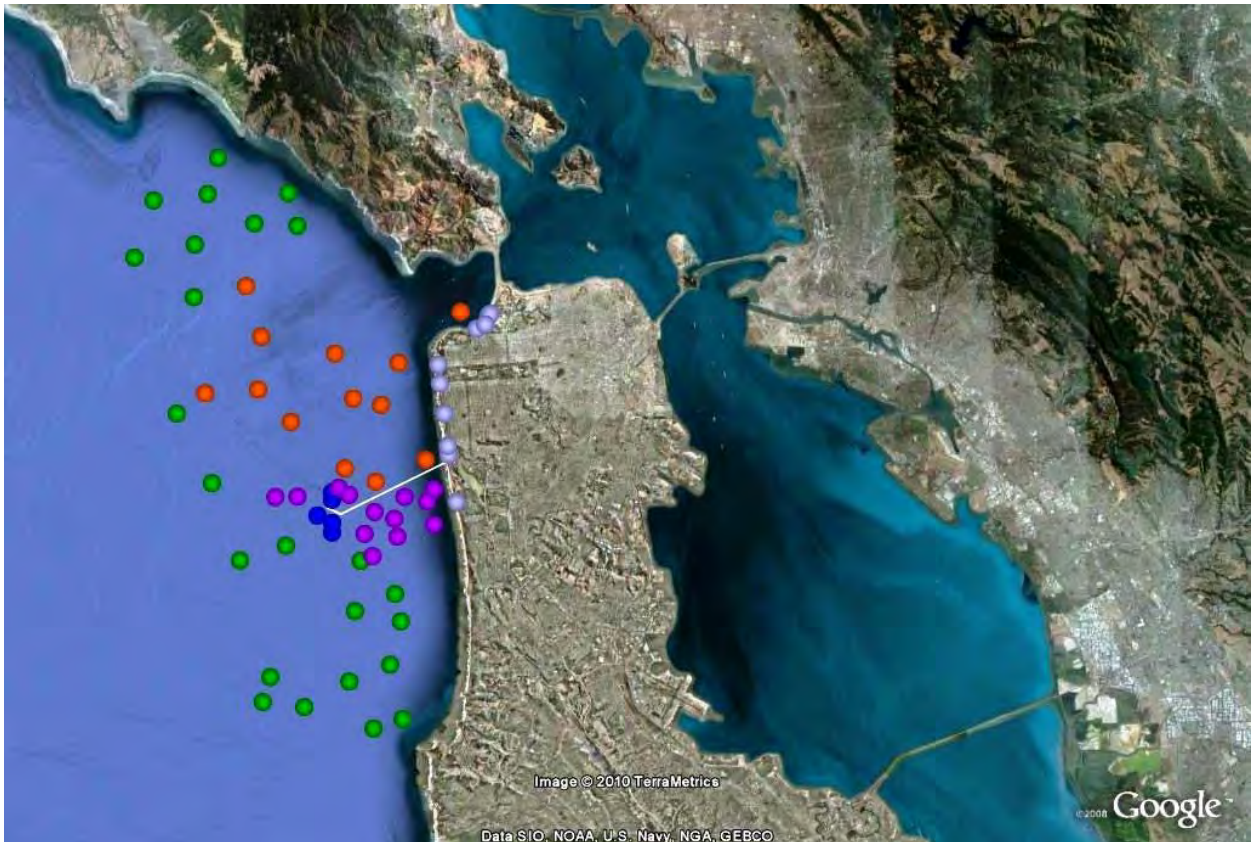


Tommy T. Moala  
Assistant General Manager  
SFPUC Wastewater Department

cc     Robyn Stuber  
       Derek Whitworth

# Southwest Ocean Outfall Regional Monitoring Program

## Sixteen-Year Summary Report 1997 – 2012



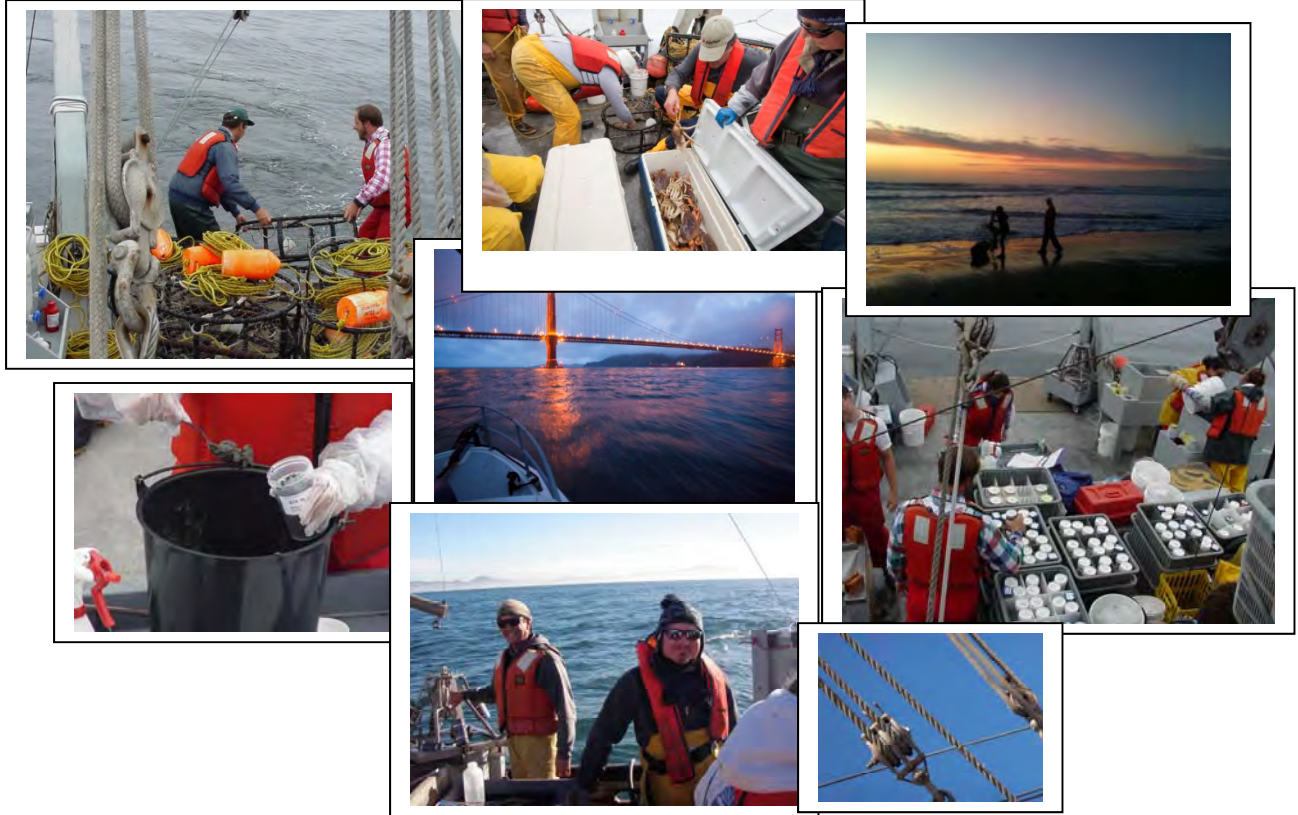
submitted to  
U.S. Environmental Protection Agency, Region 9  
and  
San Francisco Bay Regional Water Quality Control Board

for  
San Francisco Public Utilities Commission, Waste Water Enterprise  
Oceanside Water Pollution Control Plant NPDES Permit CA0037681

by  
SFPUC Natural Resources and Lands Management Division  
Oceanside Biology Laboratory  
April 2014

# SOUTHWEST OCEAN OUTFALL REGIONAL MONITORING PROGRAM

## SIXTEEN-YEAR SUMMARY REPORT 1997 – 2012



Submitted to  
U.S. Environmental Protection Agency, Region 9  
and  
California Regional Water Quality Control Board, San Francisco Bay Region

For  
San Francisco Public Utilities Commission, Waste Water Enterprise  
Oceanside Water Pollution Control Plant NPDES Permit CA0037681

By  
San Francisco Public Utilities Commission  
Natural Resources and Lands Management Division  
OCEANSIDE BIOLOGY LABORATORY

April 2014



# EXECUTIVE SUMMARY

This report summarizes and explores trends in sixteen years of environmental monitoring data collected from 1997 through 2012 as part of the Southwest Ocean Outfall Regional Monitoring Program. The monitoring is designed to detect environmental impacts related to the discharge of treated combined sewer effluent from the Oceanside Water Pollution Control Plant (WPCP) and associated Westside Wet Weather Facilities owned and operated by the City and County of San Francisco. The combined sewer system collects and treats sanitary flow, industrial effluent, and storm water. All dry weather flows (average 14 MGD) and wet weather flows up to 43 MGD receive secondary treatment. Wet weather flows above 43 MGD receive primary treatment. Flows up to 175 MGD are discharged approximately 3.75 miles offshore in the Pacific Ocean through the Southwest Ocean Outfall (SWOO) and flows in excess of 175 MGD result in combined sewer discharges into shoreline waters including some recreational beaches. All discharges to the environment have received treatment at least equivalent to wet weather primary effluent. The facilities and discharges are regulated under the National Pollution Discharge Elimination System (NPDES) provisions of the Clean Water Act through a permit jointly administered by the U.S. Environmental Protection Agency, Region IX and the State Regional Water Quality Control Board, San Francisco Bay Region. The Oceanside NPDES permit mandates extensive monitoring to assess compliance with broad goals of the Clean Water Act (maintain fishable and swimmable waters) and the California Ocean Plan (prevent degradation of beneficial uses).

Specific details of the monitoring requirements have varied over the sixteen-year period, but have always included two main components:

- The Beach Monitoring Program involves measurements of bacteria concentrations at recreational beaches and notification to the public when State standards are exceeded or when a combined sewer discharge occurs.
- The Offshore Monitoring Program involves collection and analysis of physical, chemical, and biological parameters in order to assess and compare the outfall region, where potential impacts may be expected, with reference conditions utilizing:
  - 1) sediment quality (physical and chemical)
  - 2) benthic infauna community structure
  - 3) demersal fish and epibenthic invertebrate community structure
  - 4) physical anomalies and bioaccumulation of contaminants in organism tissues

## MONITORING INDICATORS

### Beneficial Uses

Water contact and non-water contact recreation at San Francisco beaches is an important beneficial use by thousands of local Bay Area residents and tourists annually. Beach water quality is generally very good at Baker Beach, China Beach, and Ocean Beach on the City's north and west shores, especially during dry weather. Bacteria concentrations (indicators of impaired water quality) that exceed State standards for water contact recreation are most frequently associated with wet weather, either because of treated combined sewer discharges or for unknown causes. Treated combined sewer discharges continue to show a strong relationship with rainfall: years with greater rainfall usually have more discharges, but the intensity of storms is the main determining factor. There has been a dramatic reduction in

treated combined sewer discharges since the completion of the westside infrastructure improvements in 1997, demonstrating the efficacy of the combined sewer system controls. The long-term design goal of eight or fewer discharges per year has been met during this study. The switch from a single (total coliform) to three bacteria indicators (total coliform, *Escherichia coli*, and enterococcus) in October 2003 has resulted in an increased frequency of beach posting and, presumably, in greater protection of public health by use of these more sensitive indicators. Implementation of confirmation posting in July 2007 has resulted in fewer postings at Ocean Beach. Lobos Creek continues to be a source of bacteria at Baker Beach, which had more posted days than the other beaches. Overall, San Francisco west side beaches were available for water contact recreation 94% or more of the time during the nine years that the three indicators have been used (2003 – 2012).

Recreational Use observations were made after combined sewer discharges to assess their effect on beach use. Because discharges most often occur during winter storms associated with shorter days and unpleasant weather conditions, few beach users are affected. However, isolated discharge events that occur in early Fall or Spring potentially impact more users since recreational use increases when days are longer and the duration of storm events is typically shorter and may contribute to good surf conditions.

### Sediment Quality

Sediment organic content (e.g., TOC, TVS, TKN) often increases at wastewater discharges. However, a BACIP analysis comparing samples from before and after the onset of effluent discharge from an impacted and control site demonstrates that the differences in means of those constituents at a reference station and an outfall station are not significantly different now than they were before the discharge began. The BACIP analysis did show a significant difference for sediment fines at outfall Station 01, which also often increase at wastewater discharges, however, examination of the data revealed that this is due to a reduction of sediment fines at the outfall station, not an increase. Mean sediment grain size has remained similar at the outfall compared to pre-discharge values.

### Community Analyses

Benthic infauna, demersal fish, and epibenthic invertebrate communities sampled in the study area represent a general assortment of native species common in sandy offshore environments in central California. Multivariate analyses demonstrate that communities at the outfall do not differ from communities in reference areas. Rarely, one or a few introduced invertebrates common in the San Francisco Estuary have appeared in benthic samples. It is perhaps surprising that exotic species are not more common offshore given that the San Francisco Estuary is considered the most invaded aquatic ecosystem in North America.

#### *Benthic Infauna*

Reference envelope analysis shows that benthic infauna indicators (abundance, species richness, diversity, evenness) at outfall stations are the same as at reference stations. Occasional excursions from reference conditions that occurred at outfall stations were generally matched by similar excursions at reference stations in the same years. Abundance has appeared high at the outfall in some years, but a BACIP analysis comparing samples from before and after the onset of discharge from an impacted and control site demonstrates that the differences in means of infauna abundance at a reference station and an outfall station are not

significantly different now than they were before the discharge began. In addition, none of the species with high abundance at outfall stations are known to be pollution tolerant or indicators of enrichment. Sporadic, occasional high abundance at stations throughout the study area appears to be due to haphazard recruitment events. Benthic infauna are currently in a high abundance cycle, particularly the polychaete worm *Spiophanes norrisi*. Total infauna abundance has been higher in the past three years than any of the previous 13 years. Cluster and ordination analyses demonstrate that, based upon abundance and species composition, benthic infauna communities at outfall and reference stations are not different. Subtle differences that do occur are of short duration.

#### *Demersal Fish and Epibenthic Invertebrates*

Under the adaptive management provisions of the NPDES permit Monitoring and Reporting Program, trawl sampling was curtailed in 2009 due to the listing of longfin smelt (*Spirinchus thaleichthys*) as a threatened species by the California Department of Fish and Wildlife. Longfin smelt were commonly caught as by-catch during SWOO monitoring program trawl sampling. In light of the information gathered through trawl sampling over two decades, we argue to drop the trawl requirement from the permit because: 1) the trawl sampling has not revealed a significant difference between outfall and reference area demersal fish and epibenthic invertebrate communities; 2) trawl sampling is not suited to finding an outfall effect; 3) the demersal fish caught are not representative of contaminant exposure to consumers of local fishes or of body burdens obtained within the Gulf of the Farallones; 4) trawl sampling results in significant mortality to demersal fish and epibenthic organisms including listed species; 5) trawl sampling destroys benthic habitat; 6) other new sources of high-quality data are available; and 7) given the absence of an outfall effect, the trawl program is expensive and burdensome to implement.

#### Priority Pollutant Analyses

Regulatory guidelines do not exist for pollutant concentrations in sediment or organisms for the offshore San Francisco region.

#### *Sediment Organic Pollutants*

DDT and derivative compounds (organochlorine pesticides) and PCB congeners (polychlorinated biphenyls) are infrequently detected and occur at low concentrations within the study area. PAH compounds (polycyclic aromatic hydrocarbons) are detected annually, but also at generally low concentrations. Reference envelope analysis showed that outfall Station 58 was above reference conditions for total PAHs in seven of the 16 years and near the upper tolerance interval bound in other years. Station 58 also had the highest percentage of sediment fines (silt and clay) ever measured in the study area. Reference Envelope exceedances at Station 58 have generally been matched by similar exceedances at northern reference stations that also have high percentages of silt and clay.

#### *Sediment Inorganic Pollutants – Trace Metals*

Reference envelope analysis demonstrated that sediment metals concentrations at outfall and reference stations do not differ. Within the SWOO study area, arsenic, chromium, cadmium, mercury, and nickel generally have higher concentrations than other metals

measured. Some trace metals occur naturally in the environment. For example, nickel may be elevated in the region due to natural geologic sources such as serpentine soils.

#### *Bioaccumulation – Pollutants in Tissues*

Organic pollutants and trace metals were found in varying levels and tended to accumulate in higher concentrations in crab hepatopancreas tissues than in muscle tissue. Public awareness and education may be appropriate to inform people that these fatty tissues may not be suitable for consumption. None of the regressions involving sediment and tissue concentrations were significant. There appears to be a trend of decreasing PCB-levels in hepatopancreas tissue from both outfall and reference areas, but those compounds are generally detected near or below detection limits and conclusions about them should be made cautiously.

The bioaccumulated pollutants and abnormalities (e.g. tumors and lesions) found in Dungeness crab from the SWOO study area may have their source in contaminated sediments and organisms from the San Francisco Estuary. Dungeness crab utilize estuarine environments during their juvenile stages. Furthermore, they are mobile predators that can range substantially both latitudinally and along inshore-offshore gradients. Thus, the bioaccumulation results reported herein are not relevant to determining an outfall effect because the origin of body burdens cannot be determined. However, the data do provide information of potential interest to important commercial and sport fisheries and for assessing public health risk.

#### STUDY AREA OVERVIEW

Sixteen years of monitoring data allow some characterization of the SWOO study area. The sedimentary environment appears to be dominated by input from the Sacramento and San Joaquin River system through the San Francisco Estuary, and by reworking from tidal currents and wave action. Sediment-laden currents funnel through the Golden Gate on ebb tides and fan out, depositing sediments along the transport path. The strong tidal currents have formed an ebb tide delta of sandbars that surround the mouth of the San Francisco Estuary. Sediment at stations surrounded by these sandbars has been predominantly medium and coarse sands. The sandbars of the ebb tide delta are predominantly fine and medium sands and are well sorted. Seaward of the sand bars are areas of fine to very fine sands with the highest average percentages of silt and clay occurring in a band just seaward of the sandbars.

Each region of grain size has a distinct benthic infauna community. The community in the coarse grain sediment surrounded by the sandbars has been numerically dominated by two small, interstitial-like polychaetes, *Hesionura coineaui difficilis* and *Heteropodarke heteromorpha*, nematodes, and the bivalve *Tellina nukuloides*. The community associated with the well-sorted fine sands of the sandbars has been numerically dominated by the polychaete *Spiophanes norrisi* and characterized by a higher percentage of Crustacea than the other infauna communities. The benthic infauna community of the very fine sands characterizing the outfall and reference regions has been numerically dominated early in the study period by the polychaetes *Spiophanes berkeleyorum* and the bivalve *Tellina modesta*, but in later years by the polychaete *S. norrisi* and the bivalve *Mactromeris catilliformis*.

Smaller sediment grains provide greater relative surface area for adsorption of contaminants and organic matter, therefore areas seaward of the sandbars with higher percentages of silt and clay might be expected to have higher contaminant concentrations.



Such a pattern of grain size distribution, sediment chemistry measures of TOC, TVS, TKN, and metals concentrations, has been observed. Thus, the location of the SWOO (just seaward of a sandbar) places it in an environment where elevated measures of sediment fines, organic matter, and contaminants might be expected even in the absence of a wastewater discharge. It is important to evaluate potential discharge impacts by comparing similar environments.

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# ACRONYMS AND ABBREVIATIONS

AAS	atomic absorption spectroscopy
AB	California State Assembly Bill
Ag	silver
Al	aluminum
ANOVA	analysis of variance
As	arsenic
ASTM	American Society for Testing and Materials
ATSDR	Agency for Toxic Substances and Disease Registry
BACIP	Before-After-Control-Impact-Paired statistical test
BDL	Below Detection Limit
BND	black necrotic disease
BOD	biochemical oxygen demand
BW	body weight
BWPC	Bureau of Water Pollution Control
°C	degree Celsius
CalEPA	California Environmental Protection Agency
Cd	cadmium
CCSF	City and County of San Francisco
CDFG	California Department of Fish and Game
CDPH	California Department of Public Health
CDOI	Corrected Delta Outflow Index
CEDEN	California Environmental Data Exchange Network
CFCP	Coastal Fish Contamination Program
CFU	colony forming unit
cm	centimeter
COP	California Ocean Plan
Cr	chromium
CR	consumption rate
CSDOC	County Sanitation Districts of Orange County
CSS	Combined Sewer System
CSD	Combined Sewer Discharge
Cu	copper
cv	coefficient of variation
CVAAS	cold vapor atomic absorption spectroscopy
CWA	Clean Water Act
DDD	dichlorodiphenyldichloroethane
DDE	dichlorodiphenyldichloroethylene
DDT	dichlorodiphenyltrichloroethane
df	degrees of freedom
DHS	California Department of Health Services
DL	Detection Limit
DMA	dimethyl arsenic
DO	dissolved oxygen
DPW	San Francisco Department of Public Works
dry wt	dry weight
DW	disk width
EAP	Ecological Analysis Package

## ACRONYMS AND ABBREVIATIONS (cont.)

EMAP	Environmental Monitoring Assessment Program
ERL	effects range low
ERM	effects range median
ERMq	effects range median quotient
FDR	false discovery rate
Fe	iron
g	gram
g/day	grams per day
GC/MS	gas chromatography/mass spectrometry
GF-AAS	graphite furnace atomic absorption spectroscopy
GFNMS	Gulf of the Farallones National Marine Sanctuary
GGNRA	Golden Gate National Recreation Area
H'	Shannon-Weiner Diversity Index
Hg	mercury
ICP-AES	inductively coupled plasma atomic emission spectroscopy
IRIS	Integrated Risk Information System
J'	Pielou's Evenness Index
kg	kilogram
km	kilometer
L	liter
LACSD	Los Angeles County Sanitation Districts
LM	Lake Merced discharge structure
L/V	Lincoln/Vicente discharge structures
m	meter
m <sup>2</sup>	square meter
MANOVA	Multivariate Analysis of Variance
MBNMS	Monterey Bay National Marine Sanctuary
MDL	method detection limit
METC	Marine Estuarine Technical Committee
mg	milligram
MgCl	magnesium chloride
MGD	million gallons per day
mg/Kg	milligrams per kilogram
mg/L	milligrams per liter
mL	milliliter
mm	millimeter
MMA	monomethyl arsenic
Mn	manganese
MPN	most probable number
MWH	MWH laboratory referenced in method/metals
n/a	not applicable or not assessed
NA	not applicable or not assessed
ND	no data or not detected
NMDS	Nin-Metric Multidimensional Scaling
NSMCSD	North San Mateo County Sanitation District



## ACRONYMS AND ABBREVIATIONS (cont.)

Ni	nickel
NOAA	National Oceanic and Atmospheric Administration
NPDES	National Pollutant Discharge Elimination System
NPGO	Northern Pacific Gyre Oscillation Index
NPS	National Park Service
NRC	National Resource Council
NRD	Natural Resources Division of the SFPUC
NRLMD	Natural Resources & Lands Management Division
NS	no sample
NWS	National Weather Service
p	probability
PAH	polycyclic aromatic hydrocarbon
Pb	lead
PCA	principal components analysis
PCB	polychlorinated biphenyl
PCR	polymerase chain reaction
PFE	pressurized fluid extraction
ppb	parts per billion
ppm	parts per million
ppt	parts per thousand
PSMFC	Pacific States Marine Fisheries Commission
QA	quality assurance
QC	quality control
RfD	oral reference dose
RL	reporting limit or risk level
RMP	Regional Monitoring Program for Trace Substances
RWQCB	Regional Water Quality Control Board
SC I	Sea Cliff I
SC II	Sea Cliff II
SCAMIT	Southern California Association of Marine Invertebrate Taxonomists
SCCWRP	Southern California Coastal Water Research Project
SD	standard deviation
Se	selenium
SF	slope factor
SFDPH	San Francisco Department of Public Health
SFEI	San Francisco Estuary Institute
SFPUC	San Francisco Public Utilities Commission
SIM	selected ion monitoring
SIMPER	similarity percentage analysis
SL	standard length
sp	species (singular)
spp	species (plural)
SQG	sediment quality guidelines
SSM	single-sample maximum
SV	screening value
SVc	screening value for a carcinogen

## ACRONYMS AND ABBREVIATIONS (cont.)

SV <sub>n</sub>	screening value for a non-carcinogen
SWRCB	State Water Resources Control Board
SWOO	Southwest Ocean Outfall
S-W	Shannan-Weiner Index
TKN	total Kjeldahl nitrogen
TL	total length
TOC	total organic carbon
TS	total solids
TSS	total suspended solids
TVS	total volatile solids
USDOE	United States Department of Energy
U.S. EPA	United States Environmental Protection Agency
USGS	United States Geological Survey
WPCP	Water Pollution Control Plant
WQB	Water Quality Bureau
WRCC	Western Regional Climate Center
wt	weight
ZID	zone of initial dilution
Zn	zinc
$\alpha$	probability of Type I error
$\mu\text{g}$	microgram
$\Sigma$	sum of
$\phi$	phi (= $-\log_2(\text{particle diameter (mm)})$ )



**SECTION 1**  
**INTRODUCTION**

# INTRODUCTION

## 1.1. FACILITIES AND REGULATORY AUTHORITY

The Oceanside Water Pollution Control Plant (WPCP) and Westside Wet Weather Facilities are owned and operated by the City and County of San Francisco (City), Public Utilities Commission. These facilities collect, treat, and discharge wastewater and stormwater from the City's western drainage into the Pacific Ocean through the Southwest Ocean Outfall (SWOO) approximately 3.75 miles offshore. San Francisco has a combined sewer system that collects domestic sanitary flow, industrial wastewater, and stormwater runoff in the same set of pipes and conveys these combined flows to treatment facilities. Because of the combined sewer system, flow through SWOO varies from an average of 14 MGD during dry weather to a peak of 175 MGD during wet weather. The Oceanside WPCP provides secondary treatment for all dry weather flows and wet weather flows up to 43 MGD and primary treatment for wet weather flows above 43 to 65 MGD. Flows in excess of 65 MGD receive flow-through treatment equivalent to wet weather primary effluent within the Westside Wet Weather Facilities and are discharged through SWOO along with the blended secondary and primary effluents from the Oceanside WPCP. Flows exceeding the maximum capacity of SWOO (175 MGD) also receive flow-through treatment, but are discharged at several locations along the shoreline. All discharges to the environment have received treatment. More details and a history of the facilities are presented in Appendix A.

These facilities and discharges are subject to regulation under the Clean Water Act through the National Pollution Discharge Elimination System (NPDES) program. Because the ocean outfall is located beyond the California territorial limit of 3 miles, regulatory authority is jointly administered by the U.S. Environmental

Protection Agency (U.S. EPA), Region 9, and the California Regional Water Quality Control Board, San Francisco Bay Region (RWQCB). The Oceanside WPCP NPDES permit (no. CA0037681) has included extensive environmental monitoring requirements (U.S. EPA and RWQCB 1997, 2003). This report provides a summary of environmental monitoring data collected over the eight-year period from 1997 to 2004 and an analysis and discussion of time related trends.

## 1.2. SOUTHWEST OCEAN OUTFALL REGIONAL MONITORING PROGRAM

The monitoring program adopted a regional perspective in 1997 (Figure 1-1). Many studies had been conducted since the late 1970s to determine impacts of the design, construction, and operation of the SWOO. Pre-design studies included investigations of biota, water quality, water circulation, and plume behavior to determine the optimum placement of the SWOO, and to collect data on baseline physical, chemical, and biological conditions of the proposed discharge site. The SWOO began operation in 1986 to transport primary treated effluent from the Richmond-Sunset WPCP. Discharge of secondary treated effluent began in September 1993 when the Richmond-Sunset WPCP was replaced by the Oceanside WPCP. Offshore monitoring programs from 1986 to 1996 were conducted under various plans that compared impacted sites near the outfall to a single reference site and they are discussed in Appendix A. The most significant findings from those previous studies (BWPC 1988, 1989, 1990, 1992a, 1992b, 1993, 1994, 1995, WQB 1997a, 1997b, Niemi and Warheit 1989, Kellogg, et al. 1998) included:

- A single reference station was inadequate to fully characterize reference conditions or to determine if observed differences between stations were attributable to natural variability or actual differences.
- Seasonal variability was the predominant factor affecting differences in water quality conditions, grain size distribution,

sediment chemistry, and abundances of invertebrates and fish in the study area.

- Interpretation of potential outfall impacts was possibly confounded by the proximity of the SWOO to the mouth of the San Francisco Estuary.
- There was little detectable evidence of the effluent plume in the water column away from the ZID.

To address these issues, City biologists and U.S. EPA biologists and statisticians discussed ways to improve the study design so that possible effects from the SWOO discharge might either be detected or determined to be environmentally negligible. The monitoring program was modified in the 1997 NPDES

permit, expanding the study area to include multiple reference sites. Including more reference sites increased the statistical power to detect differences between sites due to effects from the SWOO. The sampling frequency was reduced to one annual event, eliminating the effects of seasonal variability on the data. Sampling was scheduled during the fall when sediments in the study area are least disturbed and when benthic infauna are most abundant.

Seven stations (01, 02, 04, 06, 25, 28, and 31) that were monitored under previous permits remained part of the program. Initially, 40 additional offshore sample sites (stations 32-71) for sediment and benthic infauna sampling were added to the study. The new stations were

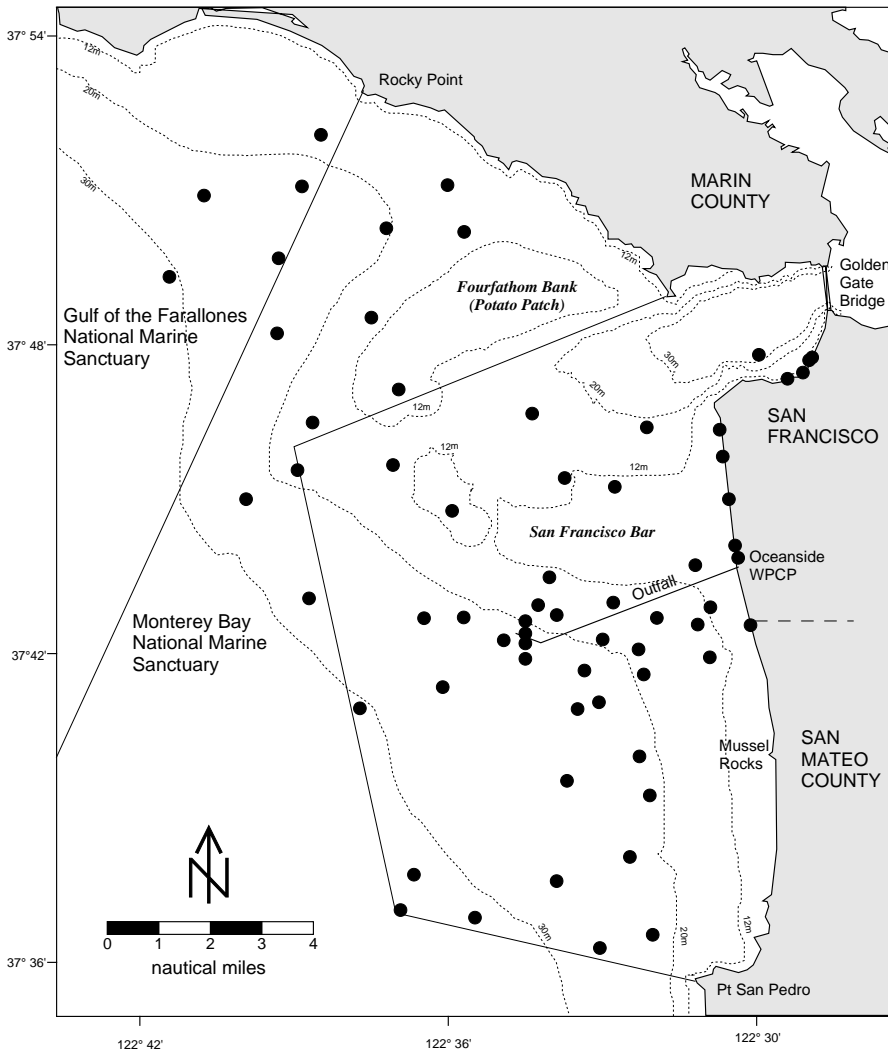


Figure 1-1  
*Study area with station locations*

located using the U.S. EPA's Environmental Monitoring Assessment Program (EMAP) random sampling site selection process in which sample sites are randomly selected in a grid pattern within the study area (Overton, et al. 1990, White, et al. 1992). The expanded study area extends from Rocky Point in Marin County south to Point San Pedro in San Mateo County. This expansion ensured the inclusion of reference locations in similar hydrological and sedimentary environments as the SWOO. In addition, the new study area spans the mouth of the San Francisco Estuary so that the effects of outflow through the Golden Gate might be detected. Three previous summary reports covering five years (WQB 2003a), eight years (NRD 2006a), and twelve years (NRLMD 2010a) of data collected under the regional monitoring program and subsequent monitoring and analysis has confirmed that potential impacts from the SWOO discharge can only be evaluated with a regional perspective (Figure 1-2).

The Southwest Ocean Outfall Regional Monitoring Program has two main components:

- 1) The Beach Monitoring Program involves measurements of bacteria concentrations at recreational beaches and notification of the public when State standards are exceeded or when a combined sewer discharge occurs.
- 2) The Offshore Monitoring Program involves collection and analysis of physical, chemical, and biological parameters in order to assess and compare outfall (potentially impacted) and reference conditions utilizing:
  - sediment quality (physical and chemical)
  - benthic infauna community structure
  - demersal fish and epibenthic invertebrate community structure
  - physical anomalies and bioaccumulation of contaminants in organism tissues

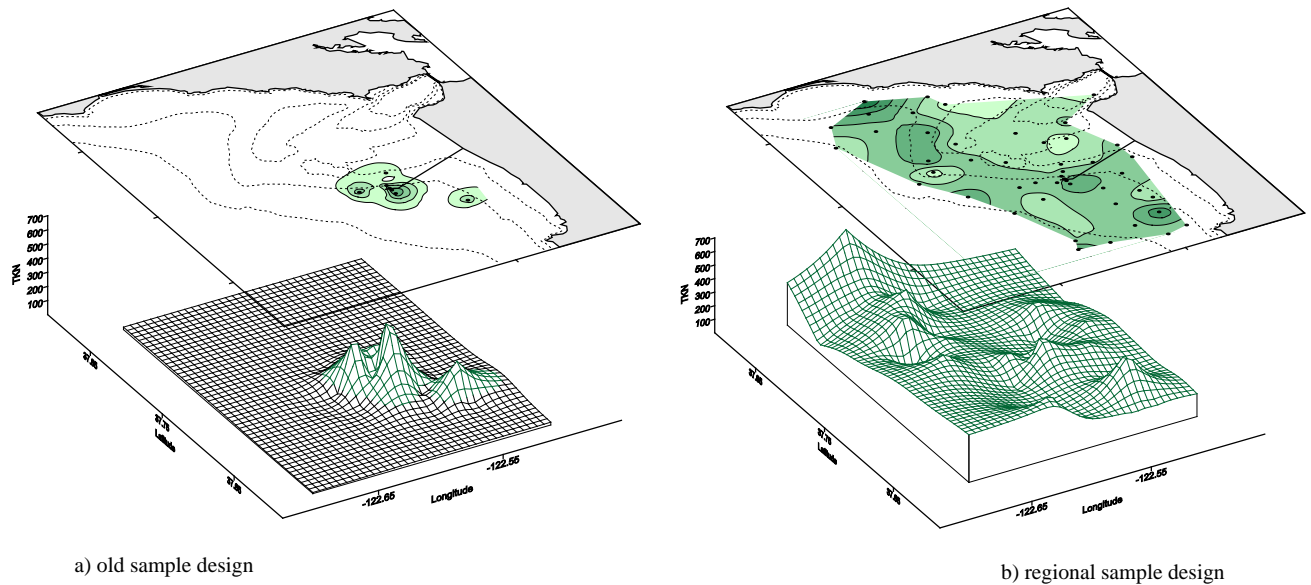


Figure 1-2a-b

*1997 sediment total nitrogen (TKN) concentrations. a) shows only stations sampled under the old sample design; b) shows those stations plus the additional stations of the regional monitoring program. There is an apparent peak of sediment nitrogen at the outfall compared to the single reference station in a), but that peak blends into the “TKN landscape” in b). The San Francisco Estuary watershed is a probable source of nitrogen in the study area. The figure pair illustrates the need for a regional perspective when interpreting potential SWOO impacts.*

### 1.2.1. SETTING

The SWOO study area lies on the continental shelf within the nearshore area of the Gulf of the Farallones. The Gulf of the Farallones is bordered by Point Reyes to the north, Point San Pedro to the south and extending about 26 nautical miles west of the Golden Gate, to the Farallon Islands. The primary influences on the near shore water quality and sediment characteristics within the Gulf include the broad changes in wind and current conditions that define oceanographic seasons, tidal currents, and outflow from the San Francisco Estuary (Brown and Caldwell 1971a,b). The San Francisco Estuary has historically been a major supplier of fine sediments to the Gulf (Noble and Gelfenbaum 1988), with the magnitude of the effects depending on the season and amount of freshwater outflow from the Sacramento-San Joaquin Rivers. Freshwater outflow and sediments transported from the estuary also have the potential to transport nutrients and contaminants into the study area. The San Francisco Estuary drains 40% of the land area of California (Conomos 1979) and nearly half of the state's total runoff (SFEP 1993) including most the agricultural, industrial, and municipal wastewater inputs from that aerial extent. Outflow from the Estuary is substantial even during drought years (Kellogg, et al. 1998). Further discussion of the setting of the study area including oceanographic seasons, El Niño and La Niña events, and marine sanctuaries is presented in Appendix A.

#### 1.2.1.1. Oceanographic Seasons

The California near shore marine climate consists of two major seasons: the California Current season during which the principal near shore current flow is southerly; and the Davidson Current season during which the principal near shore current direction is northerly.

The California Current season usually occurs between February or March and November and comprises an upwelling and an oceanic period.

The upwelling period begins about February or March and extends into late summer. Persistent west and northwest winds result in upwelling of deep, cold, nutrient rich waters into the Gulf of the Farallones. Weather systems are seldom stationary and upwelling may occur sporadically during this period. In the late summer and fall (August or September to November) the northwest winds subside and upwelling ceases. The oceanic period occurs between the cessation of upwelling and the start of the Davidson Current season and is a time when both ocean surface temperatures and salinities are at maxima.

From approximately November to February or March the northward flowing Davidson Current displaces the California Current offshore. During the rainy season, low-pressure systems offshore produce south and southwest winds along the central California coast. They produce onshore currents that are blocked by the northwest trending coast and gain a northerly direction that generates the Davidson Current. Because the low pressure systems do not remain stationary, the Davidson Current does not occur at all times and the end of the Davidson Current period can be diffuse and difficult to pinpoint (Bolin and Abbot 1963, Pavlova 1966, Schwartzlose and Reid 1972).

#### 1.2.1.2. Large Scale Oceanographic Phenomena

The intermittent oceanographic phenomena known as El Niño and La Niña have global weather consequences and may significantly impact water quality and sediment transport in the Gulf of the Farallones by altering normal seasonal climate patterns. El Niño events are characterized by warmer than normal sea-surface temperatures in the equatorial Pacific Ocean. La Niña events are characterized by colder than normal sea-surface temperatures in the equatorial Pacific Ocean. Both types of events can vary in strength and local effects are difficult to predict. Locally, El Niño winters have included both greater than normal precipitation and drought. Oceanographically, the primary local effects of wet El Niño events are intensified storms

and sustained southwest winds that reduce upwelling and result in higher than normal sea surface temperatures (USGS 1999). An unusually strong El Niño event occurred during in 1997-1998 (NOAA 1999a), with over two times the normal annual rainfall recorded in San Francisco (WRCC 1999). This event was followed by a La Niña that caused unusually strong upwelling of cold, nutrient-rich waters off the northern California coast (USGS 1999). Another, particularly strong, La Niña occurred in 2007-2008 and an El Niño event occurred in 2009-2010.

The Pacific Decadal Oscillation (PDO) is a long-lived climatic pattern affecting the northern Pacific above about 20° N, with a warm (or positive) phase and a cool (or negative) phase, that are thought to alternate about every 20-30 years. The underlying mechanisms of PDO are not well understood. The PDO has been in the warm/positive phase since 1977, the effects of which are generally increased biological production in coastal waters of Alaska and inhibited production off the west coast of the contiguous United States. A regime change to the cold/negative phase may have begun in 2008.

The North Pacific Gyre Oscillation (NPGO) is a newly described pattern of climate change that significantly correlates with previously unexplained fluctuations in salinity, nutrients and chlorophyll (Di Lorenzo et al. 2008). Fluctuations in the NPGO are driven by regional and basin-scale variations in wind-driven upwelling and horizontal advection.

#### 1.2.1.3. National Marine Sanctuaries

Three national marine sanctuaries lie partially within or adjacent to the Gulf of the Farallones. Data collected from the SWOO regional monitoring program provide important information relevant to the marine habitat management goals of these marine sanctuaries.

The SWOO is surrounded on three sides by the boundary of the Monterey Bay National Marine Sanctuary (MBNMS). An exclusion zone which extends off the north coast of San Mateo County and the City and County of San

Francisco between Point Bonita and Point San Pedro was originally created to encompass the SWOO, the shipping channel providing access to and from San Francisco Bay, and the Golden Gate dredged material disposal site associated with the shipping channel (NOAA 1992). Ten stations of the SWOO Regional Monitoring Program lie within the MBNMS.

Adjacent to the northwest MBNMS boundary, the Gulf of the Farallones National Marine Sanctuary includes the Farallon Islands on the western edge of the Gulf and near shore tidal flats, rocky intertidal areas, wetlands, subtidal reefs, and coastal beaches north of San Francisco. Five stations of the SWOO Regional Monitoring Program lie within the GFNMS.

Cordell Bank National Marine Sanctuary is an offshore sanctuary near the edge of the continental shelf at the northern most end of the Farallon Ridge.

### **1.3. MONITORING INDICATORS**

The Clean Water Act was enacted to ensure that the nation's waters remain fishable and swimmable. The California Ocean Plan has the goal of preventing the degradation of beneficial uses of water bodies. In conformance with those mandates, the primary objectives of the SWOO Monitoring Program are 1) to evaluate near shore bacteria concentrations and inform the public when water-borne bacteria along the City's shoreline are elevated, either from combined sewer discharges or other sources; and 2) to assess potential impacts on ecological communities in the receiving water environment from the presence of the SWOO discharge.

#### **1.3.1. BENEFICIAL USES**

(see Beach Monitoring Program, Section 3)

When the capacity of the combined sewer system is exceeded during wet weather periods, treated discharges occur onto recreational beaches that have several beneficial uses related to recreation and marine habitat. Ocean Beach, China Beach, and Baker Beach occupy the western and northern shores of the City and are



part of the Golden Gate National Recreation Area. These beaches provide recreational activities for Bay Area residents and tourists throughout the year. Walking, jogging, surfing, kite surfing, stand-up paddle boarding, and fishing are the primary activities that occur along the City's beaches. Weekly sampling for bacteria (total coliform, *Escherichia coli*, and enterococcus) in the surf zone at these beaches enables the City to provide information to the public about compliance with state water contact recreation standards. The City maintains a Recreational Water Quality Hotline (1-877-SF BEACH) and a web site (<http://beaches.sfwater.org>) with current water quality information.

### 1.3.2 SEDIMENT QUALITY

(see Marine Sediments, Section 4)

Physical and chemical sediment measurements are important components of marine monitoring programs designed to assess the environmental effects of wastewater discharges around ocean outfalls (Bilyard 1987). Wastewater discharges may change the properties of bottom sediments next to outfalls, which in turn may affect the natural biological communities. Such discharges are generally high in suspended solids and organic matter and, with the addition of urban runoff and industrial discharges, may contain high levels of metals, petroleum hydrocarbons, and other toxic compounds. Most sewage-related contaminants enter the marine environment as particulate matter associated with fine suspended sediments (Parker and Lee 1981, Reed et al. 1986). In addition to potentially altering the chemistry of marine sediments, wastewater discharges may physically affect the grain size distribution of bottom sediments around the outfall by adding large inputs of fine sediment and organic matter, as well as by increasing sediment re-suspension near the discharge (Reed et al. 1986). These physical and chemical sediment changes may also affect benthic organisms.

Generally, fine-grained suspended and bottom sediment particles (silt and

clay) accumulate greater concentrations of contaminants than coarser particles, especially those contaminants with low water solubility. Fine-grained particles have greater relative surface area and properties than coarser particles, allowing different physiochemical sorption and ion exchange of contaminants. A large part of the benthic community is supported by the food found in organic matter associated with fine-grained sediment particles. Sampling of the sediment surface layer provides information on the horizontal distribution of parameters such as particle size distribution and geochemical composition for the most recently deposited sediment (Mudroch and Azcue 1995).

Assessment of the solids and organic content of the sediments in the study area was determined through measurements of total solids (TS), total volatile solids (TVS), total organic carbon (TOC), and total Kjeldahl nitrogen (TKN). Maps of sediment grain size were created to determine physical impacts of the outfall and to help characterize any chemical and biological impacts of the outfall (as these effects may be greater where sediments are finer).

Bottom sediments are naturally variable, with their composition fluctuating in response to a variety of natural and anthropogenic factors (Boesch and Rosenberg 1981, Zmarzly et al. 1994). Factors affecting sediments include:

- 1) natural environmental conditions such as wave disturbance, currents, storms, and El Niño-La Niña events
- 2) ecological interactions such as feeding, burrowing and tube-building/cementing activities
- 3) anthropogenic disturbances such as wastewater discharges, oil spills, dredging, and construction

This study compares outfall stations with similar reference stations, as well as with historical data preceding the SWOO construction, with the goal of distinguishing SWOO-related effects from other effects.

### 1.3.3. COMMUNITY ANALYSES

(see Benthic Infauna, Section 5 and Demersal

## Fish and Epibenthic Invertebrates, Section 6)

The structure of biological communities (species composition, abundance, diversity) can be affected by wastewater discharges. Community properties are measured for comparison of outfall and reference areas and to determine the presence of balanced indigenous populations within and beyond the zone of initial dilution.

### 1.3.3.1. Benthic Infauna Communities

Benthic infauna are invertebrates, most of them sedentary, that live in sediments. Infauna near wastewater outfalls can be chronically exposed to sewage-derived pollutants and organic particulates that become incorporated into the bottom sediment (Khan 1980), and abundance of benthic infauna has been shown to fluctuate in response to organic input and toxicant concentrations (Swartz et al. 1986, Stull 1995). Community measurements are important components of marine monitoring programs designed to assess the environmental effects of wastewater discharges around ocean outfalls (Bilyard 1987).

Exposure to organic particulates and pollutants may result in the biological uptake of nutrients or toxicants by infauna species (Segar and Stamman 1986, Swartz et al. 1984), which can lead to changes in community population characteristics such as abundance and diversity (Pearson and Rosenberg 1978). Some infauna species respond to organic input with enhanced growth and reproduction, since the addition of low levels of organic particulates provides an additional food source for the organisms (Word et al. 1977). Filter-feeding organisms can be negatively affected by the physical clogging of feeding mechanisms (McCave 1981). The silting over of feeding and larval recruitment grounds may also affect infauna populations. In environments with high organic input, opportunistic deposit and detritus feeders (Word 1978) that can tolerate organically enriched sediments reproduce quickly and take advantage of an environment with reduced

biological competition (Pearson and Rosenberg 1978). Therefore, both species composition and diversity measures are important aspects in the evaluation of potential community impacts from an outfall (Levinton 1972, Reish 1980).

The Pearson/Rosenberg model (Pearson and Rosenberg 1978) predicts changes in benthic infauna communities subjected to organic inputs such as wastewater discharges. Under the most extreme conditions of organic input, a barren zone exists in which there is an absence of infauna. With increased distance from the input source, abundance increases reaching a maximum in the opportunist zone, which is dominated by a few extremely abundant opportunistic species. At a further distance from the input source, a transition zone exists where dominance and abundance decrease while diversity and species richness reach maximum values. Transition zone communities can be unpredictable due to seasonal recruitment of species that may cause large fluctuations in abundance and species richness. A normal community is present at some greater distance from the input source and is characterized by a decrease in abundance, species richness, and diversity compared to the transition zone. This study examines species composition, diversity measures and abundance to evaluate whether Pearson/Rosenberg-style impacts are occurring in the area of the outfall.

In addition to organic enrichment, a wastewater outfall may affect benthic infauna through the introduction of increased levels of toxic substances. Bioaccumulation of pollutants by infauna may occur by absorption across surface membranes, ingestion of sediments, diffusion across respiratory surfaces or by selective absorption by certain tissues (e.g., DDT in fatty tissues). The biological uptake of contaminants by infauna species can lead to acute and/or chronic toxicity effects. Acute effects may kill organisms in the short term; chronic effects reduce reproductive capacity and/or can affect larval development and growth and may affect populations over time (Anderson et al. 1983). Various organic pollutants and trace

elements have been associated with diseases and abnormalities of benthic invertebrates as well as fishes and marine mammals that feed on them (Sinderman 1979, Malins 1982). These compounds also affect the community structure (Grassle et al. 1981, Rygg 1985) so their impact can be detected using infaunal community analysis. In addition, benthic infauna are a primary food source for demersal fishes and epibenthic invertebrates, and may play a role in the transfer and bio-magnification of toxic substances to higher trophic levels (Parsons and Takahashi 1984, Spies 1984, Malins et al. 1985a,b). This study looks for such effects both by looking for bioaccumulated pollutants and physical abnormalities in individual organisms at higher trophic levels (Section 7) and by looking for community effects (Section 6).

Benthic communities fluctuate in response to a variety of natural and anthropogenic factors (Boesch and Rosenberg 1981, Zmarzly et al. 1994), which include:

- 1) natural environmental conditions such as temperature, wave impact, currents, storms, and El Niño-La Niña events
- 2) ecological interactions such as predation, competition, feeding and burrowing activities, and seasonal cycles of reproduction and recruitment; and
- 3) human disturbances such as wastewater discharges, oil spills, dredging, and construction

Although univariate analyses (diversity and abundance measures) describe changes in the community structure, multivariate methods (pattern analysis) are effective in describing those changes in relationship to both anthropogenic and natural factors (Thompson et al. 2000, Zmarzly et al. 1994). This study uses both univariate and multivariate analyses.

#### 1.3.3.2. Demersal Fish and Epibenthic Invertebrate Communities

A comprehensive evaluation of Demersal fish and epibenthic invertebrate communities within the SWOO study area was presented in the twelve-year summary report (NRLMD 2010a).

No trawling has been conducted since that report due to the listing of the longfin smelt as threatened by the State of California. Section 6, Demersal Fish and Epibenthic Invertebrates of the current report presents arguments to drop the trawl requirement from the offshore monitoring program because 1) the trawling has not revealed significant differences in outfall or reference area demersal fish or epibenthic invertebrates communities; 2) the trawl sampling program is not suited to finding outfall effects; 3) the demersal fish specimens collected are not necessarily representative of contaminant exposure to consumers of local fishes or of body burdens obtained within the Gulf of the Farallones; 4) the trawl sampling results in significant and unnecessary mortality to demersal fish and epibenthic organisms including listed species; 5) the trawl sampling destroys benthic habitat; 6) other new data sources of high quality are available; and 7) given the absence of outfall effects demonstrated by the data, the trawl program is expensive and burdensome to implement.

#### 1.3.4. PRIORITY POLLUTANT ANALYSES (see Marine Sediments, Section 4 and Physiological Effects and Bioaccumulation, Section 7)

Regulatory guidelines do not exist for pollutant concentrations in sediment or organisms for the region offshore of San Francisco

##### 1.3.4.1. Organic Pollutants

The deposition of organic pollutants, including polycyclic aromatic hydrocarbons (PAHs), polychlorinated biphenyls (PCBs), and pesticides, into marine sediments is a potential source of contamination for marine invertebrates, fishes, and the organisms (including humans) who ingest them (Malins et al. 1980, 1986).

Pesticides are only slightly soluble in water and tend to concentrate in sediments where they can be ingested by marine organisms and accumulate in fatty tissues (Malins et al. 1980).

The organochlorine pesticide DDT, which has been banned in the US since 1972, can cause neurotoxic and carcinogenic effects. 4,4' DDT and derivatives (4,4' DDD and 4,4' DDE) are extremely persistent in the environment, resist metabolism, have a strong affinity for lipids, and biomagnify in aquatic food webs (Gobas et al. 1993, Suedel et al. 1994).

PAHs, often derived from petroleum products, constitute most of the “oil and grease” regulated as a conventional pollutant under the Clean Water Act (Levensen and Barnard 1988). Petroleum and petroleum by-products may be composed of up to 40% PAHs, some of which are extremely carcinogenic (Malins et al. 1980, Hutzinger 1982). Once deposited in sediments, PAHs are strongly adsorbed to particulates and are generally unavailable for either de-sorption to the water column or microbial degradation, particularly when sediments are anaerobic (Wilson & Jones 1993 as cited in Law & Biscaya 1994). Anthropogenic sources of these organic pollutants include sewage outfalls, storm drains, petroleum operations, industrial discharges, shipping, and atmospheric fallout (Anderson and Gossett 1988). Asphalt sealants have recently been identified as a source of PAHs in runoff (WEF 2005). Sources of atmospheric contamination include burnt plant material from forest fires, charcoal used outdoors in food grilling, and car exhaust (SFEI 2000). In some marine environments PAHs occur naturally through seepage of oil from the ocean floor (Levensen and Barnard 1988).

PCBs are a group of over 200 organic chemicals manufactured from 1929 to 1979 and used in hydraulic fluids, lubricants, plasticizers, insulators in electrical transformers, and in carbonless copy paper; smaller quantities were also used as pesticide extenders and in inks, waxes, and other products (SFEI 2000). These pollutants were banned in 1979 because they were found to be extremely toxic in long-term exposures and can cause developmental abnormalities, disruption of the endocrine system, impairment of immune function, and cancer in organisms near the top of the marine

food chain, including humans who consume fish (SFEI 2000). Although manufacture of PCBs is now banned, the runoff from PCB-contaminated streams and urban areas continues to deliver these pollutants to the environment (SFEI 2004). PCBs accumulate in the fatty tissues of marine biota, where they are resistant to biological degradation (Malins et al. 1980, Hutzinger 1982).

This study examines both sediments and organisms for levels of organic pollutants of concern.

#### 1.3.4.2. Inorganic Pollutants – Metals

The deposition of trace metals in the marine environment is a concern because metals bioaccumulate in marine organisms and contribute to a variety of chronic health problems and developmental anomalies (Bryan 1971, McDermott et al. 1976, Sinderman 1979). Sources of trace metals include wastewater outfalls, erosion of soils and minerals, river discharge, and atmospheric fallout. Trace metals either remain in solution in the water column, or are held in suspension on the surface of fine-grained particles (Eisma and Irion 1988 as cited in Stevenson 2000). Studies indicate that some quantities of suspended material do not accumulate on the seafloor but remain in the water column and may be transported by currents that can distribute the material over long distances (Stevenson 2000). In order to affect aquatic organisms, metals must be in a form that is biologically available (Waldichuk 1985).

Many metals occur naturally depending upon the nature of the geo-chemical and rock-forming environment (Keller 1976). Sampling stations in the study area lie on the Farallon platform, which has a granitic or high-grade metamorphic basement that is exposed locally at Montara, Point Reyes, and the Farallon Islands (Cooper 1973). Granitic material naturally contains various concentrations of cadmium, chromium, copper, fluorine, iodine, lead, lithium, molybdenum, selenium, and zinc (Keller 1976). Nickel may be elevated in the region due to natural geologic sources such as

serpentine soils (SFEI 1996). Mercury is still entering San Francisco Bay from the leavings of mercury mining operations in upland watersheds during the late 1800s and early 1900s (SFEI 2002). This study examines both sediments and organisms for levels of metallic contaminants of concern.

#### 1.3.4.3 Bioaccumulation – Pollutants in Tissues (see Physiological Effects and Bioaccumulation, Section 7)

Measures of contaminant bioaccumulation and disease (which can result from immunological breakdown) characterize the effects of pollutants on individual fish and invertebrates. There is often a link between environmental contamination by organic compounds and marine vertebrate and invertebrate health problems (including increases in epizootic lesions, epithelial and hepatic tumor growth, ‘fin rot’, mutagenesis, lowered immune response) and decreased fitness (Sinderman 1979, Malins 1982, Malins et al. 1985a,b). Contamination with metals may result in reduced reproductive potential and developmental anomalies in fishes and invertebrates (Bryan 1971, Westernhagen and Dethlefsen 1975). There is also a connection between high metal concentrations and a variety of chronic health problems in fishes, including microbial diseases, skeletal anomalies due to interference with calcium metabolism, inhibited or accelerated enzyme activity due to interference with metal based enzyme systems, fin erosion, and behavioral changes (Bryan 1971, McDermott et al. 1976, Sinderman 1979). Interactive effects between metals and other contaminants frequently found in sewage effluent can increase their toxicity to marine organisms (Rhodes et al. 1985). This study analyzes levels of organic and metal contaminants in organism tissues as well as examining individual animals for abnormal characteristics that may be the result of toxic exposures.

Additionally, transfer of contaminants through the food chain has the potential to

cause human health concerns when fish or invertebrates at higher trophic levels are used as food. This concern for human health in relation to bioaccumulated pollutants led to the development of water quality criteria by regulatory agencies. These guidelines are not strict water quality limitations above which human health impact may be expected; however, as the number of fish consumption studies increases, greater effort has been made to identify the level of pollutant concentrations in edible tissues to determine human health risks.

### **1.4. REGIONAL MONITORING PROGRAM HISTORY**

#### 1.4.1. 1997 NPDES PERMIT (1997 to 2002)

##### 1.4.1.1. Beach Monitoring

Beach monitoring studies consisted of sampling at shoreline stations for bacteria concentrations and observational surveys of shoreline recreational use activities. Bacteria data collected for eight years prior to the issuance of the 1997 Oceanside NPDES permit included analysis for total and fecal coliform and enterococcus bacteria. A review of those data indicated the total coliform bacteria analysis was more conservative in measuring near shore bacteria contamination than the fecal coliform analysis, and timelier than the enterococcus bacteria analysis in providing results. Shoreline bacteria requirements in the 1997 NPDES permit therefore included analysis of only total coliform bacteria. One shoreline sampling site (station 21.1), at the foot of Sloat Boulevard, was added to the program in 1997 because the surfing community identified the site as an area of high use. Public notification of impaired water quality at recreational beaches was significantly enhanced during the life of this permit.

- A Recreational Beach Water Quality Hotline was established in 1998 to alert the public when water quality conditions were impaired or a combined sewer discharge had occurred. The hotline is reached by dialing 415-242-2214 (local)

or 1-877-SFBEACH (toll free).

- Beginning in 2001, recreational beach water quality status and data have been published on the internet at the SFPUC web site <http://beaches.sfwater.org>.

#### 1.4.1.1.1. Recreational Use Study

In order to address concerns raised by the San Francisco Chapter of the Surfrider Foundation, the 1997 permit included a provision requiring the City to complete a comprehensive Recreational Use Study along Ocean Beach. The purpose of the study was to make an assessment of the number of water contact users along Ocean Beach and to determine the impact from combined sewer discharges on water contact recreation. The study was conducted over a two-year period from October 1998 through September 2000. Results from the study (WQB 2001b), determined that water contact and non-water contact (including surf fishing) recreational activities along Ocean Beach are extensive. Of the 154,054 people observed during the two-year study, the majority of users (83%) were involved in non-water contact recreation; and of those involved in water contact recreation, up to 25% were surfers. The number of users observed participating in water contact recreation following a combined sewer discharge event represented less than one percent of all water contact users observed during the study. The two-year investigation concluded that most discharge events occur in mid-winter and have little impact on recreational use, as little use was observed during the cold, short days of winter. Isolated combined sewer discharge events that occur in early spring have the potential to impact more users as beach use increases when days become longer and the duration of storm events are shorter, contributing to good surfing conditions.

#### 1.4.1.1.2. Offshore Monitoring

The expanded offshore monitoring program implemented with the 1997 permit provided the opportunity to fully characterize the study area and to better evaluate potential SWOO impacts

with a regional perspective. While the program continued to identify potential impacts from the SWOO, the new sample design incorporated appropriate reference conditions and could therefore better address confounding effects of outflow from the San Francisco Estuary. Adaptive management incorporated into the permit allowed for periodic changes to the sample design throughout the permit cycle as necessary to obtain more meaningful and useful data.

Fisheries sampling was initially conducted throughout the study area using a stratified-random design independent of the sediment and infauna stations. Fisheries sampling was later modified to include a subset of benthic stations with the goal of detecting trends and relating fish and epibenthic invertebrate community data to sediment and infauna data.

#### 1.4.1.1.3. Adaptive Management during the 1997 Permit

The 1997 permit allows for dynamic implementation of the monitoring program to maximize the relevance and usefulness of the data gathered.

- Sample collection at reference station 35 was not possible during the 1998 survey, rocks prevented the grab sampler from closing completely, and no acceptable sample was collected. Several attempts to collect sediment and benthic infauna samples from alternative locations in close proximity to station 35 were also unsuccessful because of the hard substratum. A substitute site (station 72) was selected for benthic infauna and sediment sampling in lieu of sampling at station 35 to complete the 1998 survey.
- Benthic sampling attempted at station 35 in 2000 was only successful for infauna. Subsequent samples for sediment analysis were unobtainable due to the rocky bottom substratum. The substitute station 72 continued to be sampled. Trawling was also unsuccessful at station 35 because of the rocky substratum (torn

trawl net) and station 34 was adopted as a replacement trawl station.

- Sampling in Trawl Stratum C located just outside the Golden Gate was inconsistent. Extreme tidal currents, rocky bottom substrate, and highly variable depth contours, prevented the collection of successful trawls from within Stratum C in 1998. In addition, the trawl net was snagged, torn, and nearly lost on an unidentified bottom obstruction in Stratum C that year.
- The trawl sampling strategy within strata was discontinued after 1998 because of difficulty in interpreting results. Sampling for demersal fish and epibenthic invertebrates was modified to correspond directly with discrete sediment/benthic infauna sampling sites. Trawl sampling was conducted at 20 of the sediment/benthic infauna sample sites.
- Station 46 was not sampled for benthic infauna after the 1998 survey and stations 44 and 49 were not sampled after the 1999 survey. Previous survey samples consisted of very coarse grain sizes that were inappropriate to live-sieve, and the infauna communities were very different from those of the reference and outfall stations, and not comparable (WQB 1998, 1999). Sediment samples at these stations continued to be collected and analyzed. Stations 41 and 42 were discontinued after the 2001 survey because of their coarse sediments and stations 44, 46, and 49 were discontinued for sediment sampling at the same time.
- The City elected to conduct whole sediment toxicity testing of sediment collected from the 2000 monitoring survey. Amphipod survival, using *Eohaustorius* spp. was measured at 24 stations that were also sampled for sediment organic and inorganic pollutant analyses and benthic infauna. Along with sediment chemistry and benthic

infauna community analysis, sediment toxicity completes the sediment quality triad environmental monitoring strategy (Chapman, et al. 1986). Sediment toxicity testing is used to assess possible contaminant effects not detected in sediment chemistry or benthic infauna analyses. Results of the investigation indicated no detectable toxicity to this amphipod species at any of the sample sites. Mean survival was greater than 90% at all stations with no statistically significant differences detected between test samples and controls (WQB 2001a). Subsequent sediment toxicity testing in the study area was not warranted because survival percentages in the 2000 survey were uniformly high.

- A pattern of high benthic infauna abundance at the near shore end of the outfall (e.g. station 57) led to the addition, in 2002, of seven stations (73, 74, 75, 76, 77, 78, and 79) along the length of the pipe to examine a possible reef effect from the structure itself.

#### 1.4.1.4. Reports Submitted Under the 1997 Permit

Environmental monitoring data collected for the Southwest Ocean Outfall Regional Monitoring Program in 1997, 1998, 1999, and 2000 were compiled, analyzed, and reported to the U.S. EPA and RWQCB in annual monitoring reports (WQB 1998, 1999, 2000, 2001a). Data collected in 2001 were presented in a five-year summary report covering 1997 through 2001 (WQB 2003a). Data collected in 2002 were submitted in a data report (WQB 2003b). A recreational use study along Ocean Beach was conducted from 1998 to 2000 and reported in 2001 (WQB 2001b).

#### 1.4.2. 2003 NPDES PERMIT (2003 to 2008)

##### 1.4.2.1. Beach Monitoring

The following changes to the beach monitoring program occurred in the October 2003:

- The number of indicators measured increased from one (total coliform bacteria) to three (total coliform, *Escherichia coli*, and enterococcus bacteria).
- The frequency of routine monitoring decreased from three times per week to once per week.
- The number of stations routinely monitored was reduced from 9 to 7.
- Recreational use observations were to be made after combined sewer discharges in order to determine their effect on beach use.

These changes were deemed to bring beach monitoring in San Francisco County more in line with monitoring conducted by other counties in California.

- In July 2007 the City adopted a confirmation approach to posting beaches at certain locations as recommended by the Beach Water Quality Workgroup of the California State Water Resources Control Board.

#### 1.4.2.2. Offshore Monitoring

The following change to the offshore monitoring component was made in the 2003 permit:

- The minimum number of trawls required for demersal fish and epibenthic invertebrate community assessments was reduced from eight to two, one from an outfall station and one from a reference station.

This change acknowledged that, based upon analyses of previous data, the mobile organisms collected by trawl net had not shown an outfall effect. In addition, even though samples were processed on board and returned to sea as quickly as possible, substantial mortality occurred that was not warranted by the information gained. By requiring minimum trawls, the presence or absence of a balanced indigenous fauna at the outfall and possible effects from the discharge can still be determined by comparison to reference conditions.

#### 1.4.2.3. Marine Mammal Report

In order to address concerns raised by NOAA Fisheries and the U.S. Fish & Wildlife Service, the 2003 permit included a provision requiring the City to produce a report that identifies "... monitoring methodologies to determine the presence in wastewater of pathogens with the potential to affect marine mammals." A report was submitted in October 2005 in fulfillment with that requirement (Casteel 2005). A thorough literature review revealed that little or no information is available on the environmental occurrence, fate, and transport of *Toxoplasma gondii*, *Sarcocystis neurona* and marine mammal morbilliviruses. An understanding of such information would be useful in support of studies that attempt to identify and link pathogens to the occurrence of infectious disease in marine mammals. That information would also be helpful in determining possible options for pathogen control in wastewater and stormwater. New and improved procedures, such as ultrafiltration, may prove to be important advancements for marine mammal pathogen detection in water. Polymerase Chain Reaction (PCR) based detection and characterization assays are likely to play important future roles in monitoring methods for marine mammal pathogens. Techniques for protozoan parasite and morbillivirus detection should be investigated in a research effort to develop environmental methods for marine mammal pathogens in wastewater and stormwater. The objective of this process would be to develop basic protocols generated from bench-scale efforts in the laboratory. Such work would most likely involve matrix spike experiments using actual water matrices. Because there are important issues in dealing with environmental samples and in the use and interpretation of some molecular methods, development of a complete method would require an appreciable laboratory effort and collaboration by investigators in the areas of environmental and veterinary microbiology, in addition to water utility professionals (Casteel 2005).



#### 1.4.2.4. Adaptive Management during the 2003 Permit

- The reef effect investigation began in 2002 led to the addition of benthic station 80 in 2004, located at the end of the North San Mateo County Sanitation District outfall.
- The coarse grained stations (42, 44, 46, and 49) were sampled again in 2004 for sediment and benthic infauna to confirm that the patterns established earlier still prevailed.
- Sample collection at reference station 35 remained inconsistent due to hard substratum and substitute station 72 has been sampled continuously.
- Beginning in 2006 sediment fines were reported as combined silt and clay rather than as separate silt and clay fractions due to the small percentage of fines in the study area.

#### 1.4.2.5. Reports Submitted Under the 2003 Permit

Environmental monitoring data collected for the Southwest Ocean Outfall Regional Monitoring Program in 2003, 2005, 2006, and 2007 were compiled, analyzed, and reported to the U.S. EPA and RWQCB in annual data reports (WQB 2004; NRD 2006b; NRLMD 2007, 2008). Data collected in 2004 were presented in an eight-year summary report covering 1997 through 2004 (NRD 2006a). Data collected in 2008 were presented in a twelve-year summary report covering 1997 through 2008 (NRLMD 2010a). A report discussing techniques for detection of pathogens of concern for marine mammals was submitted in October 2005 (Casteel 2005).

#### 1.4.3. 2009 NPDES PERMIT (2009 to present)

##### 1.4.3.1. Beach Monitoring

The following change to the beach monitoring program occurred in October 2012:

- Recreational use observations are now

made at every beach visit (sampling, posting, de-posting) instead of just after treated combined sewer discharges.

##### 1.4.3.2. Offshore Monitoring

##### 1.4.3.3. Adaptive Management during the 2009 Permit

- The coarse grained stations (42, 44, 46, and 49) were sampled again in 2010 for sediment and benthic infauna to confirm that the patterns established earlier still prevailed. Stations 44 and 46 have not been processed for infauna.
- Trawl sampling was curtailed due to the listing of the longfin smelt as threatened by the State of California.

##### 1.4.3.4. Reports Submitted Under the 2009 Permit

Environmental monitoring data collected for the Southwest Ocean Outfall Regional Monitoring Program in 2009, 2010, and 2011 were compiled, analyzed, and reported to the U.S. EPA and RWQCB in annual data reports (NRLMD 2010b, 2011, 2012). Data collected in 2012 are analyzed and reported herein.

## **1.5. ACKNOWLEDGMENTS**

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SFPUC Natural Resources and Lands Management Division  
Oceanside Biology Laboratory

1997-2012 Summary Report – Pat Conroy Executive Editor

Michael Kellogg	Executive Summary, Introduction
Pat Conroy	Methods
Diane O'Donohue	Beach Monitoring Program
Patricia McGregor	Marine Sediments
Dot Norris	Benthic Infauna
Heather Peterson	Demersal Fish and Epibenthic Invertebrates
Laura Targart	Physiological Effects and Bioaccumulation

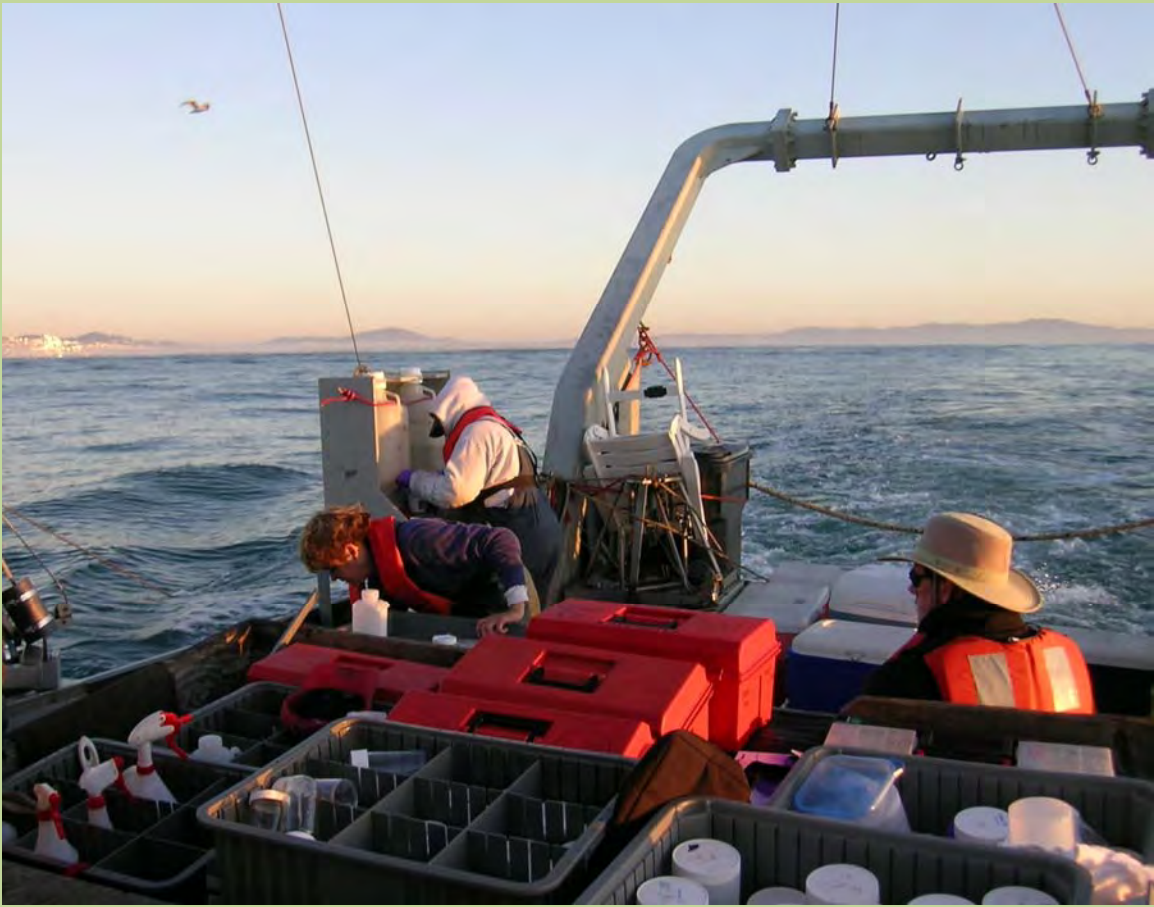
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Ellen Natesan	NRLMD Planning and Regulatory Compliance, Manager
Rod Miller	WQD Director of Laboratories
Laura Pagano	Wastewater Enterprise, Regulatory Manager
Michael Kellogg	NRLMD Supervising Biologist, Marine Biology
Paul McGregor	WQD Supervising Biologist, Microbiology
Ken Lee	WQD Supervisor of Laboratories, Wastewater
Tony Rattonetti	WQD Supervising Chemist, Inorganic Analyses
David Fok	WQD Supervising Chemist, Organic Analyses
Dolson Kwan	WQD Supervising Chemist, Process Chemistry
Hsiao-Lung Chang	WQD Database Administrator
Alex Liu	WQD Database Administrator

Consultants

Jim Christmann	Monterey Canyon Research Vessels, Inc. R/V <i>Shana Rae</i> , Santa Cruz, CA (1997, 1999-2012)
Scott Francis	West Coast Seaworks R/V <i>White Lightning</i> , Alameda, CA (1998)
Susan McCormick	Benthic Infauna Sorting and Quality Assurance, Georgetown, CA (1997-2012)
Dr. Michael Johnson	Data Analysis, Davis, CA
ToxScan, Inc.	Organic Pollutant Analysis, Watsonville, CA (1997 & 1998)
Frontier GeoSciences	Inorganic Pollutant Analysis, Seattle, WA (2000)

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## **SECTION 2**

## **METHODS**

# METHODS

Specific methods for each component and methods common to multiple components of the SWOO monitoring program are discussed below.

## 2.1. BEACH WATER QUALITY MONITORING

### 2.1.1. FIELD SAMPLING

The City sampled ten shoreline stations between the Golden Gate Bridge and Mussel Rocks during the years 1997 to 2008. These stations are mapped in Figure 2-1 and detailed location information is given in Table 2-1. In general, stations were sampled three times per week from 1997 through September 2003 and once per week thereafter. Station 15E has been collected regularly since October 2002. Stations 20, 21, and 22, which were originally collected regularly, are currently collected only when combined sewer discharges occur (this has been true of station 22 since 1998 and of stations 20 and 21 since October 2003).

Water samples were collected from the surf into sterile containers, stored at or below 10°C, and was analyzed in the laboratory within six hours of collection for bacterial concentration assessment. This report includes recreational use observations taken since 2008 whenever combined sewer discharges occurred; on each day of re-sampling after the discharge (when feasible), and on the day the beach was posted and de-posted. Sample collection and transfer to the laboratory included chain of custody documentation and procedures. When bacteria levels at a site exceeded relevant standards, the site was re-sampled daily until counts were again within limits regarded as safe for recreational activities.

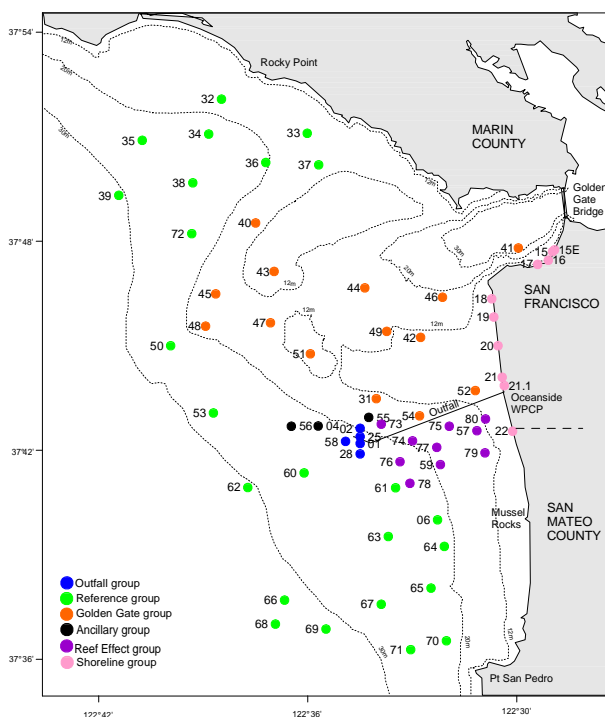


Figure 2-1  
Shoreline and offshore sampling stations.

### 2.1.2. LABORATORY SAMPLE ANALYSIS

City and County of San Francisco personnel at the Millbrae Water Quality Bureau Laboratory analyzed water samples for total coliform, *Escherichia coli*, and enterococcus bacteria. From 1997 through September 2003, all samples were analyzed for total coliform bacteria using membrane filtration method 9222B in Standard Methods for the Examination of Water and Wastewater (Eaton et al. 2005). Sample aliquots of 1 ml and 10 ml were processed, yielding a detection limit of 10 mpn/100mL (most probable number per 100 milliliters of sample). Beginning in October 2003, total coliform and *Escherichia coli* bacteria were measured using the Colilert®-18 Quanti-Tray® (IDEXX Laboratories, Inc.) formulation of the enzyme substrate test (Clesceri et al. 2005) and enterococcus bacteria were measured using Enterolert™ Quanti-Tray® methods (IDEXX Laboratories, Inc.). These marine samples were diluted for analysis, and a 10mL aliquot was used per test.

Throughout the study period, quality assurance practices as outlined in Standard Methods for the Examination of Water and Wastewater (Rice *et al.* 2012) were strictly followed, and annual State of California Environmental Laboratory Accreditation Program certification was maintained.

## 2.2. OCEAN FIELD SAMPLING

Annual offshore sampling was completed in the months of September or October during the Oceanic period of the California Current season (see Oceanographic Seasons 1.2.1.1). Historical studies show that the Oceanic period is the time of greatest infaunal abundance and the greatest level of settled fine sediments, which improves the chances of detecting priority organic and inorganic pollutants in the sediments. Sampling was conducted aboard the research vessel *Shana Rae*, Monterey Canyon Research Vessels, Inc., Santa Cruz, California in all years except

1998, when sampling was conducted from the R/V *White Lightning*, operated by West Coast Seaworks, Alameda, California. Stations were located in the field using a differential global positioning system. Benthic collection stations for sediment and infauna are mapped in Figure 2-1, with their exact coordinates given in Table 2-2.

### 2.2.1. BENTHIC MONITORING

The sampling scheme incorporated between 47 and 55 stations per year; pre-1997 data are available for five of these stations. Infauna and sediment data are available beginning in 1982 for stations 01, 02, 04, and 06; and in 1991 for station 31 (BWPC 1984, 1988-1990, 1992a,b, 1994, 1995; WQB 1997a,b). The station depths ranged from 10.5 to 36 meters for the twelve-year period. At least three grab samples were collected at each station using a 0.1 m<sup>2</sup> Smith-McIntyre bottom sampler. One grab sample was used for benthic infauna analysis. The

Table 2-1  
Shoreline sampling stations

Beach (Station)	Latitude	Longitude	Description
<u>Baker Beach</u>			
15	37°47.700'	122°28.980'	At the point where Lobos Creek enters the surf (exact location varies over time as the stream meanders)
15E	37°47.754'	122°28.920'	In the surf opposite small path from upper parking lot to beach
16	37°47.460'	122°29.100'	In the surf directly opposite the Sea Cliff 2 pump station
<u>China Beach</u>			
17	37°47.340'	122°29.400'	Near the Sea Cliff 1 pump station
<u>Ocean Beach</u>			
18	37°46.350'	122°30.720'	Foot of Balboa Street
19	37°45.828'	122°30.660'	Lincoln Way overflow structure
20	37°45.000'	122°30.540'	Foot of Pacheco Street
21	37°44.100'	122°30.420'	Vicente Street overflow structure
21.1	37°43.860'	122°30.360'	Foot of Sloat Boulevard
22	37°42.552'	122°30.120'	Lake Merced overflow structure at Fort Funston

Table 2-2  
Offshore sampling station coordinates

Station	Depth (meters)	Latitude	Longitude
1	26	37°42.21'	122°34.52'
2	23	37°42.63'	122°34.50'
4	25	37°42.70'	122°35.70'
6	25	37°40.00'	122°32.25'
25	25	37°42.23'	122°34.52'
28	27	37°41.90'	122°34.48'
31	14	37°43.50'	122°34.00'
32	21	37°52.08'	122°38.48'
33	19	37°51.10'	122°36.01'
34	24	37°51.08'	122°38.85'
35	28	37°50.90'	122°40.75'
36	22	37°50.26'	122°37.20'
37	17	37°50.19'	122°35.69'
38	27	37°49.68'	122°39.30'
39	31	37°49.32'	122°41.43'
40	16	37°48.53'	122°37.50'
41	17	37°47.81'	122°29.96'
42	13	37°45.24'	122°32.76'
43	12	37°47.13'	122°36.96'
44	17	37°46.66'	122°34.37'
45	20	37°46.49'	122°38.64'
46	18	37°46.40'	122°32.14'
47	16	37°45.66'	122°37.08'
48	22	37°45.56'	122°38.93'
49	14	37°45.41'	122°33.74'
50	29	37°45.00'	122°39.93'
51	13	37°44.77'	122°35.93'
52	13	37°43.72'	122°31.19'
53	30	37°43.07'	122°38.71'
54	16	37°42.99'	122°32.79'
55	19	37°42.94'	122°34.25'
56	26	37°42.69'	122°36.47'
57	16	37°42.56'	122°31.15'
58	26	37°42.26'	122°34.92'
59	22	37°41.59'	122°32.20'
60	29	37°41.35'	122°36.11'
61	26	37°40.92'	122°33.48'
62	33	37°40.94'	122°37.72'
63	28	37°39.53'	122°33.69'
64	25	37°39.24'	122°32.08'
65	27	37°38.05'	122°32.47'
66	33	37°37.70'	122°36.67'
67	29	37°37.58'	122°33.89'
68	35	37°37.02'	122°36.93'
69	33	37°36.87'	122°35.48'
70	25	37°36.54'	122°32.02'
71	30	37°36.28'	122°33.05'
72	27	37°48.22'	122°39.33'
73	18	37°42.75'	122°33.89'
74	21	37°42.28'	122°32.99'
75	17	37°42.69'	122°31.94'
76	26	37°41.67'	122°33.35'
77	21	37°42.08'	122°32.30'
78	25	37°41.05'	122°33.07'
79	16	37°41.93'	122°30.91'
80	13	37°42.90'	122°30.90'

remaining two grab samples were homogenized and used for physical and chemical sediment analyses. Grabs samples with disturbed or unevenly distributed surfaces were discarded and resampled.

#### 2.2.1.1. Sediment

If the sediment within the grab did not meet a minimum overall depth penetration of five centimeters the sample was discarded, and another sample was collected. From 1997 through 1999, the top two centimeters of sediment were composited for physical and chemical analyses. Starting in 2000, the top five centimeters of the grab sample were used for analyses to correspond with methods in the San Francisco Bay Regional Water Quality Control Board (RWQCB) Regional Monitoring Program for Trace Substances (RMP).

The top two or five centimeters (as described above) of each sediment grab sample were scooped into a Halar<sup>®</sup> coated stainless steel bucket, and homogenized using a Halar<sup>®</sup> coated spoon. Halar<sup>®</sup> is tough, smooth, and chemically inert and is used to prevent sample contamination. Unusual sediment texture and/or odor were noted if present. From this homogenized sample, five sub-samples were taken: a 200 gram sub-sample was transferred into a polyethylene container for physical and chemical analyses; a 200 gram sub-sample was transferred into a polyethylene container for inorganic analyses; a 200 gram sub-sample was transferred into a glass jar for organic analyses, and two 100 gram sub-samples were transferred into two glass amber jars for TOC and TKN analyses. All sample containers were pre-labeled and pre-cleaned. The homogenation bucket, utensils, brushes, and grab were all cleaned between stations using Alconox soap followed by four rinses, one each with sea water, 1% HCL solution, 100% methanol, and deionized water.

In all years, benthic stations were sampled for physical and chemical analyses except for organic priority pollutants; from 1997 through 2000, the number of stations sampled for organic

priority pollutants gradually increased. From 2001 forward all stations were sampled for organic priority pollutants.

A field blank for each container type (glass and polyethylene), filled with ultra-purified water (milli-Q® in 1997 and 1998, NANOpure® thereafter), was opened for the duration of sampling at one outfall station per day. Field blanks serve as a control for atmospheric contamination for the inorganic and organic analyses. All samples were stored on ice, transported to the laboratory, and held at 4°C prior to processing and analysis.

### 2.2.1.2. Benthic Infauna

Grab samples with undisturbed surfaces and at least 7cm sediment depth were processed for benthic infauna community assessment. If the sediment within the grab did not meet a minimum overall depth penetration of seven centimeters the sample was discarded and another sample was collected. The benthic infauna sample depth criterion is based on the vertical distribution of organisms in the sediment. Generally, seven centimeters is sufficient to capture 95% of benthic infauna

Table 2-3

*Analytical methods for bacteria, sediment, and tissue samples*

Analysis	Method	Reference	Limits
Total Coliform, Enterococcus, and E. coli bacteria	Enzyme Substrate Coliform Test Quanti-tray method	Rice et al. 2012 9222 B	10 mpn/100 mL
Grain Size	Dry sieving; hydrometer analysis of the portion passing the #230 sieve	Plumb 1981	1 ppm
Total Solids	Dried at 103 - 105 °C	Plumb 1981	0.1 ppm
Total Volatile Solids	Ignition at 550 °C	Plumb 1981	0.1 ppm
Total Organic Carbon	High Temperature Combustion Method	Rice et al. 2012 5310b	2 ppm
Total Nitrogen	Total Kjeldahl nitrogen by acidification and ammonia distillation followed by titration	Rice et al. 2012 4500(C)-NH3	0.2 mg/Kg
Organic Priority Pollutants	Isotope Dilution GC/MS (EPA Method 1613)	U.S. EPA 1993 Rice et al. 2012	Appendices D-4, G-2
Inorganic Priority Pollutants	a) Digestion (EPA Method 3050b) b) ICP-MS with Collision/Reaction cell (EPA 6020) c) Cold Vapor AA (EPA 7471A)	Tetra Tech 1986a U.S. EPA 1983 and 1993c Rice et al. 2012	Appendices D-6, G-3

organisms and species inhabiting fine sands to a depth of 20 cm (U.S. EPA 1987). Acceptable samples were live-sieved through nested 1.0 mm and 0.5 mm stainless steel mesh sieves. The material retained on each sieve was washed with seawater into separate sample jars. Jars were pre-labeled inside and outside with serial number, station number, and sieve mesh size; the sampling date was added to the external label at collection time. Animals adhering to the sieves were carefully removed with forceps and added to the sample jar for that sieve size. Seawater was decanted from each sample jar through a 0.25 mm Nitex® mesh screen-lid. An isotonic solution of magnesium chloride (MgCl<sub>2</sub>) was added as a relaxant for a minimum of fifteen minutes. In 2006, the relaxant was changed to MgSO<sub>4</sub>, since it was found to be more effective than MgCl<sub>2</sub>. The relaxant solution was then decanted through the 0.25 mm mesh screen-lid and replaced with a 10% solution of sodium borate-buffered formalin in seawater (a fixative). In 1997 and 1998, 3-4 drops of rose bengal (a protein-specific biological stain) were added to each sample jar to facilitate later sorting. This practice was discontinued in 1999 because the stain interfered with subsequent taxonomic identification of infauna. Sample jars were transported in plastic trays to the laboratory for processing.

Sediments at stations 41, 42, 44, and 46 were too coarse to pass through the 0.5 mm sieve screen making them impractical to live-sieve on board. Samples collected from these stations were fixed directly in a 10% solution of sodium borate-buffered formalin in seawater in separate, labeled 5-gallon plastic buckets fitted with water-tight lids, and sieved later with tap water in the laboratory. These stations were not collected for infauna between 2000 and 2003, and 2005 to 2008 (nor was the sample collected at station 46 in 1999 processed), because the observed sediment and benthic infauna characteristics at these stations were so different from other stations in the study area that they did not provide useful data for comparison to the outfall stations. These stations were sampled again in

2004 to confirm that these differences were still present. The following stations (year) were not sampled due to inclement weather: station 35 (2005, 2006), stations 45, 48 and 49 (2006).

## 2.2.2. DEMERSAL FISH AND EPIBENTHIC INVERTEBRATES

Trawl sampling for the SWOO monitoring program was conducted from 1982-2008 as a means of characterizing the resident fish and epibenthic invertebrate assemblages. The number, locations, and seasonality of trawl sampling have varied throughout this time period as NPDES permit requirements have changed. From 2003-2008, permit requirements reduced fishery sampling to single trawls at one outfall station (Station 1) and one reference station (Station 6) (see NRLMD, 2010a). With notification to the U.S. EPA, trawl sampling was curtailed in 2009 due to listing of Longfin smelt (*Spirinchus thaleichthys*) as a threatened species by the California Department of Fish and Wildlife.

## 2.2.3. BIOACCUMULATION AND PHYSICAL ANOMALIES

Organisms used to assess bioaccumulation of organic compounds and trace metals were collected from the outfall and reference areas. Outfall specimens were generally collected from station 01 and rarely, when necessary, also from stations 02, 25, and 28. Reference specimens generally came from stations 06 and 66 (both south of the outfall), with additional collections as needed from stations 32, 35, 39, 50, 53, 62, 65, and 70. Assessment of fish tissues for bioaccumulation was discontinued beginning in 2009. Dungeness crabs (*Cancer magister*) were collected using commercial crab pots set for at least 24 hours. Male crabs were placed in labeled burlap sacks, stored live on ice, and transferred to the laboratory where they were dissected within 48 hours of collection. Female crabs were infrequently utilized, when



insufficient numbers of males were collected.

From 1997-2003, only fish and macro-invertebrates collected in community analysis trawls were examined for tumors and gross physical anomalies at the time of taxonomic identification. Beginning in 2004, biologists examined all fish collected for community analysis and bioaccumulation analyses, and all of the Dungeness crab, whether collected in crab pots, or in trawls for community analysis or bioaccumulation analyses.

#### 2.2.4. OFFSHORE STATION GROUPS

Some figures and discussion refer to Golden Gate, reference, outfall, ancillary, and reef effect station groups or regions. These groupings (Figure 2-1) allow comparisons among similar stations and are based on proximity to the outfall, depth, sediment characteristics, benthic infauna cluster analysis and diversity patterns. To assess impacts from the SWOO discharge, outfall stations must be compared to a group of reference stations that are in a similar sedimentary environment and depth. The reference station group (stations 06, 32, 33, 34, 35-39, 50, 53, and 60-72) and the outfall station group (stations 01, 02, 25, 28, and 58) are both generally characterized by well-sorted very fine sand with a variable percentage of silt and clay, and similar infauna communities. The outfall group can be distinguished from the reference group by proximity to the outfall.

The Golden Gate station group comprises those stations on or near the sand bars with predominantly fine sands (stations 31, 40, 41, 43, 45, 47, 48, 51, 52, and 54) and those stations on or inside the sand bars with predominantly medium and coarse sands (stations 42, 44, 46, and 49). Stations within the Golden Gate group are generally shallower than stations in either the reference or outfall groups, have little or no silt and clay, and have different infauna communities. Stations 04, 55, 56, 57, and 59 constitute the ancillary group. Stations 73-79 were added in 2001 to assess the potential effects of the outfall structure on the benthic community

and are referred to as the reef effect stations (note that stations 57 and 59 were included in the reef effect stations in the analysis of the benthic infauna). Station 80 was added in 2004 to determine the effect of the North San Mateo County Sanitation District outfall, and for the present report it has been included with the reef effect stations. These groups were initially defined based upon benthic infauna cluster analysis and diversity patterns evident in the 1997 data set (WQB 1998). They have been slightly altered over the intervening years, and care must be taken in comparing this report with previous ones. Additional groupings of stations, identified and discussed in each section, are based upon cluster analysis and are referred to as cluster groups.

### 2.3. LABORATORY PROCESSING

Staff of the City and County of San Francisco, Public Utilities Commission, Natural Resources & Lands Management Division processed all samples, except for benthic infauna sample sorting (see Benthic Infauna 2.3.2), organic contaminant analysis of tissue in 1997 and 1998, and inorganic contaminant analysis of sediment in 2000 (see Priority Pollutants Analyses 2.3.5). All samples were preserved according to established protocols and analyzed within the recommended storage limits. Many of the methods used were taken from U.S. EPA guidance documents for biological monitoring programs associated with 301(h) waivers (Tetra Tech 1985 a-f; 1986 a-c).

#### 2.3.1. SEDIMENT

After thorough mixing in the laboratory, each sediment sample was split into portions for total Kjeldahl nitrogen (TKN) analysis, total volatile solids (TVS) analysis, total organic carbon (TOC) analysis, and grain size analysis. Total solids and TVS analyses were performed within thirty days of collection. TKN samples were stored at or below -20°C until being transported

to the City and County of San Francisco's process laboratory within 14 days of collection. All chemical analyses except TOC were performed on wet samples.

Grain size analysis was performed using dry sieve and hydrometer methods. In the dry sieve method, a series of sieves were stacked in order, coarsest on top, above a catch pan. Sieves sizes used were 4.75 mm, 2.0 mm, 1.0 mm, 1.5 mm, 0.25 mm, 0.125 mm, and 0.0625 mm. The entire stack was placed in a shaker apparatus. Air-dried sediment samples were placed in the top sieve and shaken for five minutes. The weight in grams of the portion of sample retained on each sieve was recorded. The fraction of sample retained in the catch pan represents the amount of silt and clay in each sample. Silt and clay fractions less than 1.0 gram were reported as a combination of silt/clay. Silt and clay fractions greater than 1.0 gram were further analyzed using the hydrometer method. The silt and clay portion of each such sample was placed in a 1L glass cylinder filled with double distilled water. A hydrometer was then placed in the cylinder and measurements of the distance the hydrometer settled over discrete time intervals were recorded. Calculations of the settlement distance and time intervals were used to determine the particle size fractions that distinguish silt from clay (beginning in 2005, silt and clay fractions were not separated using the hydrometer method). Duplicate analyses were conducted on 10% of the total number of sediment samples as a quality control check. A summary of the methods used in the physical and chemical sediment analyses is shown in Table 2-3.

### 2.3.2. BENTHIC INFAUNA

Samples were fixed in buffered formalin for 48 to 96 hours, then rinsed with tap water and transferred to a preservative solution of 70% ethanol. The City's benthic contractor (Susan McCormick, Georgetown, Ca) sorted the samples using a stereo dissecting microscope into five major taxonomic groups: Polychaeta,

Mollusca, Crustacea, Echinodermata and Others (all other taxa). In some samples in some years, polychaetes were further sorted by family. The 0.5 mm and 1.0 mm sieve fractions for each sample were processed separately, but the results were combined for statistical analyses. Sample residues were retained and 10% of them were subjected to a sorting efficiency quality control check. If any sample showed less than 95% sorting efficiency, the samples processed by the same sorter immediately before and after it were re-sorted and evaluated. If either of those samples failed the 95% sorting efficiency check, then all samples processed by that sorter were re-sorted.

The organisms from each sample were morphologically identified to the lowest possible taxon using published taxonomic keys and literature, in-house voucher sheets and reference collection, museum collections, and materials developed by the Southern California Association of Marine Invertebrate Taxonomists (SCAMIT). Species level identifications clearly increase the precision of station comparisons (Furse et al. 1984, Rosenberg et al. 1986) and are valuable in building a regional database for future long-term comparisons (Lenat and Barbour 1994). Organisms that did not fit described species were given a provisional identification and/or were sent to specialists for identification. Species identification sheets and a reference collection of representative species developed in-house aid in identification. For quality control purposes, randomly chosen samples from each taxonomist were re-identified by a different taxonomist. Discrepancies in taxonomic identifications were reviewed and resolved by discussion and re-examination of the specimens. Misidentifications by either the original or QC taxonomist resulted in review of all specimens identified by that taxonomist as belonging to the problem taxon or closely related taxa.

### 2.3.3. DEMERSAL FISH AND EPIBENTHIC INVERTEBRATES

See NRLMD (2010a) for details of demersal fish and epibenthic invertebrate laboratory processing.

### 2.3.4. BIOACCUMULATION

Individual crabs (*Cancer magister*) were divided in half; one half was used for organic analysis and the other half for metals analyses. Generally, each of three replicates consisted of ten crabs.

Crabs were rinsed with Barnstead NANOpure® Type I ultrapure deionized water (NANOpure water) prior to dissection, and dissected using clean room conditions in Nuair model NU-201-324, series 13, positive flow dissection hoods using techniques recommended by TetraTech (1986a). Instruments contaminated with mucus or carapace material did not contact tissue sampled for analyses. Hepatopancreas tissue and muscle tissue from the legs and claws were dissected from Dungeness crab.

Dissections for organic analyses were performed on new, kilned aluminum foil that was changed between each replicate. Stainless steel or Teflon-coated instruments used for dissections were rinsed with ultra-purified water and methanol during all dissections and between the processing of individual organisms. The dissected tissues were placed in new, pre-cleaned, glass jars (untreated in 1997 and 1998 and kilned thereafter). They were sealed, labeled, and stored at or below -20°C.

Dissections for metals analysis were performed on pre-cleaned plastic trays that were changed between each replicate. Teflon-coated or plastic forceps were used for metals dissections. Results from laboratory blanks indicate that these procedures and instruments did not contaminate the samples (Appendix G-3). Instruments used in dissections were rinsed with an Alconox<sup>®</sup> detergent solution, rinsed

with tap water and then rinsed with NANOpure water between and during dissections. Dissected tissues were placed in new polypropylene jars that were rinsed with ultra-purified water, labeled, and stored at or below -20°C.

### 2.3.5. PRIORITY POLLUTANT ANALYSES

#### 2.3.5.1. Organic Analyses

All tissue samples and the sediment samples targeted for organic pollutant analyses were processed for extraction, isolation of pollutants, and final quantification of the extractable components. Tissue and sediment samples were processed and analyzed by a contract laboratory (ToxScan, Inc., Watsonville, California) in 1997 and 1998, and by City personnel in subsequent years. Sediment and tissue extracts were analyzed for polycyclic aromatic hydrocarbons (PAH), polychlorinated biphenyl (PCB) congeners, and the pesticide DDT and its analogs. PCB congeners are individual chemicals that are based on substitution of the biphenyl molecule with varying numbers of chlorine atoms. Forty-seven out of a possible 209 congeners were measured in 1997 and forty-two were measured in 1998. a fixed set of 45 congeners was measured annually from 1999 – 2008; since 2009, a set of 52 PCB congeners has been assessed annually. These 52 congeners were selected to closely parallel those analyzed in sediments and tissue in the RWQCB Regional Monitoring Program for Trace Substances (SFEI 1999a). All pollutants were extracted using U.S. EPA Method 3545 (Pressurized Fluid Extraction – PFE) and analyzed using a modification of U.S. EPA Method 1625, Rev. B – Semi-volatile Organic Compounds by Isotope Dilution GC/MS, (U.S. EPA 1993), using selected ion monitoring (SIM).

Reporting limits (RL) for each compound expressed in terms of dry weight and analyzed in sediment samples are presented in Appendix D-4. U.S. EPA Method 1664 was used to determine percent lipids in tissue samples, and U.S. EPA Method 160.3 for percent solids.

Detection limits for tissue samples are listed in Appendix G-2.

All of the Tetra Tech (1986b) recommended quality assurance/quality control (QA/QC) procedures were employed. Reference materials were processed and analyzed concurrently with each set of samples. Organic concentrations were measured in blanks, duplicates, and spiked samples.

#### 2.3.5.2. Trace Metals Analyses

Tissue and sediment samples were processed and digested according to the Tetra Tech modification of U.S. EPA Method 3050 with recommended procedures for monitoring priority pollutants in marine sediments and tissues (Tetra Tech 1986a, U.S. EPA 1983 and 1993). Sediment samples were analyzed by a contract laboratory (Frontier GeoSciences, Seattle, Washington) in 2000; in other years, they were processed by the City chemistry laboratory. Samples were mechanically homogenized, and aliquots were digested under a Class-100 laboratory hood using a wet oxidation technique (perchloric/nitric acid digestion) and trace metal clean techniques.

Elemental concentrations of aluminum (Al), arsenic (As), cadmium (Cd), chromium (Cr), copper (Cu), iron (Fe), lead (Pb), manganese (Mn), nickel (Ni), and zinc (Zn) were measured by emission spectroscopy using inductively coupled plasma instrumentation (ICP-AES) according to U.S. EPA Method 200.7. Mercury (Hg) concentrations were determined using atomic absorption spectroscopy (AAS) with cold vapor techniques. Selenium (Se) and silver (Ag) were determined using ICP-AES for sediment samples, and for tissue samples in 1997 and 1998. Since 1999, however, hydride generation AAS (U.S. EPA Method 270.3) was used to determine selenium concentrations in tissue samples, and graphite furnace AAS (U.S. EPA Method 272.2) was used to detect silver concentrations in tissue samples. The method detection limits for sediment samples are listed in Appendix D-6. Detection limits for tissue

samples are listed in Appendix G-4.

In 2006, MWH Laboratories (Monrovia, CA) assessed arsenic speciation in organism tissues – for total arsenic, As; trivalent arsenic, As(III); inorganic arsenic, As(In); monomethyl arsenic, MMA; and dimethyl arsenic, DMA; including pentavalent arsenic, As(V) determined as the difference between As(In) and As(III) – to determine contribution of organic, relatively non-toxic form of arsenic to total arsenic body burdens. Total arsenic (As) was determined using U.S. EPA Method 1638 (ICP-MS); all other forms were determined using U.S. EPA Method 1632 (AAS).

All of the Tetra Tech (1986b) recommended quality assurance/quality control (QA/QC) procedures were employed. Reference materials were processed and analyzed concurrently with each set of samples. Elemental concentrations were measured in digestion blanks, duplicates, and spiked samples.

## 2.4. DATA ANALYSIS

### 2.4.1. SUMMARY STATISTICS AND UNIVARIATE ANALYSES

#### 2.4.1.1. Sediment

Sediment grain size data were converted to phi ( $\phi$ ) units where  $\phi = -\log_2$  (diameter of the particle in millimeters) (Appendix D-1a,b). This transformation is widely used in sediment size analysis because it produces a normal distribution. Grain size data for each station were summarized to calculate skewness, kurtosis, the median, the mean, and the standard deviation of the mean (Appendix D-2). Inorganic and organic pollutant data were converted to dry weights.

#### 2.4.1.2. Biological Community Measures

The following community measures were calculated (using Primer v6) for each benthic infauna and trawl station: abundance (number

of individuals), species richness (number of species), Shannon-Weiner diversity index, Simpson's or Swartz's dominance index, and Pielou's evenness index. A detailed discussion of these indices is found in Appendix B.

Abundance of each species was measured as total count of individuals for combined 0.5 mm and 1.0 mm sieve of each 0.1 m<sup>2</sup> sample (Appendix E-2). Taxonomic community analyses and computations presented in this report include all taxa identified and enumerated, except that diversity analyses exclude higher or incompletely identified taxa whose inclusion would artificially inflate diversity (e.g. a few juvenile or incomplete specimens identified as *Glycinde* spp. would not be included in diversity calculations if numerous specimens of *Glycinde polygnatha* and/or *Glycinde* sp. SF1 were present at the same station). This conservative procedure can result in an underestimate of species richness, but avoids inflation that would occur by inclusion.

#### 2.4.1.3. Bioaccumulation

Statistical analyses were generally conducted using wet weight values; dry weights were used as noted. In statistical analyses, any sample that tested below detection limits for a compound was treated as though the compound level was equal to the detection limit. This assumption simplifies calculations and yields a conservative result.

In looking for patterns of causality, statistical comparisons were made using one- or two-tailed Student's T-tests, as noted, with unequal variance,  $\alpha = 0.05$ . Trends and correlations were determined using linear regression with  $\alpha = 0.05$ . Linear regressions used the Pearson product moment correlation coefficient. Tables of t-test results show probabilities.

#### 2.4.1.4. Mapping

Contour maps of physical grain size, phi size, sediment chemistry values and biological community measurements were plotted spatially

to show patterns in the study area using Surfer<sup>®</sup> for Windows 8.0 contouring and 3D surface mapping software by Golden Software, Inc. Figures 3-1, 5-6, and 5-7 were created using ArcMap for ArcGIS 10.1 by ESRI.

### 2.4.2. MULTIVARIATE ANALYSES

Detailed discussions of principal components analysis, principal coordinates analysis, cluster analysis, reference envelope analysis, and regression are found in Appendix B.

#### 2.4.2.1. Ordination and Cluster Analyses

Ordination and cluster methods are used to distinguish groups of entities (such as stations) according to similarity or dissimilarity of attributes (such as community composition or grain size parameters) (Tetra Tech 1982). Ordination analysis displays the sampling stations as points in a multidimensional space. The distance between the points in the space is proportional to the dissimilarity of the attributes found at the respective stations. The different dimensions of the ordination space, called axes, define independent gradients of change in the data. The axes are ordered so that the first axis defines a maximal amount of the change; the second axis defines a maximal amount of the remaining change, and so on for subsequent axes.

Cluster analysis defines groups of stations with similar community or grain size composition. The results are displayed in a hierarchical tree-like structure called a dendrogram. Stations that cluster together are more similar to each other than they are to stations in other cluster groups.

Principal components analysis (PCA) with varimax rotation (Dillon and Goldstein 1984) is an ordination technique applied to physical grain size and metals data at all sediment stations. The resulting axes define the main independent gradients of change in the sediment data. The axis scores were used as independent variables in multiple regression analysis with infauna and

trawl ordination scores.

Non-metric Multi-Dimensional Scaling (NMDS) analysis attempts to map the samples in 2 or 3 dimensional space in relation to their similarities of the Bray-Curtis index (Bray and Curtis 1957). The map is then compared to the PCA ordination of abiotic parameters such as sediment composition and chemistry to find patterns of similarity. A BEST (Biota and Environmental STEPwise) test can then define the relevance of any observed pattern (Appendix B.2.2.1). All taxa, including rare species, were included in the analysis and an overall square root transformation was done on the data prior to analysis. This reduces the stress coefficient (a measure of significance) of the plots and the effect of the rarer species on the distribution of stations on the NMDS plot.

In order to produce hypotheses concerning the possible causes of the community patterns identified by the ordination and cluster analyses, further analyses were performed to correlate the community patterns with sediment chemistry, grain size, Delta outflow data (IEP 2005) and oceanographic conditions (NOAA 2005). The community gradients defined by the ordination analysis often correlate with external environmental factors, which may be suspected of causing the associated benthic community changes. The relationships between the ordination axis scores and environmental parameters were observed by the ANOSIM (analysis of similarity) method from the Primer v6 program, which compares the similarity indices of the two distributions for statistical relevance. Parameters used in the ANOSIM analyses for benthic infauna were grain size, sediment organic content (Total PAHs and TOC and TKN) and trace metals (Clark and Gorley 2006).

#### 2.4.2.2. Reference Envelope Analysis

Reference envelope analysis involves comparing parameters at potentially impacted stations with the population of parameters at

individual reference stations and avoids the pseudoreplication (Hurlbert 1984) inherent in the common approach of using analysis of variance (ANOVA) to compare indicator variable means for reference and outfall locations (Smith 1995, 1998). When comparing an indicator value for a potentially impacted station with a population of reference indicator values, it is useful to compare the indicator value to a quantile on the tail of the population distribution (Smith 1995).

For the reference population, the actual value of the quantile of interest is usually unknown, but if appropriate sample data are available one could estimate the quantile (See Appendix B, 2.4. Reference Envelope). In this report, quantiles of 0.10 (for indicators expected to decrease with impact) or 0.90 (for indicators expected to increase with impact) and  $\alpha = 0.05$  were used for comparison with reference. These quantiles were subjectively chosen to balance between environmental protection (sensitivity to impacts) and avoidance of false indications of impact (Smith 1998). Using these parameters, values that exceed the tolerance interval bounds are a potential indication of impact because, 95% of the time, only 10% of the reference population would exceed them in the absence of an impact.

#### 2.4.2.3. BACIP Analysis

The general BACIP (Before-After-Control-Impact Paired) experimental design involves sampling at predetermined “control” and “impact” areas before and after the onset of the potentially impacting activity (Bernstein and Zalinski 1983, Stewart-Oaten et al. 1986). A change in indicator values at a potentially impacted location after the onset of the impacting activity does not necessarily indicate that an impact has occurred, since indicator values can change naturally over time. With this statistical design, it is assumed that large-scale environmental factors causing natural temporal changes in indicator values will have a similar effect in both the impact and control areas. Thus, the test for impact is a test for changes in the after-impact period that do *not* take place

in both the control and impact areas. The null hypothesis of the BACIP statistical test is that the average differences between impact and control sites will be the same in the before- and after-operational periods. The details of the sampling design can vary, although all credible designs should involve multiple sampling periods both before and after the impact. As a paired test, the comparison involves a single impact and a single control location. This is because a point source impact (such as an outfall) will create gradients of change in the vicinity of the impact, and the severity of the impact at different locations on the gradient is of interest rather than the impact to the larger area.

An assumption of the BACIP test is that the differences within each group are normally distributed. When the data are positively skewed as total abundances usually are, log transformation will make these differences more normal. Using the log is equivalent to testing for the same ratio of abundances before and after impact. Student's T test was used to evaluate the differences in the log values prior to discharge and after discharge.

The City and County of San Francisco began pre-discharge benthic infauna studies in 1982. Since that time, one outfall station (station 01) and one reference station (station 06) consistently remained part of the sampling program. A BACIP analysis of infauna abundance at these two stations was performed to provide some information on the degree to which the wastewater discharge may have affected organism abundance.



## **SECTION 3**

# **BEACH WATER QUALITY MONITORING**



# BEACH WATER QUALITY MONITORING PROGRAM

## 3.1. INTRODUCTION

The San Francisco Public Utilities Commission (SFPUC) and the San Francisco Department of Public Health (SFDPH) jointly administer the beach water quality monitoring program in San Francisco. Both agencies participate in sample collection; the SFPUC Microbiology Laboratory performs bacteriological analyses. The SFPUC is responsible for public notification when water quality does not meet State standards for water contact recreation, while the SFDPH is responsible for ensuring compliance with the California Sanitation, Healthfulness and Safety of Ocean Water-Contact Sports Areas Regulations, Title 17, California Code of Regulations.

Shoreline bacteria monitoring provides a measure of water quality conditions for near shore waters. Because measuring all pathogens is impractical, bacteria indicator organisms are used with the assumption that high numbers of bacterial indicators imply the presence of fecal contamination. The beach monitoring program is designed to monitor compliance with State standards for water contact recreation along recreational beaches in San Francisco (City) and to alert the public when water quality conditions exceed those standards. In addition, the public is alerted and water quality conditions are monitored when treated combined sewer discharges occur from the City's combined sewer system at locations where water contact recreation occurs.

Bacteria concentrations are measured along Baker Beach and China Beach on the City's north shore and along Ocean Beach, which extends along the western shore south to Fort Funston (Figure 3-1). This chapter provides a summary of those measurements spanning sixteen years.



Figure 3-1  
Shoreline sampling stations and combined sewer system discharge sites

### 3.1.1. REGULATIONS AND STANDARDS

The U.S. Environmental Protection Agency (U.S. EPA) strives through regulation to ensure that all waters of the nation are “fishable and swimmable” as required under the goals of the 1972 Clean Water Act. The California State Water Resources Control Board established water quality standards in the California Ocean Plan (SWRCB 2012) that apply to discharges into California territorial waters up to three miles from shore. Water quality objectives and standards were developed to ensure (1) the reasonable protection of beneficial uses (including water contact recreation) and (2) the prevention or detection of nuisance conditions.

The single sample maximum (SSM) standard (Table 3-1) is based on a variety of epidemiological studies from: (1) the U.S. EPA's draft “Implementation Guidance for Ambient Water Quality Criteria for Bacteria” that recommends enterococcus bacteria monitoring for marine systems (U.S. EPA 2002), (2) the U.S. EPA's “Water Quality Standards for Coastal Recreation Waters:

Using Single Sample Maximum Values in State Water Quality Standards” that addresses the appropriate use of SSM values for the presence of fecal contamination, specifically for the indicator organisms *Escherichia coli* (*E. coli*) and enterococci (U.S. EPA 2006) and (3) changes to the California Ocean Plan that provide standards for total and fecal coliform

Table 3-1

Single sample maximum criteria for the three bacteria indicators monitored

Indicator Bacteria	Single Sample Maximum Limit (per 100 mL)
Total Coliform	10,000
Fecal coliform ( <i>E. coli</i> )	400
Enterococcus	104

and enterococcus bacteria. The SSM standard is also in the California Department of Health Services regulations that implement Assembly Bill 411 (AB 411) for high-use public beaches with storm drains that discharge during dry weather. Although San Francisco beaches are not regulated under AB 411, use of this standard maintains consistency with other California beaches (U.S. EPA and RWQCB 1997). *Escherichia coli* and enterococcus bacteria were added to the beach monitoring program under the Oceanside Water Pollution Control Plant, Southwest Ocean Outfall, and Westside Wet Weather Facilities 2003 NPDES permit (Table 3-2).

### 3.1.2. SOURCES

Elevated bacteria concentrations at the shoreline may occur for a variety of reasons: naturally from soil bacteria, organic decay, marine mammals and birds, and storm or surface water runoff; or from anthropomorphic influences such as sewage, domestic pet waste, waste from boat holding tanks, ship or boat bilge-water purging, and street runoff (CDPH 2006).

Storm water runoff and street runoff are concerns the City has addressed through the development of the combined sewer system (CSS) control structures (see Introduction Section 1 and Appendix A). The CSS controls all dry weather flow including street runoff. All dry weather flow receives secondary-level treatment prior to discharge to receiving waters. Under normal wet weather conditions, all flow including storm water runoff is contained and treated prior to discharge. During intense rainstorms, treated combined sewer discharges may occur along the shoreline. The structures that store and transport wastewater and storm water are designed to allow settling of solids and to retain floatable materials, thus, shoreline discharges have received flow-through treatment equivalent to wet weather primary effluent. However, combined sewer discharges are not disinfected and may introduce waterborne pathogens into shoreline waters that could cause illness to those involved in water contact recreation. Treated combined sewer discharges that affect recreational beaches can occur from the Sea Cliff I pump station located at China Beach, the Sea Cliff II pump station located at

Table 3-2

Shoreline monitoring indicators and sampling frequency, 1997-2013

Time Period	Indicator(s) monitored			Minimum Monitoring Frequency
	Total coliform	<i>E. coli</i>	Enterococci	
June 1997 to September 2003	√			3 times per week
October 2003 to present	√	√	√	1 time per week

Baker Beach, or the Lincoln, Vicente, or Lake Merced Discharge structures located along Ocean Beach (Figure 3-1).

Lobos Creek is a contributor of bacteria on Baker Beach, but the source of bacteria in the Creek is not known. The creek is a spring brook about 0.5 mile long that is fed from an aquifer underlying the Richmond and Sunset districts and that originates as surface flow between 17<sup>th</sup> and 18<sup>th</sup> Avenues. It is densely vegetated and supports numerous wildlife species that might contribute bacteria loads. There are also about 30-35 homes that border the creek on the south (upper reach) and west (lower reach) banks. In addition, the City has two combined sewer crossings, one underground at 17<sup>th</sup> Avenue and one above ground at 22<sup>nd</sup> Avenue. Internal inspections and creek sampling above and below each crossing have indicated that they are not leaking. Extensive sampling has shown intermittent elevated bacteria levels throughout the creek, but has not identified a potential source.

### **3.2. PROGRAM DESCRIPTION**

The NPDES permit for the Oceanside Water Pollution Control Plant, Southwest Ocean Outfall, and Westside Wet Weather Facilities requires routine shoreline monitoring, and monitoring after a treated combined sewer discharge occurs. Routine shoreline monitoring consists of ankle-depth, surf zone grab samples once per week, year-round at seven ocean shoreline stations, and combined sewer discharge monitoring consists of sampling at those locations in closest proximity to the treated discharge. In addition to shoreline bacteria monitoring, the permit requires that recreational use surveys (which tally users by full, partial or non-water contact activities) be conducted after any combined sewer discharge. The public must also be notified whenever bacteria levels exceed State standards and whenever a combined sewer discharge occurs. Bacteria water quality data, beach-posting data, and recreational use data collected from 1997 to 2011 are available in previous reports (WQB 1998, 1999, 2000,

2001a, 2003a, 2003b, 2004; NRD 2006a, 2006b; NRLMD 2007, 2008, 2010a, 2010b 2011).

Seven sites are monitored weekly (daily when counts are elevated) at Baker Beach, China Beach, and Ocean Beach (Figure 3-1). Samples are analyzed for three different bacteria indicators of impaired water quality (total coliform, *Escherichia coli*, and enterococci). Results for samples collected are not available until the following day because of the time required to culture the bacteria to obtain an estimate of their presence. Beaches are posted and the public is notified 18 to 24 hours after an elevated concentration of bacteria occurs. This is done in case the elevated bacteria concentrations persist. In order to provide as rapid a response as possible the City proactively posts (and de-posts) beaches and makes public notifications based upon preliminary bacteria counts made available before final results are confirmed. The public is better served overall by timely notifications based upon preliminary counts than by the necessary delay needed to act upon confirmed counts.

#### **3.2.1. PUBLIC NOTIFICATION**

The City has implemented several public notification methods for dissemination of shoreline bacteria information. Permanent, large, yellow signs are installed along recreational beaches to explain that beaches will be posted if water quality is impaired. These information signs are written in English, Spanish, and Chinese (Figure 3-2).

##### **3.2.1.1. Recreational Beach Water Quality Hotline**

The current status of beach water quality in San Francisco is available on the SFPUC Recreational Beach Water Quality Hotline 415-242-2214 or 1-877-SFBEACH (toll-free). The hotline is updated whenever new sample results are available. The information provided includes the date and results (posted/not posted) of the most recent samples and additional information related to combined sewer system discharges as warranted.

### 3.2.1.2. Internet

The current status of beach water quality in San Francisco is also available on the internet at <http://beaches.sfwater.org>. The site uses color-coded symbols on a map of the City to provide an at-a-glance view of water quality status. Additional information including

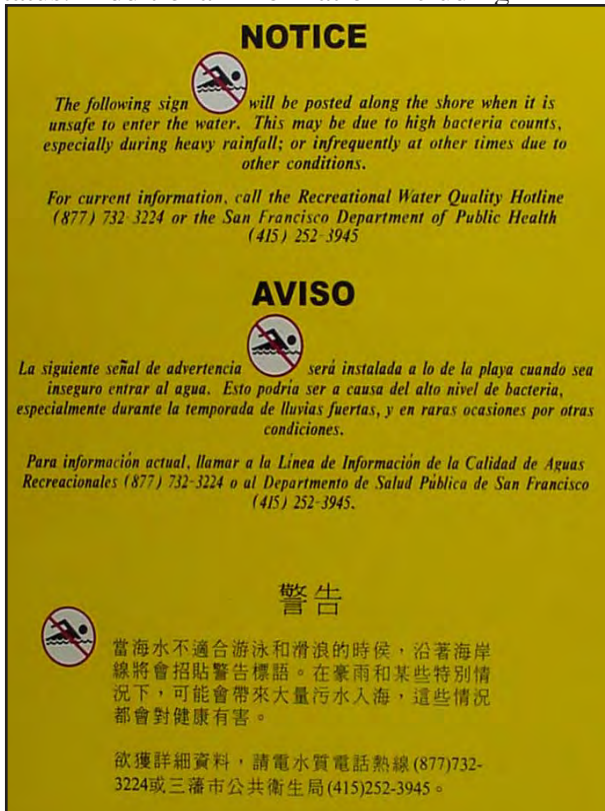


Figure 3-2

Permanent public notification sign located at major beach access points

monitoring program description, monitoring station locations, beach descriptions and photos, and data tables with recent sample results are available by clicking on the station symbols. The tables of sample results allow beach users to make an informed decision about water quality at San Francisco beaches.

### 3.2.1.3. E-mail

E-mail notifications are sent to affected agencies, user groups, and interested parties whenever a beach is posted or de-posted, including whenever a combined sewer discharge

occurs that affects a recreational beach. The public can subscribe to e-mail notifications of beach postings (and de-postings) at <http://beaches.sfwater.org>.

### 3.2.1.4. Beach Posting

In the event of a treated CSS discharge or when routine monitoring indicates that water conditions are not suitable for water contact recreation, beach access areas are posted with an international “No Swimming” sign (Figure 3-3). The “No Swimming” signs posted along Ocean Beach, located at beach entry points and at each traffic signal light stanchion, are easily visible to passing motorists on the Great Highway. Signs at Baker Beach and China Beach are located in parking areas and at trail entrances. The sign at Fort Funston is located near the entrance of the foot path. The ‘No Swimming’ signs are removed and the public is notified when bacteria concentrations indicate water conditions are within State standards for water contact recreation.

Table 3-3

Discharge structures and associated shoreline stations sampled when a discharge occurs

Combined Sewer System control structures	Stations
Seacliff I (China Beach)	17
Seacliff II (Baker Beach)	15, 15E, 16
Lincoln (Ocean Beach)	18,19, 20
Vicente (Ocean Beach)	20, 21, 21.1
Lake Merced (Ft Funston)	22

#### 3.2.1.4.1. Posting Due to a Combined Sewer Discharge

Whenever a treated CSS discharge occurs that affects Ocean Beach (including Fort Funston), China Beach, or Baker Beach the affected beaches are posted with “No Swimming” signs and samples are collected as soon as practical after a discharge occurs (Table 3-3). Beaches remain posted, recreational use surveys are conducted, and samples are collected daily until the discharge ceases and all three bacteria indicators are below State levels for

water contact recreation. The SFPUC web site <http://beaches.sfwater.org> is flagged with a flashing symbol to indicate the discharge location(s). The flashing symbol remains active until the discharge has ceased and all affected sample locations are below SSM standards. If elevated bacteria levels persist more than three days after the last discharge, the flashing symbol is removed and the station(s) remain(s) posted.



*Figure 3-3  
International no swimming symbol indicating that water quality does not meet recreational water quality standards*

### 3.2.1.4.2. Posting Due to Elevated Bacteria Counts

Because a single elevated indicator may be spurious and historical data indicate that such counts typically do not persist, a confirmation approach to posting beaches that lack sources of pollution (see Section 3.2.1.4.2.1) was adopted on July 1, 2007. For those beaches, confirmation is provided by a second elevated indicator in the same sample, an elevated indicator at a linked station (if applicable), or an elevated

indicator in a repeat sample. Linked stations are hydrologically connected such that, during routine monitoring, single indicator exceedances at both stations provide the necessary confirmation for posting.

1. Station not requiring confirmation to post: Baker Beach at Lobos Creek (station 15) is posted with “No Swimming” signs when one or more of the bacteria indicators are above the SSM standards for water contact recreation. Once posted, the station remains posted and is re-sampled daily until all three indicators fall below SSM standards for water contact recreation.
2. Stations requiring confirmation to post: Ocean Beach (stations 18 and 19 are linked, station 21.1), China Beach (station 17), Baker Beach (stations 15E and 16 are each independently linked to station 15). Stations in this category are only posted with “No Swimming” signs when one or more of the following confirmation criteria are met:
  - 1) More than one indicator exceeds its respective SSM at a single station.
  - 2) One or more indicator exceeds its respective SSM at a linked station on the same date.
  - 3) One or more indicator exceeds its respective SSM in a repeat sample.

If a single indicator exceeds SSM standards and confirmation criterion 1) or criterion 2) is not met, the station is re-sampled and evaluated using criterion 3). Once posted, stations remain posted and are re-sampled daily until all three indicators fall below SSM standards for water contact recreation. In contrast to routine monitoring, treated CSS discharges at these beaches are discrete events that require a separate response not involving confirmation (see Section 3.2.1.4.1).

### *3.2.1.4.2.1. Rationale for Confirmation before Posting*

The Beach Water Quality Workgroup of the California State Water Resources Control Board charged its Monitoring and Reporting Subcommittee with: 1) evaluating the existing monitoring program mandated by State law (AB 411), 2) reviewing the latest scientific findings regarding effectiveness and reliability of that program, and 3) recommending changes to increase public health protection and public notification for those engaged in ocean water contact activities. Although the beach water quality monitoring program conducted in San Francisco by the SFPUC and SFDPH is not regulated by AB411, implementing the recommendations of the Monitoring and Reporting (M&R) Subcommittee provides optimal information delivery to the public and will also assure that San Francisco's program is consistent with other programs throughout the State, to the extent practical. The Guidance Document prepared by the M&R Subcommittee and adopted by the Beach Water Quality Workgroup recognizes three types of beaches, each with distinct water quality issues:

- 1) Open Coastal Beaches with tidal flushing and unimpeded swell energy, with no known sources of contamination affecting water quality and a history of good water quality.

The M&R Subcommittee recommends employing a confirmation approach to posting this type of beach when a single sample standard is exceeded. Routine beach monitoring data have shown that San Francisco's ocean beaches have a history of good water quality, with no known source of contamination, and with sufficient circulation that the confirmation approach is appropriate. A single exception is indicated below. In contrast to routine monitoring, treated CSS discharges at these beaches are discrete events that require a separate response not involving confirmation.

- 2) Beaches with storm drain, creek, or river discharges during the summer.

The M&R Subcommittee recommends permanent posting for beaches within this category. In San Francisco, Lobos Creek entering the sea at Baker Beach represents a known or potential source of dry weather contamination. Confirmation is not required at this site. Permanent information signs are posted at this location.

- 3) Beaches in an enclosed harbor, bay, or estuary.

The M&R Subcommittee recommends an approach requiring "best professional judgment" for these beaches. None of the beaches covered in this report qualify as an enclosed beach.

### 3.2.2. RECREATIONAL USE MONITORING

San Francisco beaches are popular recreation areas used by the local and regional community and tourists throughout the year. Surveys that document the number of people at the beach provide an indication of the users potentially impacted by combined sewer discharges.

Under the terms of the 1997 NPDES permit the City was required to complete a comprehensive Recreational Use Study along Ocean Beach. The purpose of the study was to make an assessment of the number of water contact users along Ocean Beach and to determine the impact from treated combined sewer discharges on water contact recreation. The study was conducted over a two-year period from October 1998 through September 2000. Results from the study, contained in a separate report (WQB 2001b), determined that water contact and non-water contact (including surf fishing) recreational activities along Ocean Beach are extensive. Of the 154,054 people observed during the two-year study, the majority of users (83%) were involved in non-water contact recreation; and of those involved in water contact recreation, up to 25% were surfers. The number of users observed participating in water contact recreation following a combined sewer discharge represented less than one percent of all water contact users observed during the study. The two-year investigation concluded that most combined sewer discharges events occur in mid-

winter and have little impact on recreational use, as little use was observed during the cold, short days of winter. Isolated combined sewer discharge events that occur in early spring have the potential to impact more users as beach use increases when days become longer and the duration of storm events are shorter, contributing to good surfing conditions.

### 3.3. RESULTS AND DISCUSSION

#### 3.3.1. BACTERIA DATA

Over 80 percent of San Francisco's rainfall occurs between November and March (Null 1995). Most of the data analyses presented here are from July 1 through June 30 in order to correspond with the rainfall year rather than the calendar year. Data in this section is presented for the full 2012-2013 rainfall season, as well as for the combined survey years of 2003 to 2013 (1997 to 2013 for CSDs).

##### 3.3.1.1. Survey Year 2012-2013

Bacteria data from July 2012 to June 2013 are presented in Appendix C-1. Data for prior years are available in earlier reports referenced above. Bacteria sampling from July 2012 to June 2013 resulted in the collection of 468 samples from the ten locations in Figure 3-1: 215 along Ocean Beach (including Fort Funston), 54 at China Beach, and 199 along Baker Beach.

3.3.1.2. The SSM standard was exceeded in 56 samples during the year: the total coliform bacteria standard was exceeded 16 times, the *E. coli* standard was exceeded 31 times and the enterococcus standard was exceeded 50 times (Appendix C-2(a-e)). Thirty-nine of these samples were associated with wet weather and treated combined sewer discharge events. Of the 17 samples not associated with wet weather, 9 (53%) occurred at Baker Beach and 7 of those were at station 15 where Lobos Creek drains into the Pacific Ocean. Of the remaining eight elevated counts during dry weather, three occurred at China Beach, and five at Ocean

Beach, where beach sources (such as pet waste, organic decay, etc.) or ground water were the most likely cause of the elevated counts. The majority (76%) of elevated bacteria counts (including those associated with treated CSS discharges) dropped below the level of concern for water contact recreation within 24 hours. Station 15 (Baker Beach at Lobos Creek) remained elevated for 96 hours, stations 15E (Baker Beach East) and 16 (Baker Beach West) for 72 hours and stations 18 (Ocean Beach at Balboa) and 21.1 (Ocean Beach at Sloat Boulevard) for 48 hours after CSS discharges in November 2012. Station 19 (Ocean Beach at Lincoln Way) remained elevated for 48 hours after a CSS discharge in December 2012. Elevated bacteria counts not related to treated CSS discharge include station 15E (Baker Beach East) which remained elevated for 48 hours in September 2012 and stations 16 (Baker Beach West) and 21.1 (Ocean Beach at Sloat Boulevard) for 48 hours in October 2012 (Appendix C-2(a-e) and C-6-C15) Survey Years 2003-2013

Monitoring for all three bacteria indicators (total coliform, *Escherichia coli*, and enterococci) began in October 2003; prior to that date only total coliform bacteria were measured (Section 3.1.1 Regulations and Standards).

The SSM standard was exceeded in 189 samples during the last 4 years (July 2009-June 2013, Appendix C-2(a-e)) since the previous Summary Report (NRD 2010a). Eighty-three of these samples (56%) were associated with wet weather or treated combined sewer discharge events. Of the 65 samples not associated with wet weather, fifty occurred at Baker Beach. These 50 elevated counts were primarily due to outflow from Lobos Creek, which drains into the Pacific Ocean. The remaining seven samples were due to unknown sources. The total number of days when there were elevated bacteria counts not associated with wet weather or a treated discharge(s) at one or more ocean beaches was eight days during the four-year period.

Venn diagrams (Chow and Rodgers 2005)

were developed to assess the degree of overlap in SSM exceedances among the three indicators (October 2003-June 2013). Enterococcus was the indicator that was most frequently exceeded followed by *E. Coli* and total coliforms at all stations (Fig 3-4 (a)). The enterococcus threshold was exceeded ten times more than *E. Coli* and total coliforms thresholds. Enterococcus alone accounted for 189 (44%) of the standards failures, all three bacteria thresholds were collectively exceeded 79 times (19%) and *E. Coli* and enterococci combined for 92 (22%). The remaining threshold exceedences comprised of *E. Coli* 16 (4%), total coliform 20 (5%), total coliform and enterococci 28 (7%) and total coliform and *E. Coli* 2 (1%) thresholds. When comparing all stations, 15 (Baker Beach at Lobos Creek) and 19 (Ocean Beach at Lincoln Way) had the highest number of exceedances (143 and 55, respectively) (Fig 3-4 (b,c)). The enterococci standard was exceeded two times as often as any other standard at station 15. Station 19 had twenty-one enterococci exceedances followed by 17 for all three bacteria thresholds. All of the exceedances at station 19, except one dry-weather enterococci and one *E. Coli* exceedance, were due to treated CSS discharges and rain.

### 3.3.2. TREATED COMBINED SEWER DISCHARGES

During periods of heavy rainfall, combined flows may exceed the capacity of the City's CSS and treated discharges onto beaches can occur from discharge structures. A treated combined sewer discharge event is defined as a discharge from the CSS through one or more discharge structures as a result of rainfall. To be considered a discrete event discharges must be separated by six hours in time from any other discharge.

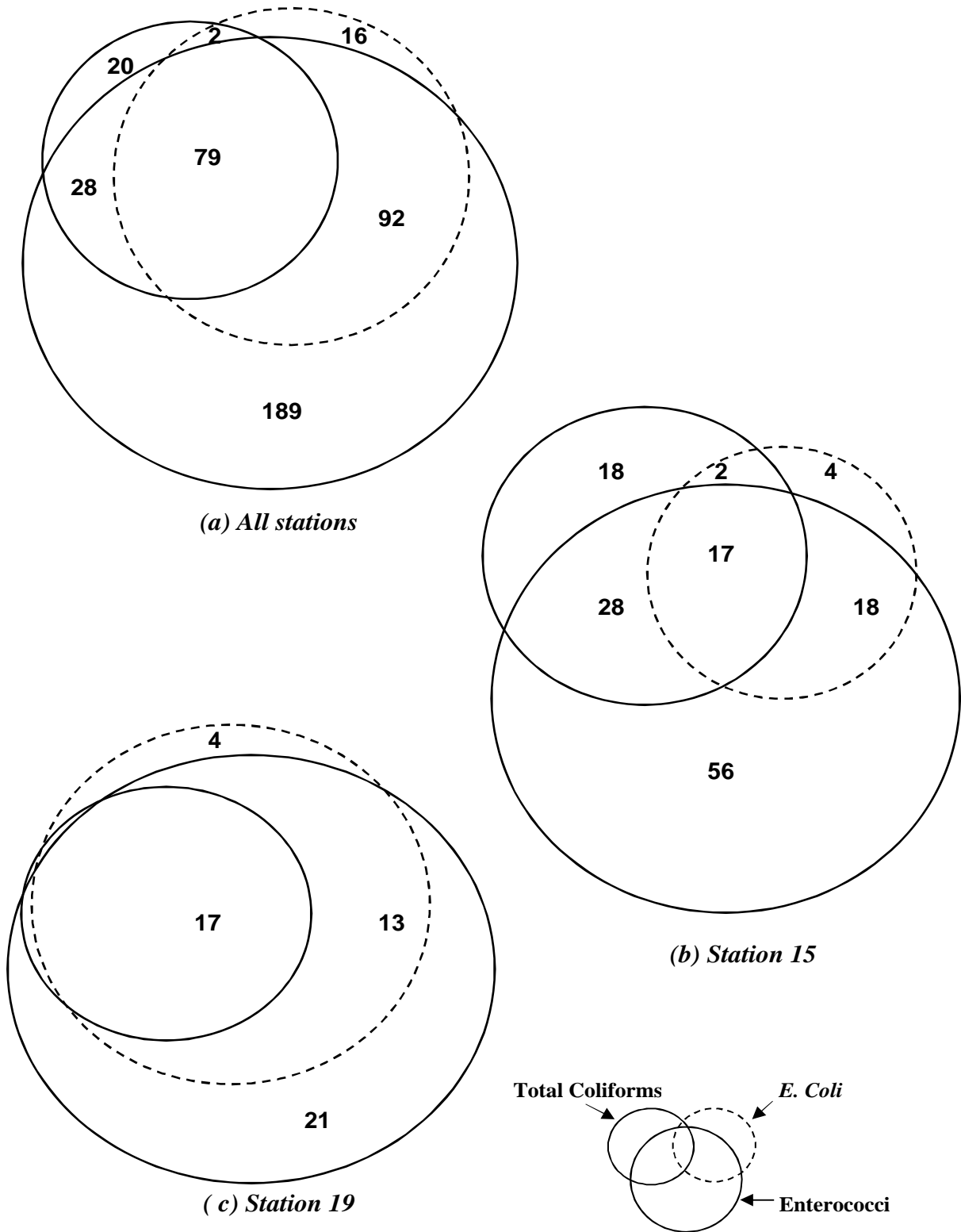
San Francisco's Combined Sewer System (CSS) is unique in coastal California. In addition to normal sanitary, commercial, and industrial wastewater flows treated by separate sewer systems, a CSS collects and treats storm water. This offers significant environmental

benefit because both storm water and urban street runoff is captured and treated. All street runoff during dry weather receives full secondary treatment, most storm flows receive full secondary treatment, and all storm flow receives treatment to at least wet-weather primary effluent equivalence before being discharged through a designated outfall. During heavy rain events, treated effluent typically comprised of 94% treated storm water and 6% treated sanitary flow can discharge into coastal waters through the CSS. A system of underground storage, transport, and treatment boxes handles major rain events, minimizing the number of combined sewer system discharges. Best management practices are implemented to maximize storage and treatment and minimize shoreline discharges.

#### 3.3.2.1. Survey Year 2012-2013

Average historic rainfall (1948-2013) is 19.99 inches for San Francisco near Oceanside (WRCC 2013). From July 2012 to June 2013 there were 55 days of recorded rainfall, totaling 19.7 inches, which is 99% of average historic rainfall (Appendix C-3). Eight discrete discharges occurred during the rainfall year. The number of Combined Sewer System discharges generally corresponds to the duration and number of rainfall events. Over 14 inches of rain (71% of total) occurred in the months of November and December; all of the discharges occurred during this same period. Six combined sewer discharge events occurred at the Lincoln, Vicente and Lake Merced discharge structures, and three at Sea Cliff I and II pump stations (Appendix C-3). December had the heaviest rainfall with 8.75 inches over 16 days. Five of the eight CSS discharges that occurred during the year took place in December. The other three events occurred in November, which had 5.98 inches on 10 days. The 2012-2013 rainy season was longer than any other in recent record, beginning in October and lasting until the end of June (the rainy season typically ends by the second week of April). No dry weather discharges occurred from any discharge





Zero values are not indicated

Figure 3-4

Venn diagrams showing the number of SSM threshold exceedances for each indicator alone and in combination October 2003-June 2013. These diagrams illustrate that enterococci account for the majority of threshold exceedances at (a) All stations combined; and at the two stations with the greatest number of exceedances (b) Baker Beach at Lobos Creek; and (c) Ocean Beach at Lincoln Way.

structures in 2012-2013 (Appendix C-3).

### 3.3.2.2. Survey Years 1997-2013

Prior to construction of the transport and storage structures, over 80% of wet weather flows were discharged untreated at the shoreline as combined sewer overflows. Rainfall at a rate of 0.02 inches per hour regularly resulted in untreated combined sewer discharges to the City’s shoreline. From November 1993 until completion of the Richmond Transport in January 1997, there were over 150 untreated combined sewer discharge events reported from the Sea Cliff II pump station that were uncontrolled and did not receive treatment. Since completion of the final CSS control structures in 1997 (Westside Transport in 1986, Lake Merced Transport in 1994 and Richmond Transport in 1997), all wet weather flows, including treated combined sewer discharges, receive the minimum equivalent of wet weather

primary treatment before discharge to the receiving waters and the number of discharges has been greatly reduced (Figure 3-5). The completed CSS captures and provides treatment to 100 percent of all rainstorms. The majority of storm water flow is discharged offshore through the SWOO. Treated combined sewer discharges, which result from less than ten percent of the City’s rainstorms, are composed on average of 6% sanitary flow and 94% stormwater.

There are relatively few instances of elevated or persistent high counts as a result of treated combined sewer discharges. Of the 117 discrete combined sewer discharge events that have occurred from 1997-2013, only 48 resulted in an elevated bacteria count (Table 3-4). All bacteria counts resulting from a discrete combined sewer discharge event on Baker Beach (stations 15, 15E and 16), Ocean Beach (stations 18, 19, 20, 21, 21.1 and 22) and China Beach (station 17) dropped below the level of concern for water

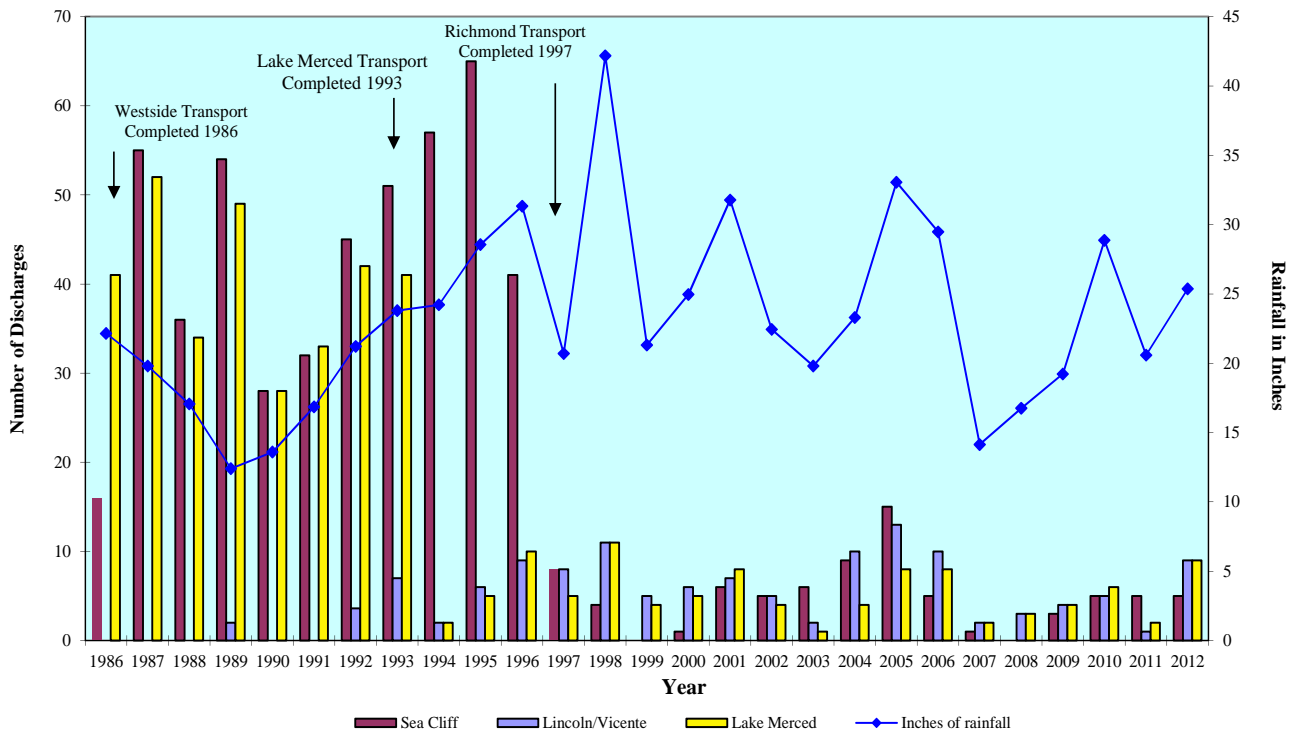


Figure 3-5

Rainfall and westside combined sewer discharges showing reduction in number of discharges with completion of infrastructure improvements in 1997. No data available for Lincoln/Vicente from 1986 to 1988; rainfall data from Golden Gate Weather Services (ggweather.com)

*Table 3-4  
Summary of total treated combined sewer discharges  
and discharges with elevated bacteria counts  
July 1997-June 2013*

	Seacliff I	Seacliff II	Lincoln/ Vicente	Lake Merced	Discrete CSDs
Discharges with elevated bacteria counts	4	17	53	15	48
Total Discharges	28	67	162	83	117

*Table 3-5  
Summary of rainfall and related discharge events  
1997-2013*

Wet Weather Season (July 1 - June 30)	Rainfall (inches)	Number of treated CSD events					Number of Discrete CSD Events*	
		Sea Cliff I Pump Station	Sea Cliff II Pump Station	Lincoln Structure	Vicente Structure	Lake Merced Structure		
1997-1998	41.14	2	10	13	13	10	14	
1998-1999	18.86	0	0	7	7	6	7	
1999-2000	23.19	1	1	6	6	5	7	
2000-2001	13.76	2	2	0	0	2	3	
2001-2002	24.40	1	1	6	6	6	6	
2002-2003	22.25	1	7	6	6	5	9	
2003-2004	18.77	2	8	4	4	4	8	
2004-2005	26.20	5	8	6	7	7	12	
2005-2006	31.83	3	9	9	9	11	13	
2006-2007	14.76	0	2	1	1	2	3	
2007-2008	18.37	0	1	4	4	4	4	
2008-2009	18.29	0	1	4	4	4	4	
2009-2010	25.80	6	7	3	3	4	7	
2010-2011	30.06	0	4	4	4	5	7	
2011-2012	17.56	2	3	2	3	3	6	
2012-2013	19.70	3	3	6	6	6	8	
Average (July 1997- June 2013)	22.81					Long Term Average		<b>7</b>
						Expected performance based on design		<b>8</b>

\*A discrete discharge event is separated by six hours in time from any other discharge.

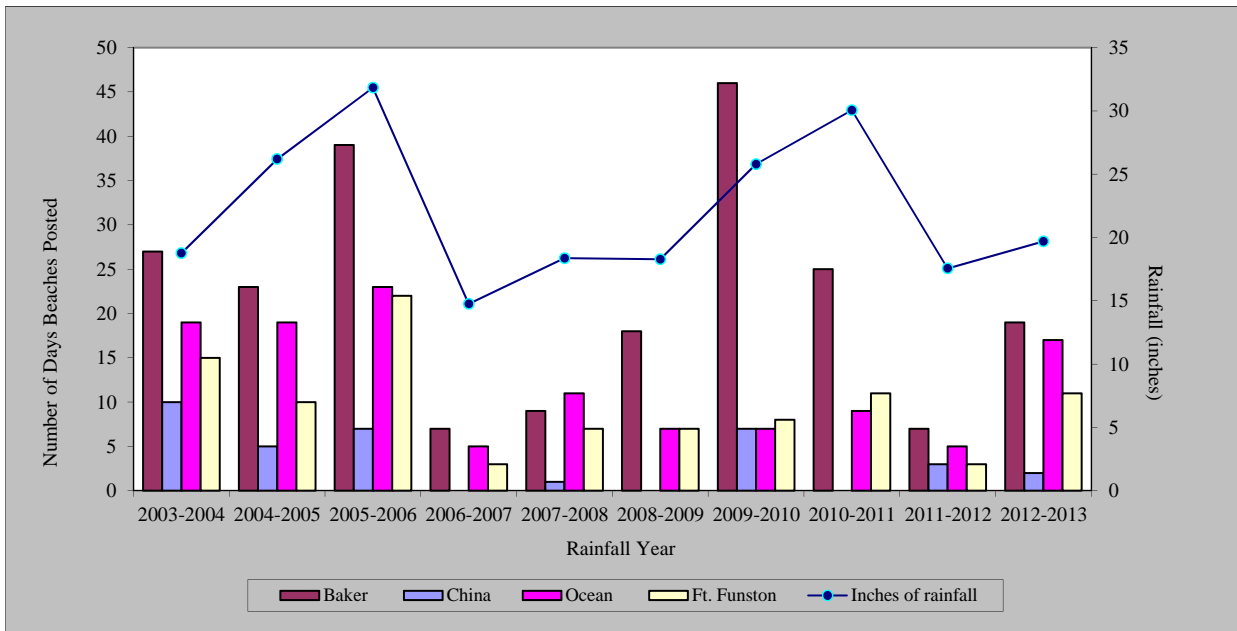


Figure 3-6  
Annual rainfall and number of beach postings July 2003-June 2013

Table 3-6  
Number of beach postings due to elevated bacteria counts or treated combined sewer discharges 2003-2013

Rainfall Year	Baker Beach		China Beach		Ocean Beach		Ft. Funston		Rainfall (inches)
	Elevated	Discharge	Elevated	Discharge	Elevated	Discharge	Elevated	Discharge	
2003-2004	12	15	3	7	6	13	3	12	18.77
2004-2005	9	14	0	5	4	15	0	10	26.2
2005-2006	25	14	1	6	4	19	0	22	31.83
2006-2007	3	4	0	0	2	3	0	3	14.76
2007-2008	7	2	1	0	0	11	0	7	18.37
2008-2009	16	2	0	0	0	7	0	7	18.29
2009-2010	35	11	0	7	0	7	0	8	25.8
2010-2011	20	5	0	0	0	9	0	10	30.06
2011-2012	3	4	2	1	0	5	0	3	17.56
2012-2013	11	8	0	2	1	14	0	12	19.7
average days posted/year	14	8	1	3	2	10	0	9	
Average percent of time beach available for water contact recreational use	94		99		97		98		

contact recreation within 48 hours.

A summary of rainfall and related discharge events from each structure is presented by wet weather season in Table 3-5. The wettest rainfall years were 1997-1998 (El Niño/La Niña years) with 41.14 inches of rain (207 % of

normal) and a correspondingly high number of combined sewer discharges (14), 2005-2006 with 31.83 inches of rain (159% of normal) and 13 combined sewer discharges and 2010-2011 with 30.06 inches of rain (152% of normal) and seven combined sewer discharges. The 2000-

2001 rainfall season had the least rain with 13.76 inches (68% of normal) and one of the least number of combined sewer discharges (3). The 2006-2007 season also had only three combined sewer discharges with 14.76 inches (74% of normal) of rain. Notwithstanding the substantial differences in rainfall over the last 12 years, the City has met the long-term shoreline discharge system design average of eight combined sewer discharge events per year (established as the Westside design goal by the evaluation of cost/benefit analysis) since completion of the final CSS control structures in 1997. The relationship of rainfall, combined sewer discharges, and enterococci counts are presented in graphic form in Appendix C6-C15 for rainfall years 2008-2009 through 2012-2013.

### 3.3.3. BEACH POSTINGS AND RECREATIONAL USE MONITORING

#### 3.3.3.1. Survey Year 2012-2013

One or more beaches were posted a total of 27 days; 17 of the posting days were due to treated combined sewer discharge events and 11 were due to elevated bacteria counts (Appendix C-4). One of the posting days was due to both a treated combined sewer discharge at Ocean Beach and elevated bacteria counts at Baker Beach at Lobos Creek. Ocean Beach (including Fort Funston) was posted for a total of 17 days, which were coincident with combined sewer discharge events. China Beach was posted for three days due to two discharge events at Seacliff I pump station. Baker Beach was posted a total of 27 days; eight days due to three discharge events at the Sea Cliff II pump station, and 11 days due to elevated counts. Of the eleven days, eight occurred at station 15 where outflow from Lobos Creek was the likely cause of the elevated counts leading to the beach postings.

Recreational use data associated with discharge events (Appendix C-5) showed a higher than average use for partial water contact (fishing) (7% vs. 3%), a lower average for water contact use (3% vs. 6%) and the same average use for non-water contact (91%) compared to

those averages associated with discharge events from the extensive Ocean Beach Recreational Use Study (WQB 2001b).

#### 3.3.3.2. Survey Years 2003-2013

The number of postings has remained fairly consistent during the last six-year period, (Fig. 3-6) with the variability in number of postings being due to the variability of rainfall and consequent discharges. The exception occurred at Baker Beach, primarily at Lobos Creek, during the 2009-2010 season. The number of beach-specific postings per rainfall year is shown on Table 3-6; postings are separated as a result of either treated combined sewer discharges or elevated bacteria counts (over the SSM). Postings due to elevated counts increased after *E. coli* and enterococci results were added to the analyses in October 2003. Postings due to elevated counts at Ocean Beach decreased after the confirmation approach was adopted in July 2007. Ocean Beach and Fort Funston were in compliance with water contact recreational use standards an average of 97% and 98% of the time during the last ten-year period; China beach 99%, and Baker Beach 94%.

### 3.4. SUMMARY AND CONCLUSIONS

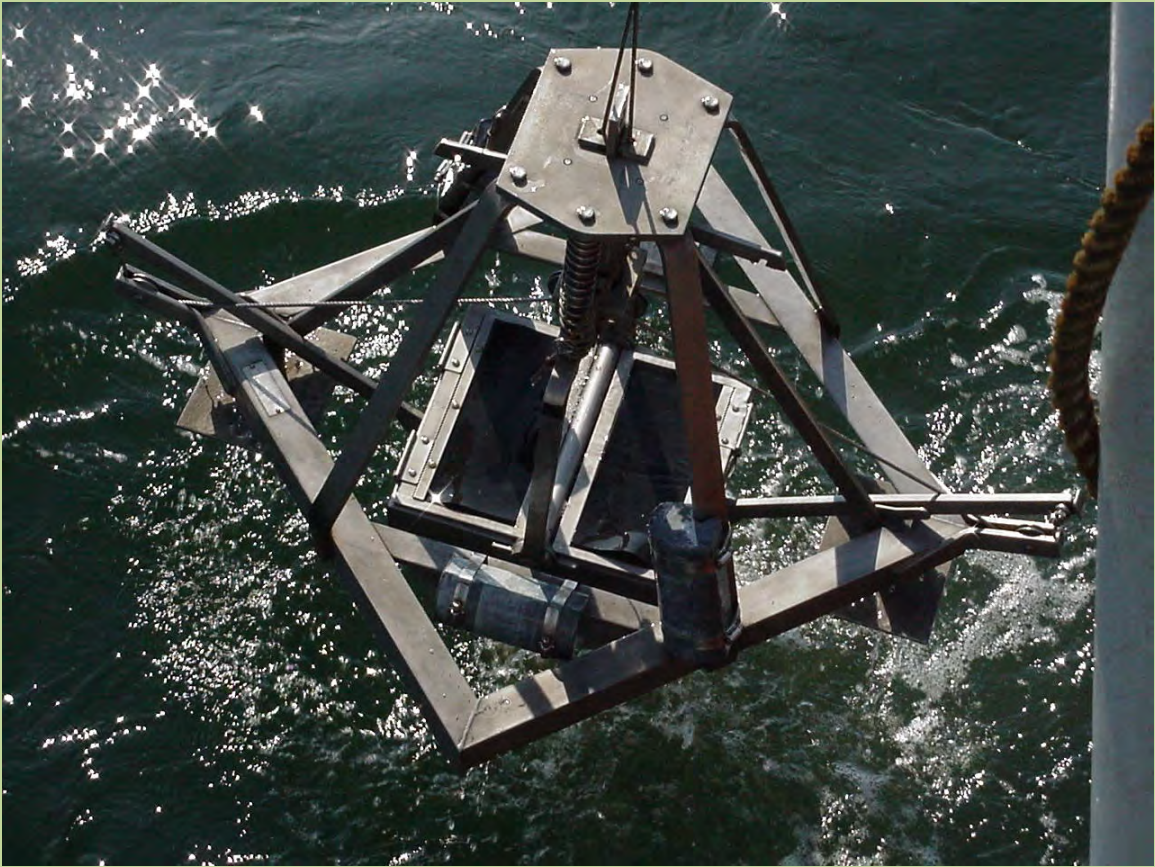
The completion of all Westside combined sewer infrastructure improvements in 1997 has resulted in fewer combined sewer discharges and improved near shore water quality along San Francisco's ocean beaches. Beach water quality monitoring data over the past 16 years continue to verify that the City's capital improvements to the CSS are successful in treating and controlling combined sewer discharges and protecting beach water quality. Recent studies (CDPH 2006) indicate that storm water is a major source of near shore water pollution, and the City's CSS provides some level of protection for the health of recreational beach users by treating and transporting the majority of storm water runoff offshore through the SWOO.

The addition of *Escherichia coli* and enterococcus analyses in 2003, while adding to the number of postings in San Francisco, is

likely more protective of public health. This is especially the case when elevated counts in dry weather are likely originating from sources other than the wastewater system.

San Francisco beaches provide activities for thousands of people throughout the year with most people visiting in the spring and fall months. The majority of beach users engage in non-water contact activities. Surveys indicated that treated combined sewer discharges in the middle of winter affect few users, since these discharges were typically associated with unpleasant weather conditions.

Sixteen years of monitoring data show that there continues to be a relationship between rainfall, elevated shoreline bacteria counts, and treated combined sewer discharges. Rainfall and combined sewer discharge events typically only cause short-term periods when bacteria concentrations exceed water contact standards.



## **SECTION 4**

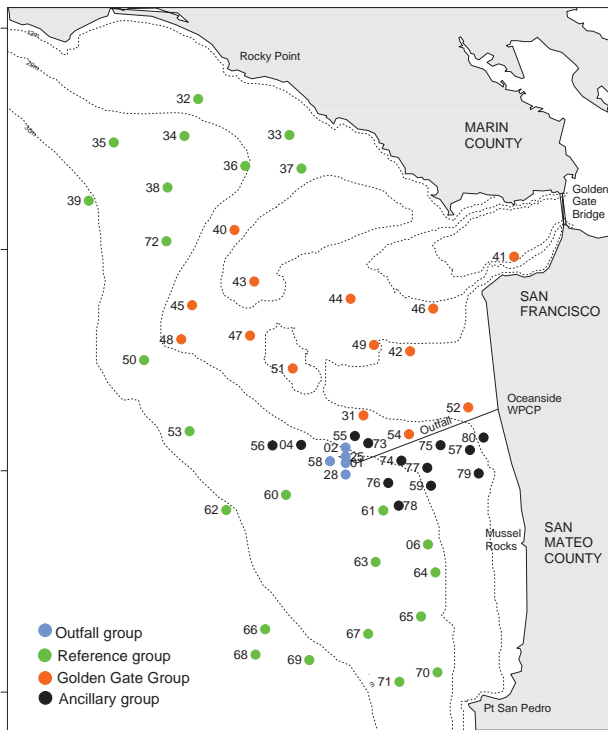
### **MARINE SEDIMENTS**

# MARINE SEDIMENTS

## 4.1. INTRODUCTION

Sediment monitoring program is used to assess the presence, magnitude, and spatial extent of changes from the South West Ocean Outfall (SWOO) to sediment characteristics, and how they relate to or affect the surrounding environmental community in the near shore waters of the Gulf of the Farallones (Figure 4-1).

Sediment grain size, and organic, and inorganic priority pollutant levels in sediments monitored from 1997 to 2012 are discussed in this section.



*Figure 4-1  
Sediment stations showing regions of the study area: outfall, reference, Golden Gate, and ancillary*

## 4.2. RESULTS AND DISCUSSION

### 4.2.1. GRAIN SIZE

#### 4.2.1.1. Survey Year 2012

Results of 2012 grain size analysis are presented in Appendix D-1 Grain sizes are expressed as percentages of pebble, granule, sand, and combined silt and clay. Sand is further categorized into coarse, medium, fine, and very fine grades (Table 4-1). Phi size is a numerical representation of the particle sizes in each grain size category (see Methods, 2.4.1.1 Sediment).

#### 4.2.1.2. Survey Years 1997 -2012

The standard deviation (SD) of grain sizes indicates how well sorted the sediments are at each station: a high SD corresponds to a wide size distribution and indicates poorly sorted sediments, and characterizes stations with relatively high percentages of silt and clay (e.g., reference stations 36, 35, 34, 32 and outfall station 28), while a lower SD corresponds to a lower spread in size distribution and indicates well-sorted sediments, and characterizes stations with high percentages of fine sand (e.g., stations 41, 43, and 80). Figure 4-2 shows the average distribution of well-sorted (low SD) versus poorly sorted (high SD) stations from 1997 to 2012. The degree of sorting depends on sediment type; fine sands are generally well sorted, while sorting becomes progressively poorer as the percentages of silt, clay, or gravel increase (Folk and Ward 1957). Although a wide range of grain sizes are found in the study area, the low percentages of silt/clay and gravel characterize a generally well-sorted sedimentary environment. The majority of the study area is characterized by fine to very fine-grained sand (phi sizes 2 to 4).

Average grain size distribution (Figure 4-3) summarizes the patterns for each year analyzed separately (WQB 1998, 1999, 2000, 2001a,



Table 4-1  
 Classification of sediment grain size categories used in the study.

Description	Size (mm)	Phi size (-Log <sub>2</sub> mm)	Comments
Gravel	> 2	≤ -1	Includes granule, pebble, and shell fragments
Coarse Sand	> 0.5 to 2	> -1 to 1	Includes very coarse sand
Medium Sand	>0.25 to 0.5	> 1 to 2	
Fine Sand	>0.125 to 0.25	> 2 to 3	
Very Fine Sand	>0.0625 to 0.125	> 3 to 4	
Silt/Clay	> 0.0039 to 0.0625	> 4	Combined Silt and Clay

2003a, 2003b, 2004: NRD 2006a, 2006b, NRLMD 2007, 2008, 2010, 2011, 2012). An area extending west from the Golden Gate Bridge and surrounded by the barrier sand bars, contained the highest average percent of coarse to medium sand. In 2004 and 2005 more

than 40 large sand waves were mapped using a multibeam sonar system, in this area, and grain size measurements indicate that the bed surfaces of the sand waves are composed of primarily coarse sand and gravel (Barnard, et al., 2006). The influence of strong diurnal tides in and out

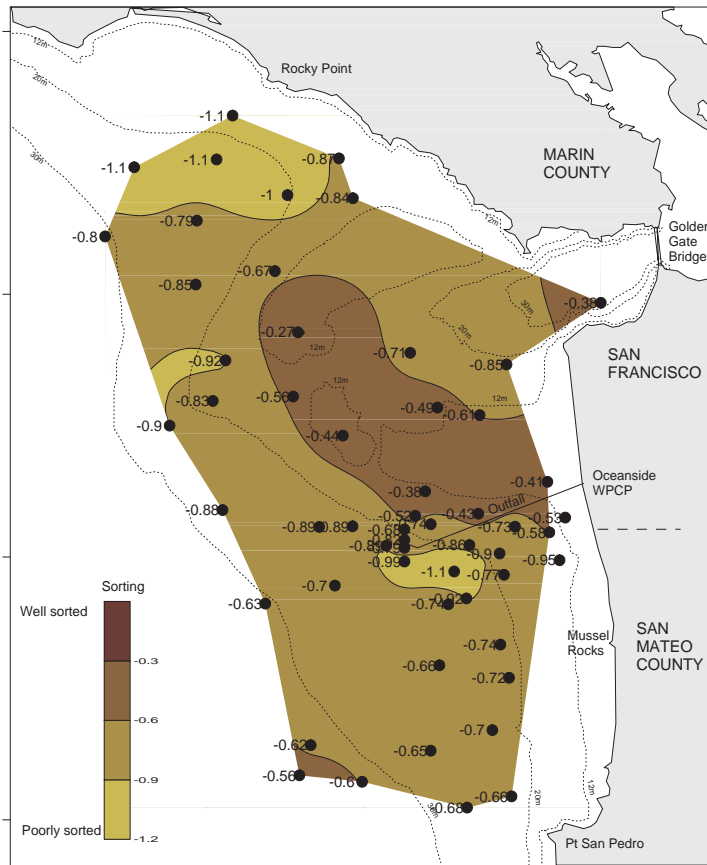
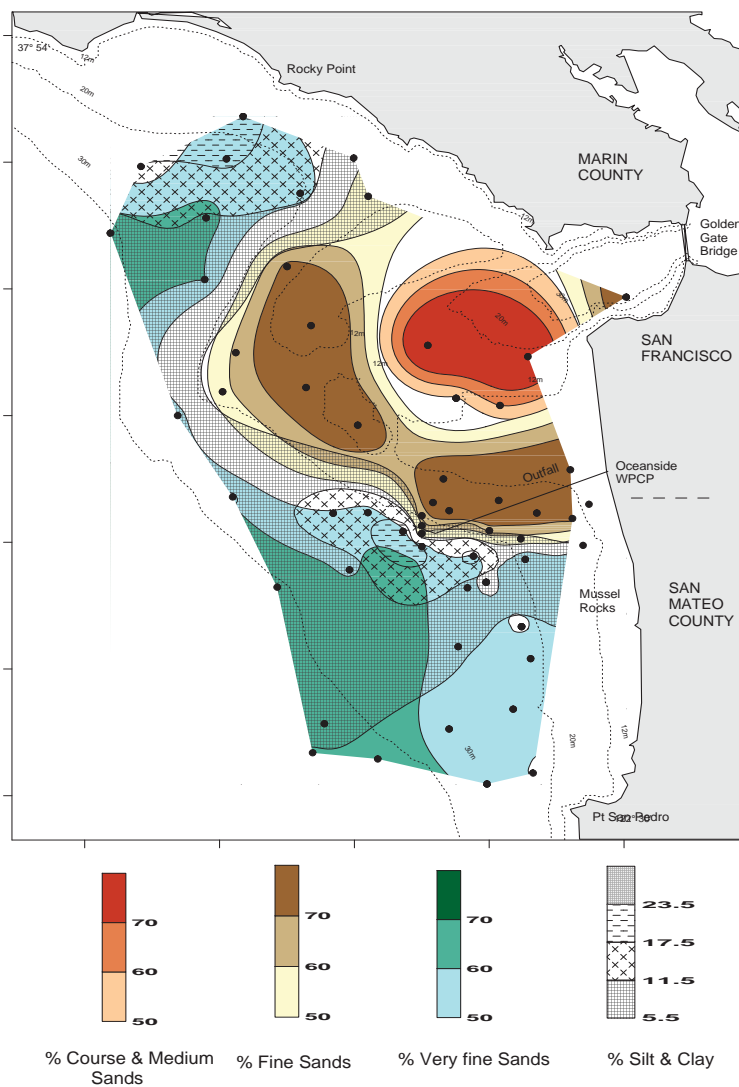


Figure 4-2  
 Average sediment sorting from 1997 to 2012

of San Francisco Bay and major storm events (Dean and Gardner 1995) act to rework the sediments in this area and may contribute to this pattern of coarser sand. Fine to very fine sands have been the dominant grain size fractions at most stations in the study area from 1997 to 2012 (Figure 4-3). Fine sands predominate on the sand bars. Greater percentages of very fine sands were evident seaward of the sand bars. Beginning in 2007 there has been a noticeable pattern of increasing fines (% silt/clay) at

reference locations in the northern reach of the monitoring area (offshore of Rocky Point) as well as near the SWOO outfall (Figures 4-4a & b). Barnard et al. (2012) attribute the observed fining of sediments in these areas to a decrease in aerial extent of the ebb-tidal delta of the San Francisco Estuary, related to a recent (1999) step-decrease in suspended sediment concentrations observed inside San Francisco Bay. Barnard et al. (2012) also speculate that the origin of the fine sediment observed in the outer reaches of the ebb-tidal delta is either finer shelf sediment or finer bay-derived sediment.



*Figure 4-3*  
Average grain size distribution from 1997 to 2012

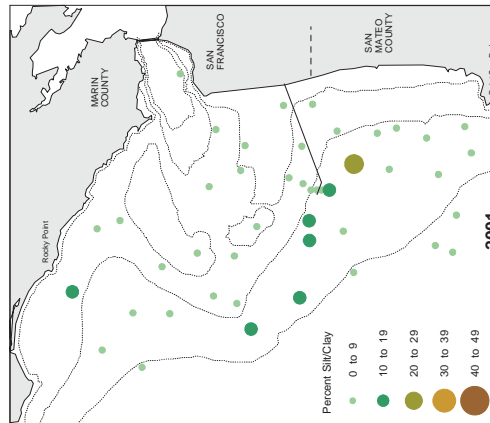
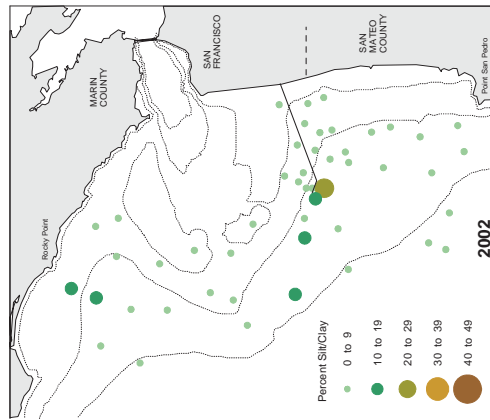
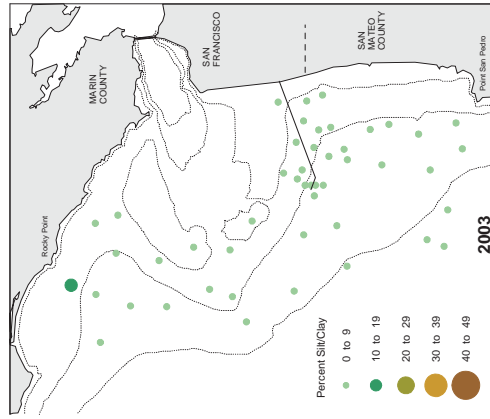
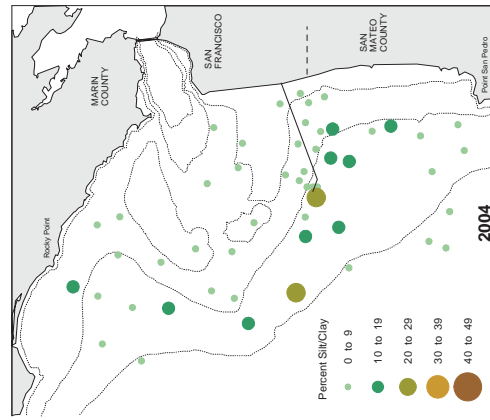
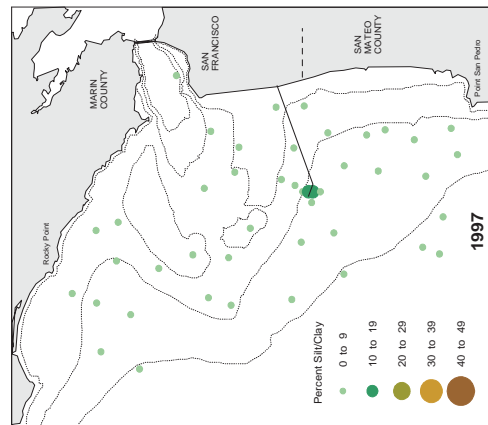
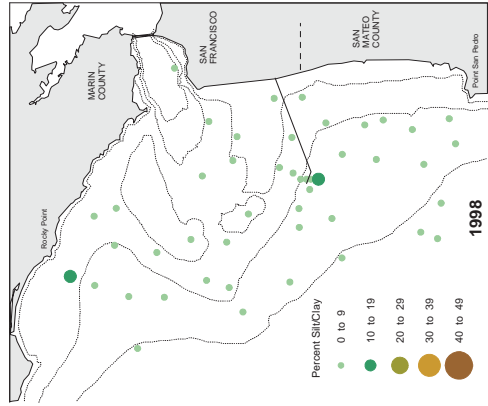
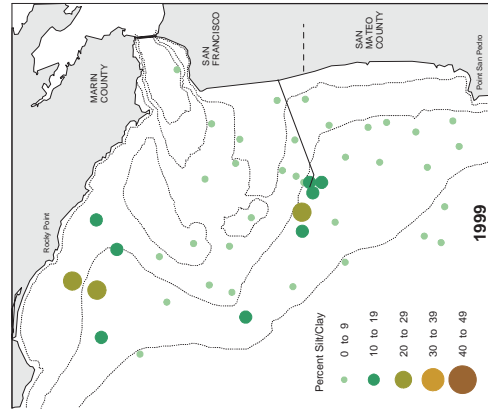
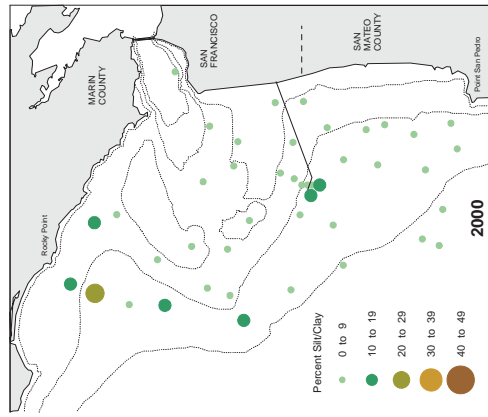


Figure 4-4a  
Percent silt and clay from 1997 to 2004

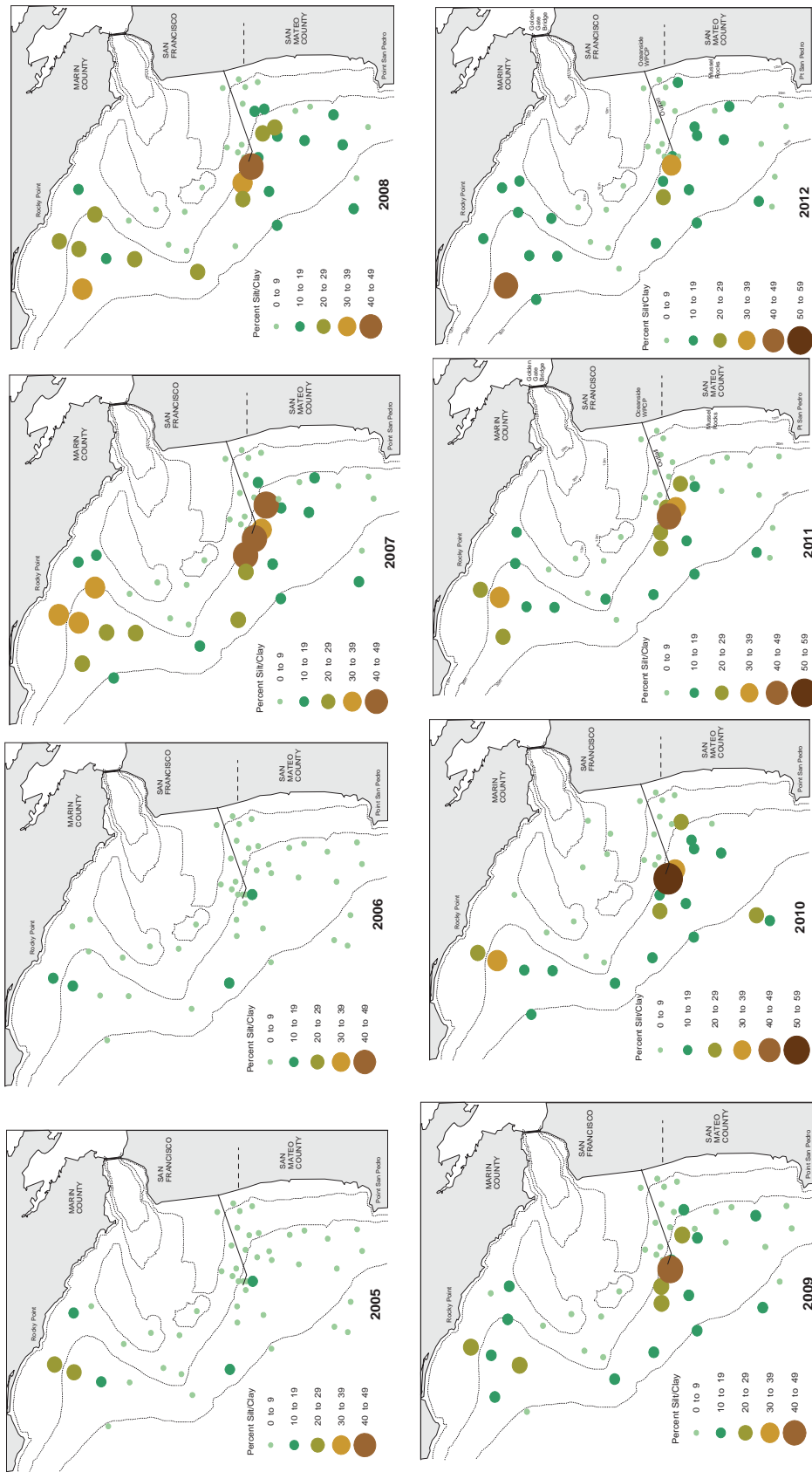


Figure 4-4b  
Percent silt and clay from 2005 to 2012

#### 4.2.1.3. Cluster Analysis

Sixteen-year cumulative cluster analyses (Figure 4-5) identified five cluster groups, using sediment physical measurements (sorting, percent silt/clay, and mean phi size) and total organic carbon (TOC). TOC is often used as a substitute for silt/clay in sediment chemistry studies. The combined analysis shows how stations changed or remained the same over time. Stations within each cluster group are more similar to each other than they are to stations in other cluster groups. Only two stations (stations 31 and 43) in the study area displayed persistent sediment characteristics over the sixteen-year study period, remaining in a single cluster group. All other stations showed some degree of variability and fell into more than one cluster group during this study period. The outfall, reference, and ancillary stations have generally clustered together.

The similar cluster patterns of the outfall stations and reference stations continue to reflect the similarity in grain size and organic carbon concentrations between these regions, validating the selection of reference stations used in the reference envelope analysis for chemical contamination (see 4.2.3). Golden Gate stations generally fell into cluster group 3, with occasional stations in cluster groups 1, 2, and 4. The ancillary station group clustered very similarly to outfall and reference stations. The majority of these stations are either near the outfall or along the SWOO pipeline.

#### 4.2.2. SEDIMENT CHEMISTRY

##### 4.2.2.1. 2012 TOC, TVS, TKN

Results for 2012 sediment chemistry data are presented in Appendix D-3. TOC and Total Volatile Solids (TVS), are measures of the amount of organic material in sediments and can reflect additions or loadings from wastewater discharges. Total Kjeldahl Nitrogen (TKN) is a measure of ammonia and organic compounds, and therefore an indicator of nitrogen

availability. TOC values in 2012 at outfall stations ranged from 1.31% to 2.95%; reference site values range from 1.05% to 8.72%. TVS values in 2012 at outfall stations ranged from 1.70 to 2.57; reference sites values range from 1.18 to 9.40. TKN values in 2012 at outfall stations ranged from 241 mg/Kg to 348 mg/Kg; reference site values range from 267 mg/Kg to 867 mg/Kg.

##### 4.2.2.2. 1997-2012 TOC, TVS, TKN

Average percent TVS values from 1997 to 2012 at outfall sites ranged from 1.5 to 2.4; stations in the northern reference area ranged from 2.0 to 3.5, southern reference stations range from 1.6 to 2.0 (Figure 4-6). Average TOC values at outfall sites ranged from 1.4% to 3.3% (Figure 4-7) Average TOC measurements between 1997 and 2012 were highest in the northern reference area (from 2.22% to 6.63%); station 32 had consistently high values during the sixteen year study. Average TKN values at outfall stations from 1997 through 2012 ranged from 221 mg/Kg to 342 mg/Kg (Figure 4- 8). Stations in the northern reference area had the highest TKN values. TKN averages at outfall stations were below that of northern reference stations, indicating negligible influence of the SWOO discharge on nitrogen loading in the outfall area. The Golden Gate stations, with the highest percentages of coarse and medium sand, generally had the lowest measurements of all three sediment chemistry parameters, with outliers at station 45 in 1997 which had particularly high values of TOC and TVS (WQB 1998).

##### 4.2.2.3. Organic Pollutants

Organic waste discharged through sewage outfalls may be composed entirely of non-toxic domestic sewage or some combination of domestic and industrial/chemical waste which may include a toxic component (Swartz et al. 1984). Sewage discharged through the SWOO is predominantly domestic with minor commercial

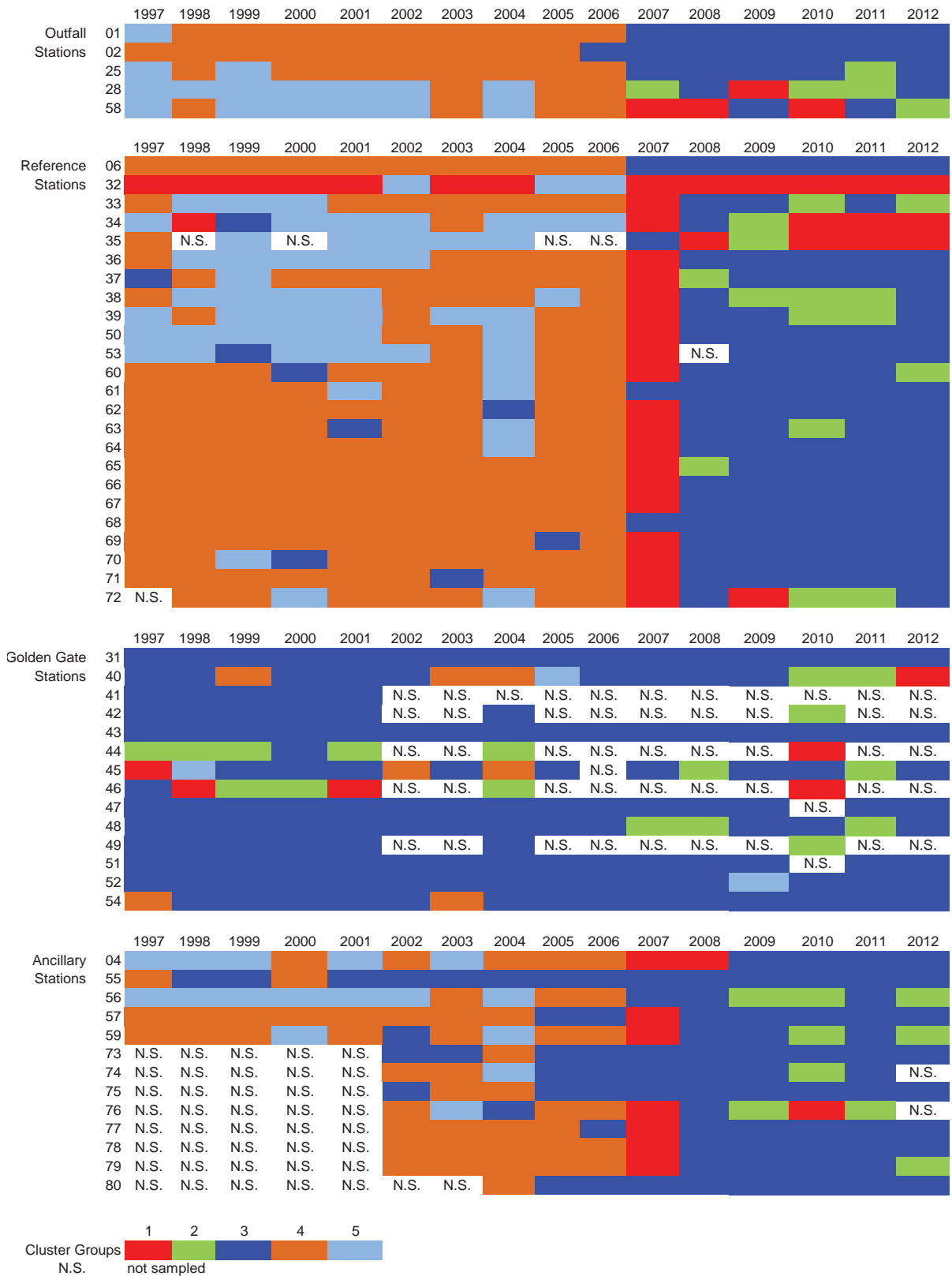


Figure 4-5  
Station-Cluster group matrix showing relationship of stations through time (1997 to 2012)

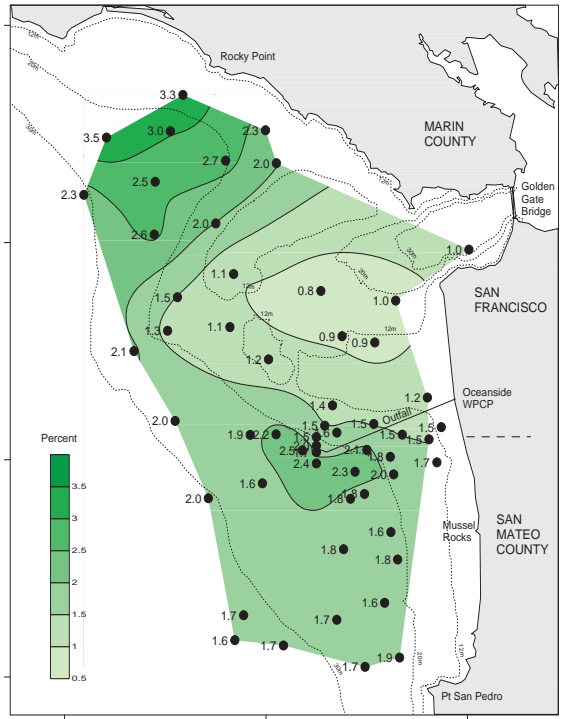


Figure 4-6  
Average percent TVS from 1997 to 2012

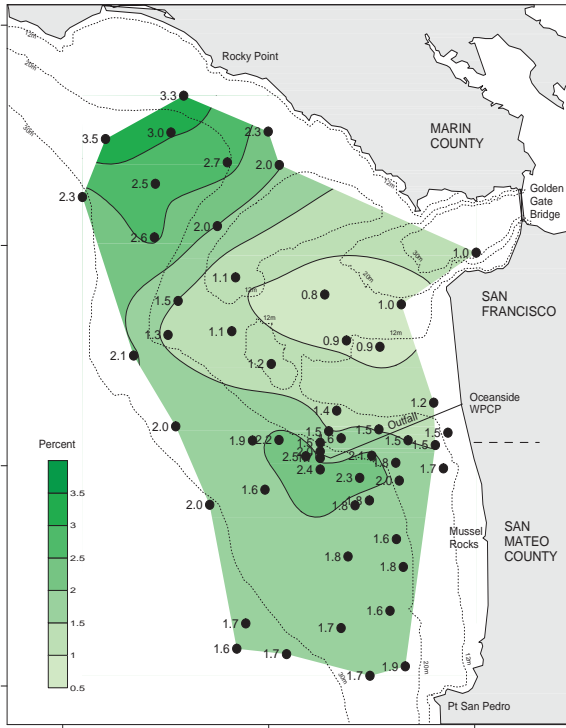


Figure 4-7  
Average TOC (mg/Kg) from 1997 to 2012

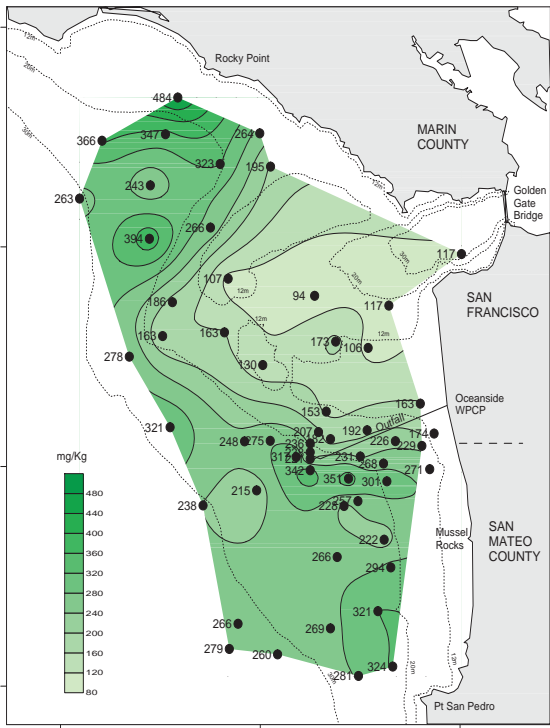


Figure 4-8  
Average TKN (mg/Kg) from 1997 to 2012

additions from hospitals, restaurants, and dry cleaners. Measureable amounts of organic pollutants were detected in sediments at forty-six (46) of the forty-nine (49) stations sampled in 2012 (Appendix D-4).

Final effluent samples analyses from the Oceanside WPCP over the last eleven years (2001 to 2012) have not exceeded NPDES permit limits for metals or organic compounds. Total polycyclic aromatic hydrocarbons (PAHs) from Oceanside WPCP final effluent also have not exceeded NPDES permit limits (Appendix D-5). Figures 4-9a & b shows the distribution of average concentrations of all PAH pollutants measured at each SWOO station over the sixteen-year period. PAHs were found to be a major organic component of the sediments in the vicinity of the outfall before construction began (pre-discharge) (de Lappe et. al. 1980) and the ratios of the PAH compounds indicated that their source was combustion processes (BWPC 1984). PAH concentrations within the study region appear to be transitory and have not followed percent silt/clay patterns (Figure 4-9).

Polychlorinated biphenyl (PCB) congeners and organochlorine pesticides (such as DDT) are infrequently detected and are only generally detected at low concentrations within the study area (Appendix D-4; WQB 1998, 1999, 2000, 2001a, 2003a, 2003b, 2004: NRD 2006a, b,; NRLMD 2007, 2008, 2009, 2010, 2011, 2012).

#### 4.2.2.4. Inorganic Pollutants

Results of trace metals analysis at all stations in 2012 are shown in Appendix D-6. Analytical techniques and detection limits are listed in Appendix D-7.

Figure 4-10 shows the average concentration of metals of concern (Copper, Mercury, Nickel, and Selenium) from outfall and reference stations from 1997 to 2012. Appendix D-8 shows summary measurements of Oceanside Water Pollution Control Plant final effluent for thirteen metals. Samples were taken quarterly, then yearly, depending on permit year. All measured concentrations were within NPDES

permit limits.

Both a standard analysis of variance (ANOVA), testing metal by metal, and a multivariate analysis of variance (MANOVA), testing a multivariate summary of metals variation, were performed on sediment metals concentrations to detect differences between reference and outfall stations during the study period (1997 to 2012) (Appendix D-9). Overall, the concentrations of metals at reference stations are the same as the concentrations at outfall stations ( $p > 0.05$ ). There is no discernible outfall effect in the sediment metals data.

#### 4.2.2.5. Ordination Analysis for Sediment Chemistry

Principal component analysis (PCA) (see Methods, 2.4.2. Multivariate Analyses) of grain size parameters and sediment chemistry identified six independent gradients of change in the study area (Appendix D-10). The first four axes from the PCA analysis encompass



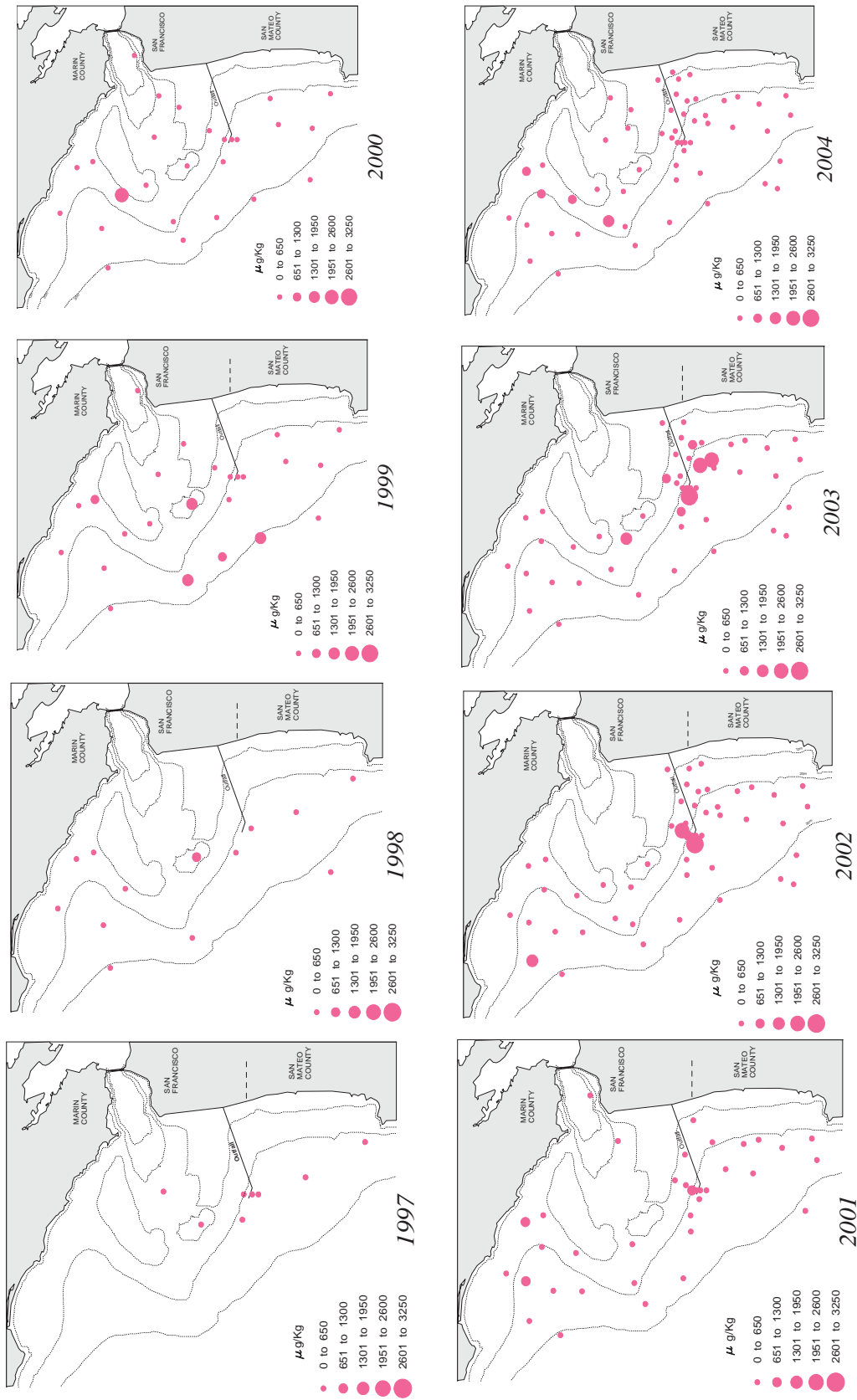


Figure 4-9a  
 Total PAHs ( $\mu\text{g/Kg}$ ) measured from 1997 to 2004. The  
 number of stations sampled increased each year.

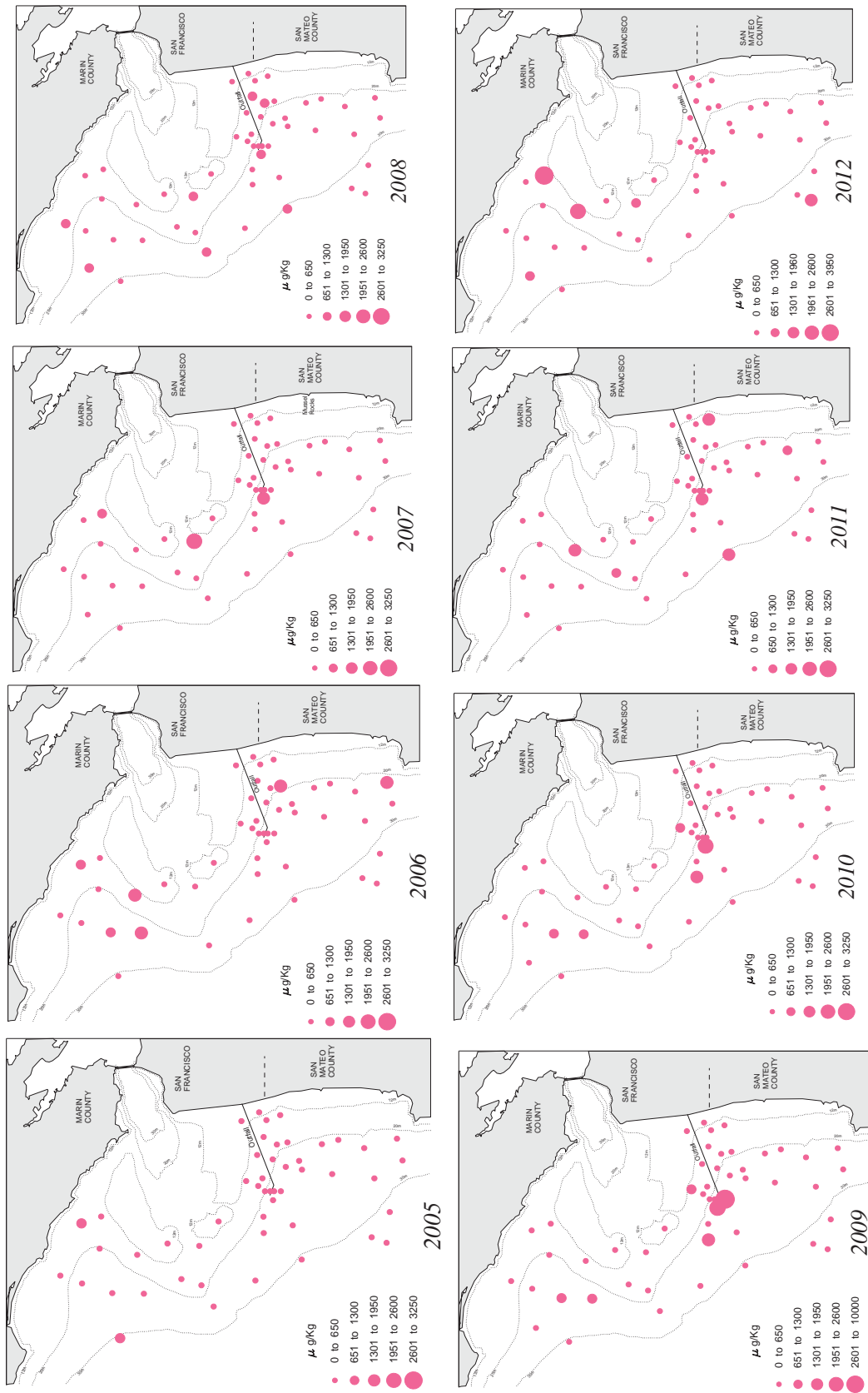


Figure 4-9b  
Total PAHs ( $\mu\text{g}/\text{Kg}$ ) measured from 2005 to 2012.

more than 87% of the variation in the data. Axis 1, accounting for 31% of the variation, combines low values of parameters associated with organic loading – silt/clay, TOC and TVS. The next three axes were all dominated by a single factor: Axis 2 (21% of the variation) by higher concentrations of metals; Axis 3 (18% of the variation) by higher concentrations of sediment nitrogen (TKN); and axis 4 (17% of the variation) by higher concentrations of organic pollutants. The remaining two axes combined explain less than 13% of the variation in the data.

#### 4.2.3. REFERENCE ENVELOPE

Reference envelope tolerance interval bounds (see Methods 2.4.2.2 Reference Envelope Analyses) were calculated using TOC, TVS, TKN, silt/clay, metals (as the sum of their ERM quotients), and total organic pollutant from reference stations (06, 32-39, 50, 53, 60-72).

Outfall station measurements are compared to the reference station tolerance interval bounds in order to assess the potential degree of impact from the SWOO discharge (Figures 4-11a, b, & c). Analysis of measurements between 1997 to 2012 from 80 outfall samples and 384 reference samples showed that the majority of the outfall region measurements were within reference tolerance bounds, indicating that overall the SWOO discharge does not have an adverse impact on sediments in the SWOO study area. Outfall station 58 and reference station 32 have exceeded tolerance interval bounds most frequently over the last twelve years; station 58 for percent TVS in 2008, for percent silt/clay in 2004, 2007, 2009, 2010, 2011, and 2012, and for sum of organic pollutants (DDT, PAHs, and PCB's) in the following years: 2002, 2003, 2007, 2008, 2009, 2010, and 2011. The exceedance in organic pollutants at station 58 may be due to the high silt and clay content. Reference station 32, also with high values of silt and clay, exceeded tolerance bounds for almost all parameters in multiple years except for sum of metals ERMq. Outfall station 02 exceeded the tolerance interval bound for sum of organic pollutants in 2001 and 2002 and for TKN in 2003. Outfall station 28

exceeded tolerance bounds for percent silt/clay in 2007, 2009, 2011 and 2012. Other northern reference stations, as well as a few southern reference stations, have exceeded tolerance bounds for various parameters. These stations typically have the highest percentages of silt and clay each year and often exceed or approach the upper tolerance bound for organic content.

#### 4.2.4. BACIP ANALYSIS

The BACIP (Before-After-Control-Impact-Paired) statistical model (see Methods, 2.4.2.3 BACIP Analysis) was used to evaluate whether or not conditions have changed at the outfall since the onset of effluent discharge. Parameters used to test the relationship between an outfall and a reference station were percent silt/clay, TOC, percent TVS, and TKN. The null hypotheses for this statistical test is that the mean differences between the outfall and reference station measurements are the same in pre-discharge and discharge periods. Stations 01 (outfall) and 06 (reference) were used in this analysis because they have the longest history of continuous sampling. While there was no significant difference in relative values between pre- and post-discharge periods for TOC, percent TVS, and TKN ( $t(23) = 1.72, 1.56, 1.81$  respectively,  $p > 0.05$ ), there was a significant difference in percent silt and clay since pre-discharge at outfall station 01 ( $t(23) = 2.30, p < 0.05$ ); demonstrating a reduction of percent silt and clay at the outfall (Figure 4-12). This data suggests that since 1998, percent silt/clay at outfall station 01 has become more like reference station 06 post-discharge. However, there is evidence of increasing sediment fines throughout the study area (4.2.1 Grain Size section).

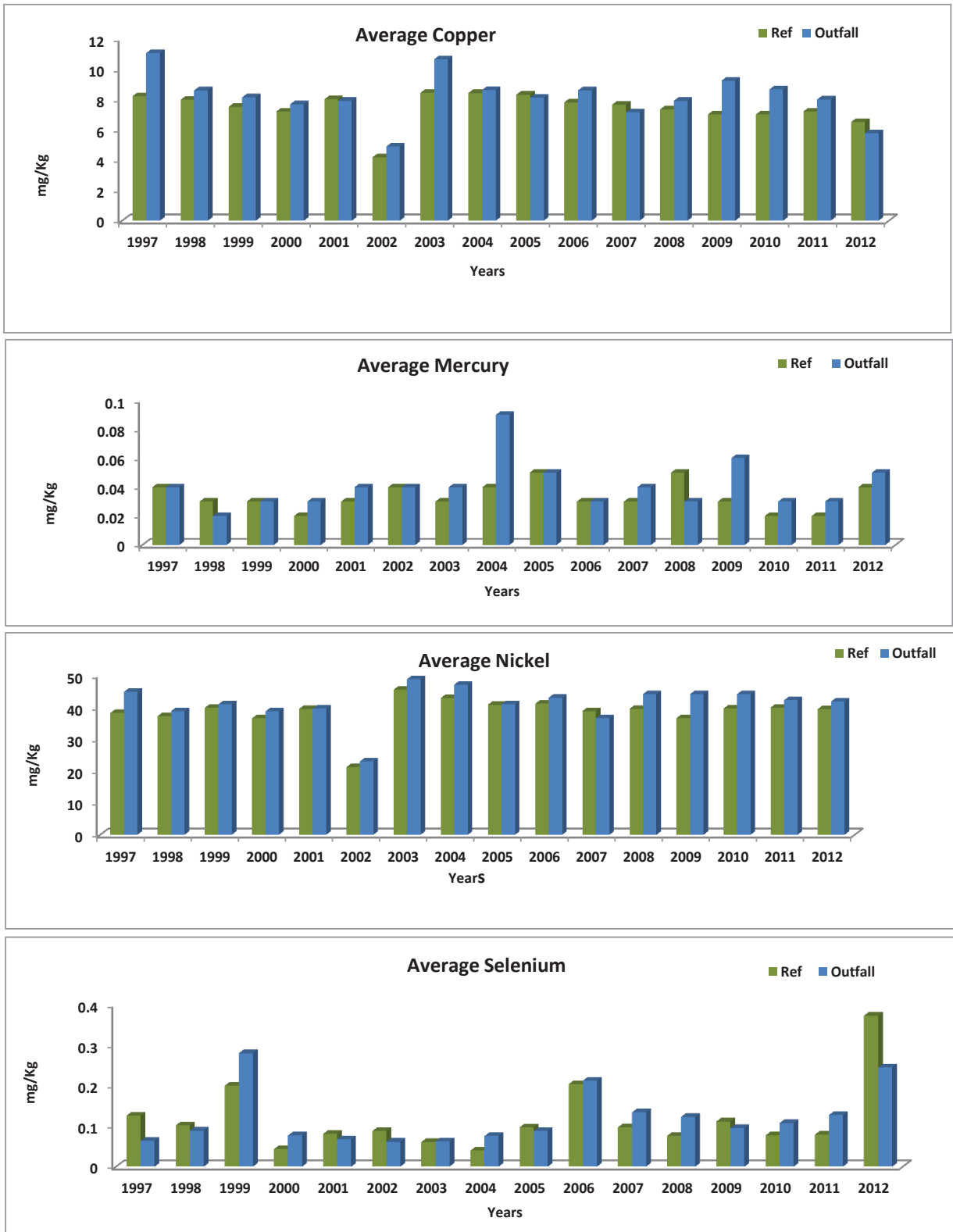


Figure 4-10

Average concentrations (mg/Kg) of copper, mercury, nickel, and selenium at SWOO reference and outfall stations from 1997 to 2012. These are metals of concern in San Francisco Bay (SFEI 2013).

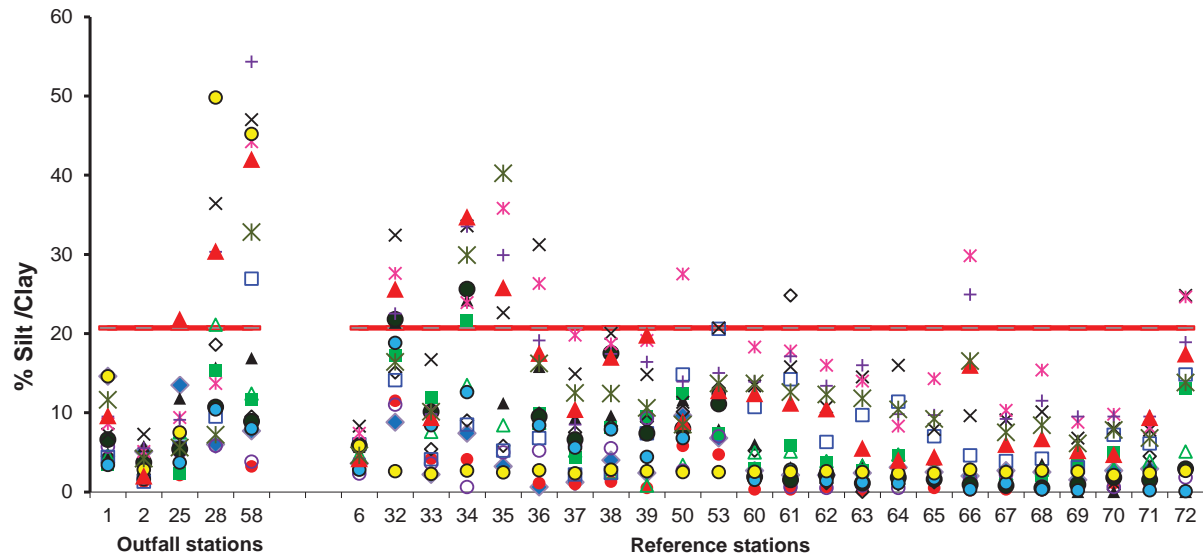
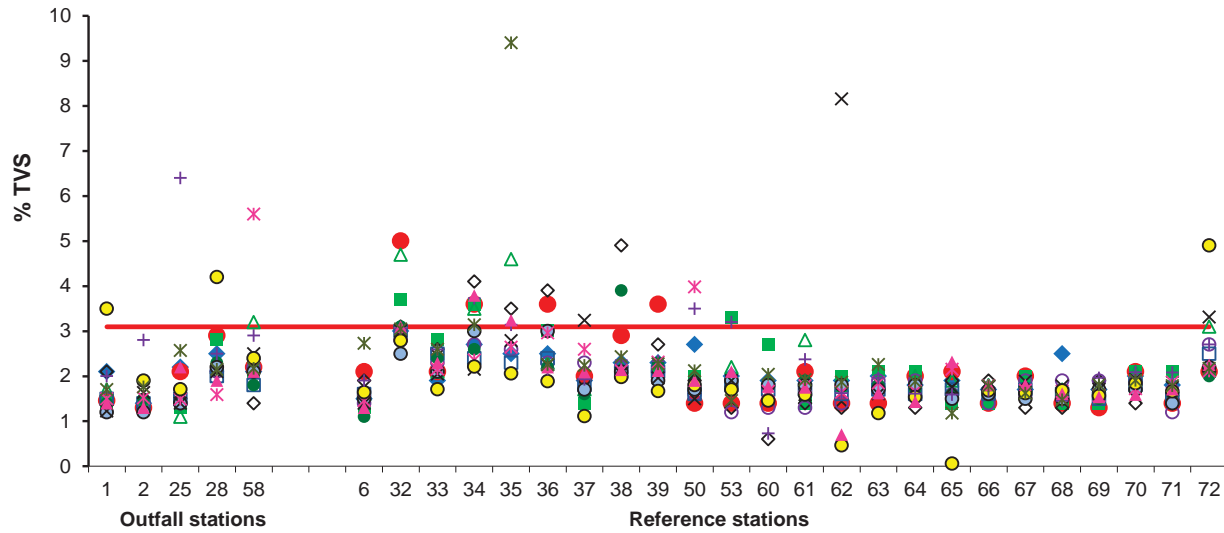


Figure 4-11a

Sediment indicators used in the reference envelope analysis and plotted with an upper tolerance interval bound  $P = .90$  and  $\alpha = .05$ . Only the upper interval bound is plotted since concern for these indicators would be an increase in the presence of a wastewater discharge. Tolerance interval bounds calculated from 384 reference samples.

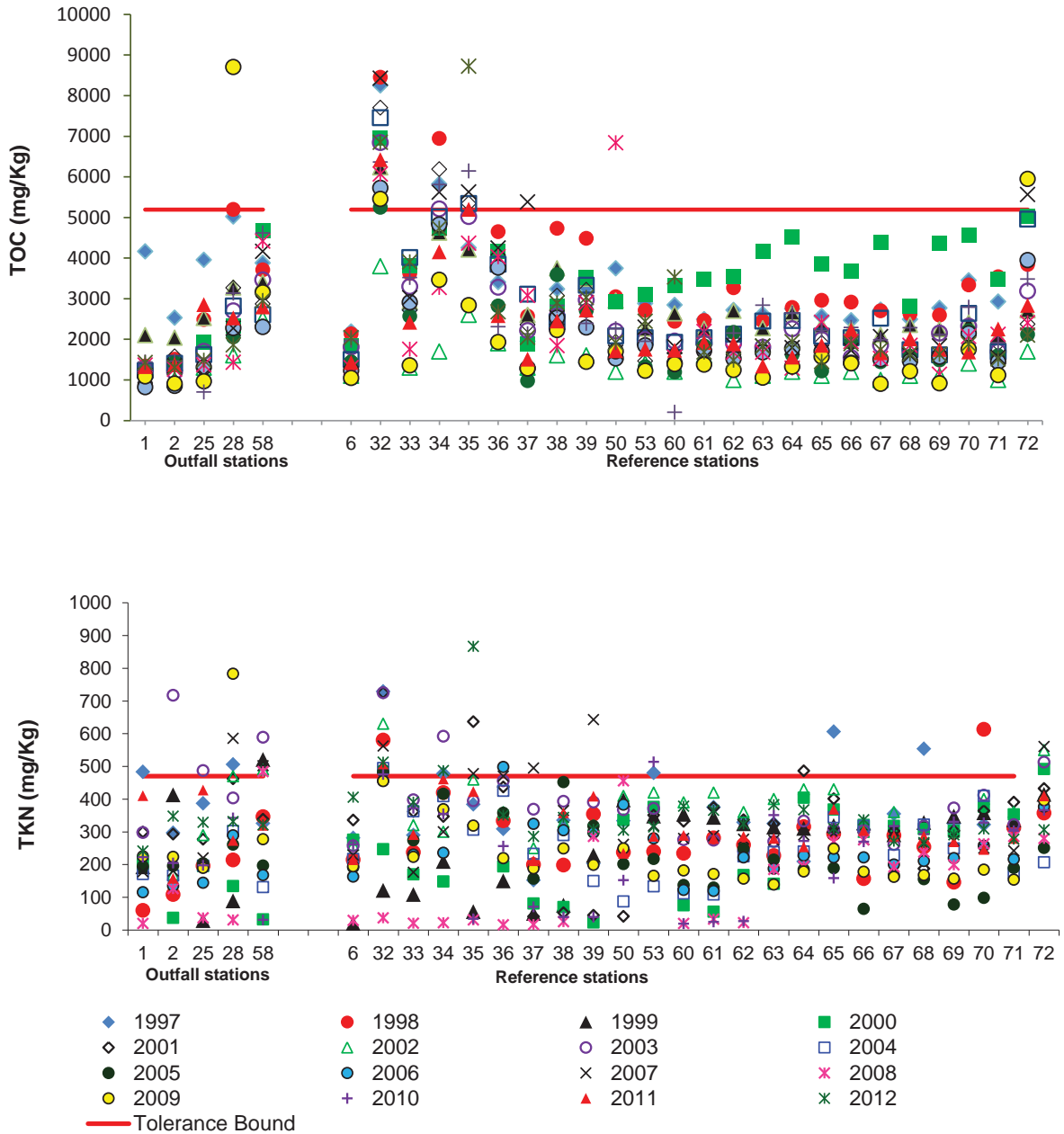


Figure 4-11b

Sediment indicators used in the reference envelope analysis and plotted with upper tolerance interval bound  $P = .90$  and  $\alpha = .05$ . Only the upper interval bound is plotted since concern for these indicators would be an increase in the presence of a wastewater discharge. Tolerance interval bounds calculated from 384 reference samples.

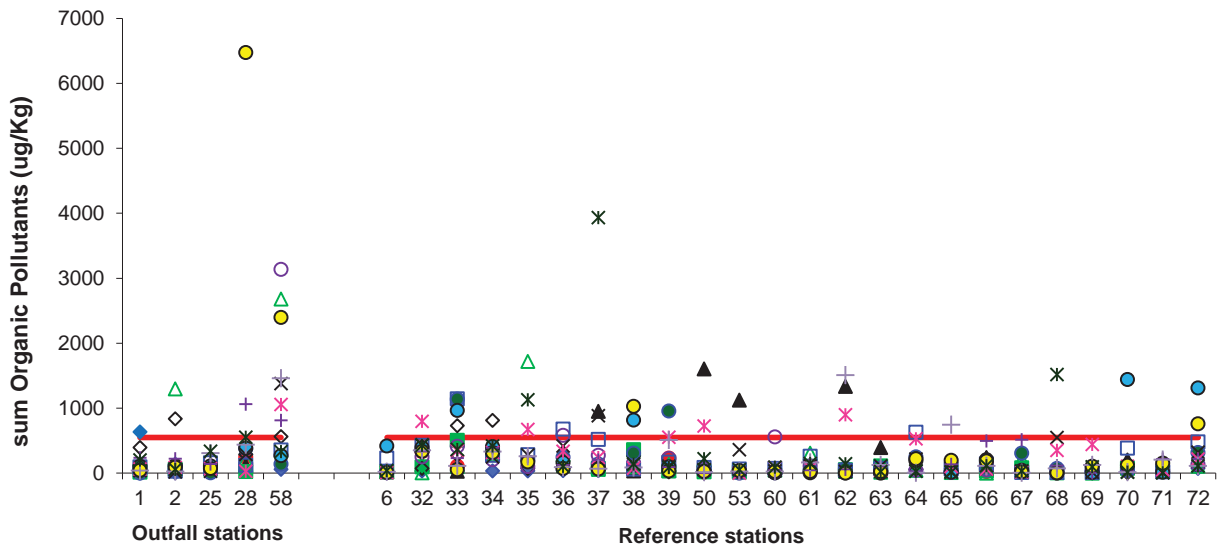
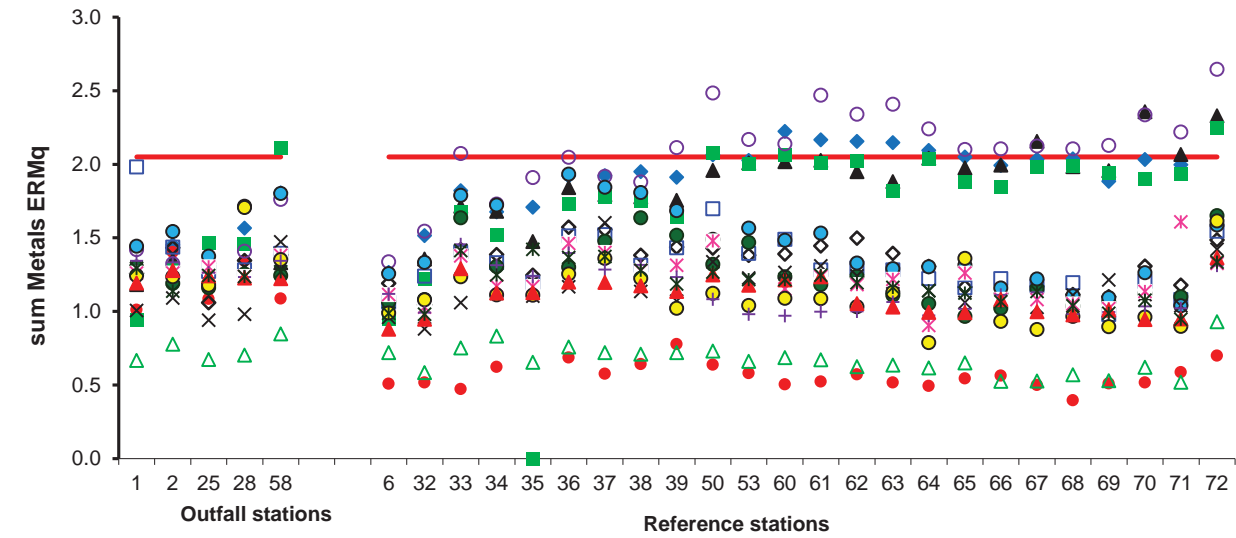


Figure 4-11c

Sediment indicators used in the reference envelope analysis and plotted with an upper tolerance interval bound  $P = .90$  and  $\alpha = .05$ . Only the upper interval bound is plotted since concern for these indicators would be an increase in the presence of a wastewater discharge. Tolerance interval bounds calculated from 384 reference samples.

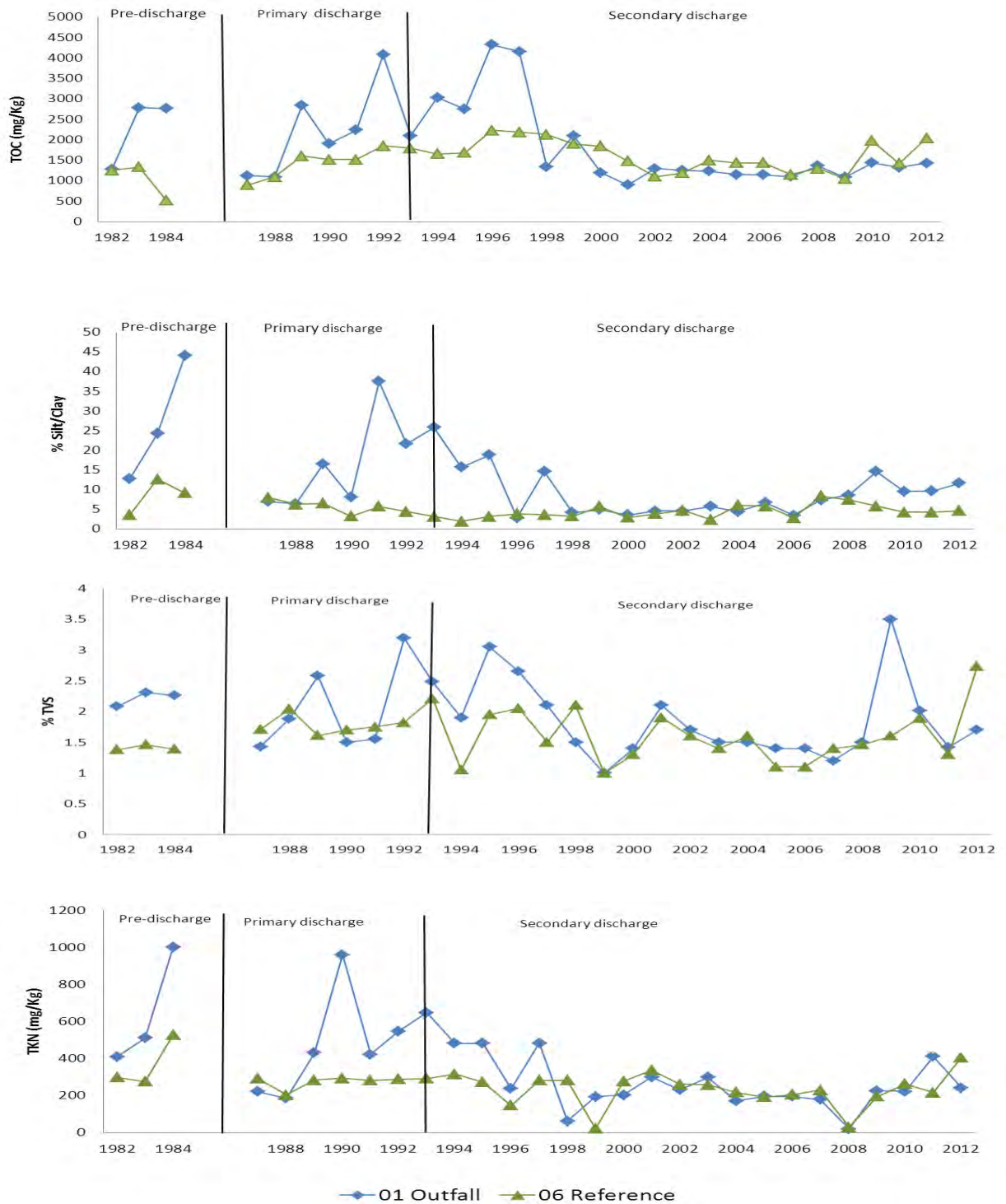


Figure 4-12  
Plots of sediment fines and organic content, 1982 to 2012



### 4.3. SUMMARY AND CONCLUSIONS

The study area is a high-energy environment comprising primarily sandy beaches and bluffs to the south of the Golden Gate, and rocky cliffs and pocket beaches to the north. The geology is controlled by active tectonics within the San Andreas and San Gregorio Fault Zones traversing directly through the region. This area is susceptible to high energy waves, being exposed to swell from almost the entire Pacific Ocean (Barnard et al., 2013).

The sedimentary environment of the SWOO study area is dominated by input from the Sacramento and San Joaquin River system through the Golden Gate, and by reworking from tidal currents and wave action. Sediment-laden currents funnel through the Golden Gate on ebb tides and fan out depositing sediments along the transport path (Carlson and McCulloch 1974, Conomos 1979). Tidal and longshore currents rework these sediments to form the barrier sandbars that surround the mouth of San Francisco Bay. The stations inside these sandbars are characterized by medium to coarse sands. Those stations that are offshore from the sandbars, and which are more or less in line with the Golden Gate Bridge and the shipping channel, had a greater percentage of medium sands than other stations at similar depths. The sandbars themselves are bathymetric highs made of fine to medium sands. Recently, an area of massive sand waves was mapped between the sand bars and the Golden Gate (Barnard, et.al, 2006). Beyond the sand bars are areas of very fine sands with the highest average percentages of silt and clay (up to 47%) just outside the sandbars, decreasing to less than 10% well beyond the sandbars.

Mean sediment particle sizes at the outfall area have not significantly changed since pre-discharge and pre-construction periods, suggesting that the SWOO outfall and effluent discharge has not affected sediment grain size distribution. Well-sorted sediments were found at stations on or near the sandbars (stations 31, 40, 41, 42, 44, 46, 43, 47, 51, 52, 54, and 57),

likely as the result of reworking by waves and tidal currents. The northern reference region (stations 32, 33, 34, 35, 36, 38, and 39) had some of the most poorly sorted sediments and relatively high percentages of silt/clay in the study area.

Smaller sediment grains provide greater relative surface area for adsorption of contaminants and organic matter, therefore areas outside the barrier sandbars with finer sediments may be expected to have higher contaminant concentrations. Such a pattern of grain size distribution, measures of TOC, TVS, TKN, and metals concentrations, has been consistent throughout this study period. Consequently, the location of the SWOO (just south of a sandbar) places it in an environment where elevated measures of sediment fine grains, organic matter, and contaminants might be expected even in the absence of a wastewater discharge. It is important therefore to evaluate potential discharge impacts by comparing similar environments. The sandbar stations, and the coarse sand stations inside the sandbars, are in such a different sedimentary environment/hydrodynamic regime from stations near the outfall that the sediment characteristics from these coarse grain stations are not relevant to interpreting an outfall effect. Reference envelope analysis, using tolerance bounds defined by multiple reference station locations, of sediment measurements from outfall stations indicates that some stations exceeded tolerance bounds in four different categories with approximately the same or lower frequency as stations in the reference region. An exception is outfall stations 58 and 28 with elevated organic pollutants in multiple years.

The number of stations analyzed for organic priority pollutants was increased every year since 1997 until all stations were included by 2001. Concentrations of 18 PAHs were measured throughout the study area and have not paralleled percent silt and clay patterns. High concentrations near the outfall in some sampling years appear to be transitory. Variations in

sediment organic contaminant concentrations between survey years may be a result of sediment movement over time via currents and winter storms. All stations have been sampled for trace metal analyses from 1997 to 2012. Overall, the concentrations of metals at reference stations are the same as the concentrations at outfall stations. There is no discernible outfall effect in the sediment metals data.

After sixteen years of monitoring the SWOO study area, sediment data have revealed no trends in sediment characteristics that would indicate that the discharge from the Oceanside Water Pollution Control Plant has adversely affected the surrounding environment. Increasing sediment fines in the northern reference area and around the SWOO outfall have been attributed to a consolidation of the ebb-tidal delta of the San Francisco Estuary, related to a recent step-decrease in suspended sediment concentrations observed inside San Francisco Bay (Barnard et. al., 2012). Continued analyses of wastewater effluent samples have not exceeded established water quality standards for the parameters measured. Physical measurements of sediment grain size and chemical organic and inorganic data further demonstrate that the discharge has not produced any discernable effects on the physical characteristics of sediment or resulted in contaminant accumulation in the vicinity of the outfall.



## **SECTION 5**

### **BENTHIC INFUANA**

# BENTHIC INFAUNA

## 5.1. INTRODUCTION

Benthic infauna communities were monitored to evaluate effects associated with the discharge of treated wastewater effluent from the Southwest Ocean Outfall (SWOO) into nearshore waters of the Gulf of the Farallones (Figure 5-1).

Wastewater discharges characteristically change the properties of bottom sediments next to outfalls, which in turn affect the natural biological communities (e.g. LACSD 1981). Relationships between benthic infauna communities and sediment (Section 4.2.1) have helped in determination of the role sediments play in governing community characteristics.

Some of the analyses used only stations that were sampled every year to make the comparisons valid. These include the annual comparisons of the 20 most abundant taxa and the annual total abundance surveys. Analyses employing the entire database include the geographic distributions of community parameters, ordination, cluster and reference envelope analyses. See cluster diagram (Cluster Analysis Section 5.2.3.2) for the stations sampled each year throughout the survey period.

### 5.1.1. STATION GROUPS

Some figures and discussion refer to reference, outfall, Golden Gate, ancillary and reef effect groups of stations (Figure 5-1). These groupings allow comparison of similar stations and are based on geographic location, common sedimentary characteristics, and infauna community measures (Methods, Section 2.2.4). Stations in the reference (green) and outfall (blue) groups are generally characterized by well-sorted very fine sand with a variable percentage of silt and clay, and similar infauna communities. The outfall group includes those stations in closest proximity to the SWOO

discharge. Stations within the Golden Gate group (orange) are generally shallower and have coarser sediments that support different infauna communities than the reference or outfall groups. Stations that compose the ancillary station group (black) are outside of the direct influence of the discharge plume but too close to be considered a suitable reference station. Reef effect stations (purple) were added to assess the potential effect of the outfall structure on the benthic communities. Station 80, which is included in the reef effect stations in this study, was added in 2004 to assess the effect of the North San Mateo County Sanitation District outfall on nearshore benthic abundance values.

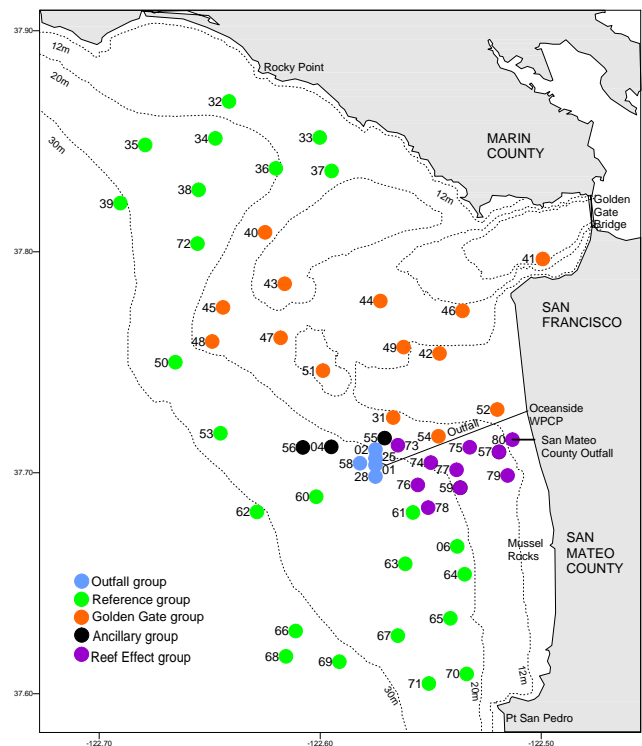


Figure 5-1

Stations sampled for infauna and station groups used in comparison analyses. Outfall stations (blue) were used to define outfall conditions that were compared to reference conditions (green). Golden Gate stations (orange) characterize conditions near the Bay outflow and reef effect stations (purple) were examined to determine if there is an effect from the outfall structure. Ancillary station (black) are outside the influence of the effluent discharge.

## 5.2. RESULTS AND DISCUSSION

### 5.2.1. GENERAL CHARACTERIZATION

#### 5.2.1.1. Survey year 2010

The processing of benthic infauna samples from stations 78 and 79 sampled in 2010 were not completed in time for inclusion in the annual data report. The infauna data for these stations are reported in Appendix E-4 and are included in the analyses of this report.

#### 5.2.1.2 Survey year 2012

Appendix E-2 lists the raw data for 2012 and Table 5-1 displays a summary of the 20 most abundant species in the 2012 survey. A total of 149,951 specimens were collected and the polychaete worm, *Spiophanes norrisi* (formally reported as *Spiophanes bombyx*), was the most abundant taxa as it was in the previous 3 years of the survey. The mollusk, *Callianax pycna* (formally reported as *Olivella pycna*) was unique in being present in all of the samples collected in 2012.

Table 5-1

The twenty most abundant species in 2012. Abundance was dominated by polychaetes, especially *Spiophanes norrisi*.

Species	Taxonomic Group*	Total Abundance	Percent Total Abundance	Occurrence at 47 Stations	Percent Occurrence
<i>Spiophanes norrisi</i>	P	89,095	59.42	45	95.74
<i>Photis</i> sp.	C	10,189	6.79	44	93.62
<i>Callianax pycna</i>	M	6,267	4.18	47	100.00
<i>Photis macinerneyi</i>	C	4,776	3.19	39	82.98
<i>Protomedeia penates</i>	C	4,476	2.98	41	87.23
<i>Scoletoma luti</i>	P	3,618	2.41	45	95.74
<i>Mediomastus</i> spp.	P	2,393	1.60	41	87.23
<i>Diastylopsis dawsoni</i>	C	1,630	1.09	46	97.87
<i>Glycinde picta</i>	P	1,629	1.09	39	82.98
<i>Owenia collaris</i>	P	1,604	1.07	28	59.57
<i>Tellina modesta</i>	M	1,577	1.05	46	97.87
<i>Glycinde</i> spp.	P	1,528	1.02	42	89.36
<i>Rhepoxynius fatigans</i>	C	1,365	0.91	33	70.21
<i>Ischyrocerus pelagops</i>	C	1,351	0.90	30	63.83
<i>Pectinaria californiensis</i>	P	992	0.66	38	80.85
<i>Onuphis</i> sp. A	P	775	0.52	41	87.23
<i>Pleurogonium</i> sp. SF1	C	710	0.47	39	82.98
<i>Apopronospio pygmaea</i>	P	619	0.41	37	78.72
<i>Pacifoculodes barnardi</i>	C	612	0.41	44	93.62
<i>Magelona sacculata</i>	P	553	0.37	36	76.60

\*Taxonomic Groups: P=Polychaeta, M=Mollusca, C=Crustacea

#### 5.2.1.3 Survey Years 1997-2012

For the study period a total of 879,666 individual organisms belonging to 628 taxa and representing 15 phyla were identified in 786 samples collected during the sixteen-year period from 1997 to 2012 (Table 5-2, Appendices E-1 and E-2, WQB 1998-2004, NRD 2006a, 2006b, NRLMD 2007-2011,).

The annelid class Polychaeta represented the greatest species richness and total abundance and contributed 41.9% of the taxa (263) and 64.1% of the total abundance (564,027 specimens) for the period (Table 5-2). Dominance of polychaete species in benthic assemblages has been documented in several studies (Reish 1983, Diener and Fuller 1995, Dorsey et al. 1995). In general, polychaete species compose over 40% of benthic infauna communities regardless of depth (Knox 1977). Mollusca were represented by 19.4% of the taxa (122) and contributed 20.1% of the total abundance (177,140 specimens). Crustacea, with 29.5% of the taxa (185), contributed 10.3% of the organisms for the period (90,468 specimens). Echinoderms

Table 5-2

Total abundance and relative contribution of major taxa from 1997 to 2012. Data includes all stations collected. Polychaetes dominate overall in species richness and abundance.

Taxon	Number of Taxa	Abundance
Polychaeta		
number	263	564,027
percentage	41.9%	64.1%
Mollusca		
number	122	177,140
percentage	19.4%	20.1%
Crustacea		
number	185	90,468
percentage	29.5%	10.3%
Echinodermata		
number	11	8,819
percentage	1.8%	1.0%
All Other Taxa		
number	47	39,212
percentage	7.5%	4.5%
TOTALS		
number	628	879,666
percentage	100.0%	100.0%

were the least abundant major group, with 1.8% of the taxa (11) and 1.0% of the organisms (8,819 specimens). All other groups combined contributed 7.5% of the taxa (47) and 4.5% of the total abundance (39,212 specimens).

The twenty most abundant organisms in the sixteen-year period (Table 5-3) accounted for 74.6% of the total individuals at stations sampled every year. The polychaete *Spiophanes norrisi* was the most abundant species, composing 30.6% of the total number of individuals. During 2010, 2011, and 2012, this species accounted for over half of the total individuals collected. During 2007 and 2008, the most dominant taxa was a bivalve clam, *Mactromeris catilliformis*, which accounted for nearly half (47.5%) of the specimens collected in 2008. The polychaete worm taxon *Mediomastus* spp. was the third most abundant with a total percent abundance of 4.2%, and it dominated the benthic community in 2005. Another

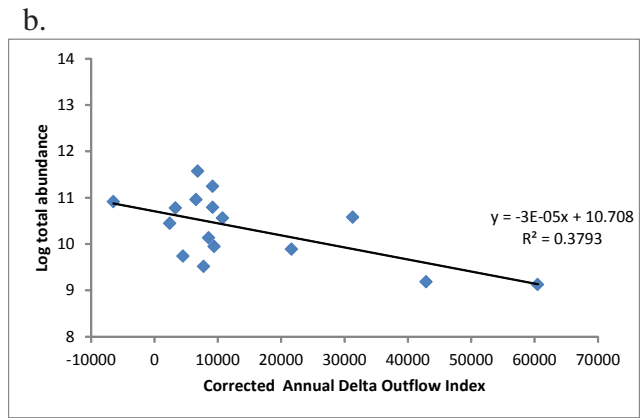
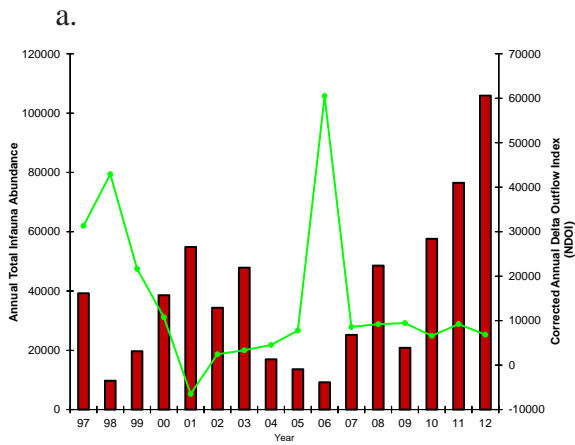
spionid polychaete, *Spiophanes berkeleyorum*, was fourth in abundance with over 11% of the specimens collected in 2000 and 2001. The bivalve, *Tellina modesta*, mostly due to high recruitment in 1997, was the fifth most abundant species. The taxa listed in Table 5-3 contributed a minimum of 47% of the annual abundance for years surveyed, as well as 74.6% of the total abundance for the survey period. These organisms include 12 polychaetes, five mollusks, and three crustacea. *Spiophanes norrisi* is unusual in sustaining several years of relatively high abundance while other taxa have only one or two years of high abundance (e.g. *Tellina modesta* in 1997, *Apoprinospio pygmaea* in 1998, *Owenia collaris* in 2003, *Scoletoma luti* in 1998, *S. berkeleyorum* in 2000 and 2001, *Mactromeris catilliformis* in 2007 and 2008). The composition of this table has changed very little since the first summary report (WQB 2003a).

Table 5-3

Twenty most abundant taxa from 1997 to 2012 with cumulative, annual percent and percent total abundance for each taxon. Comparison includes only stations sampled every year. In 2010, 2011 and 2012 the polychaete worm, *Spiophanes norrisi*, was dominant accounting for over half the specimens collected in those years.

TAXON	Taxa*	Cumulative Abundance 1997-2012	Annual Percent Abundance																	% Total Abundance 1997-2012
			1997	1998	1999	2000	2001	2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012		
<i>Spiophanes norrisi</i>	P	189,317	1.2	1.0	2.4	6.0	9.6	27.2	25.6	10.2	5.2	2.1	6.8	5.9	25.0	64.8	63.4	57.5	30.6	
<i>Mactromeris catilliformis</i>	M	43,436	5.6	3.2	3.9	3.8	7.4	5.8	3.7	1.5	2.6	6.1	18.1	47.5	0.4	0.2	2.2	0.2	7.0	
<i>Mediomastus</i> spp.	P	25,830	7.6	6.6	7.3	3.1	2.7	7.8	3.7	8.4	12.1	7.0	5.7	3.8	5.2	2.7	2.6	1.8	4.2	
<i>Spiophanes berkeleyorum</i>	P	21,742	1.6	1.2	6.6	11.1	11.6	1.2	2.3	1.0	1.2	4.8	7.4	1.1	0.9	6.4	0.4	0.1	3.5	
<i>Tellina modesta</i>	M	20,585	17.9	2.4	2.6	4.5	3.6	3.9	2.1	3.7	2.1	1.9	0.6	2.6	5.7	0.8	1.7	1.2	3.3	
<i>Scoletoma luti</i>	P	20,321	8.9	12.7	3.1	0.9	2.0	1.8	1.9	6.1	4.9	7.2	2.9	2.3	6.4	2.4	3.0	2.7	3.3	
<i>Owenia collaris</i>	P	15,425	0.3	0.2	0.2	0.7	3.1	5.3	18.2	5.0	1.9	0.0	0.0	0.1	0.2	0.0	0.1	1.3	2.5	
<i>Protomedeia penates</i>	C	13,033	0.4	0.0	0.1	1.2	6.7	4.0	2.0	1.9	0.3	0.9	0.8	0.5	2.2	0.5	0.9	3.9	2.1	
<i>Pectinaria californiensis</i>	P	12,683	3.6	1.3	2.8	4.3	8.2	1.1	1.1	0.7	0.5	0.8	4.4	1.2	0.1	0.4	0.6	0.8	2.1	
<i>Callianax pycna</i>	M	12,679	0.1	0.8	0.1	1.5	2.1	1.1	2.9	0.3	0.4	0.3	0.0	0.5	4.2	1.9	4.7	2.9	2.0	
<i>Apoprinospio pygmaea</i>	P	12,110	1.5	15.2	3.7	2.3	1.6	1.2	1.2	1.1	4.9	7.3	7.2	1.9	1.0	1.6	0.9	0.5	2.0	
<i>Onuphis</i> sp. A	P	12,072	2.9	2.8	6.5	3.1	2.1	1.3	2.0	2.0	3.0	3.8	2.9	3.4	1.9	0.7	0.9	0.6	2.0	
<i>Photis</i> spp.	C	11,670	0.1	0.1	0.3	0.5	1.3	0.8	1.1	0.4	0.5	0.1	0.4	0.7	1.5	0.9	1.0	7.3	1.9	
<i>Leukoma staminea</i>	M	9,357	8.3	1.7	1.2	4.0	1.9	0.7	1.5	0.5	1.5	0.2	1.0	1.7	1.6	0.1	0.3	0.2	1.5	
<i>Kurtiella tumida</i>	M	8,580	3.2	3.1	0.7	1.2	2.8	4.5	1.9	6.4	1.8	0.7	0.3	0.3	1.1	0.2	0.3	0.3	1.4	
<i>Magelona sacculata</i>	P	8,041	0.6	2.8	1.6	2.6	0.7	0.5	1.1	0.9	3.8	4.1	3.7	0.5	1.3	2.2	1.2	0.4	1.3	
<i>Glycinde picta</i>	P	7,210	0.3	0.1	0.1	0.5	1.7	1.7	0.9	1.6	0.7	0.3	0.7	2.6	1.5	0.9	1.7	0.9	1.2	
<i>Rhepoxynius fatigans</i>	C	6,071	0.4	0.6	0.3	0.2	0.3	1.2	1.3	4.0	3.8	2.9	1.6	0.4	2.0	0.4	0.7	1.2	1.0	
Maldanidae	P	5,851	4.7	1.4	3.1	1.1	0.5	1.3	2.2	1.8	1.8	1.8	0.6	0.1	0.3	0.0	0.0	0.0	0.9	
<i>Magelona hartmanae</i>	P	5,625	0.5	1.9	1.2	3.9	0.7	0.4	0.6	0.5	1.2	1.2	1.1	0.3	2.1	1.1	0.6	0.3	0.9	
Totals			69.6	59.2	47.6	56.5	70.7	72.6	77.3	57.8	54.3	53.6	66.3	77.6	64.8	88.3	87.1	84.2	74.6	

\* = P = Polychaeta, M = Mollusca, C = Crustacea



Figures 5-2a & b

Annual total abundance plotted with the corrected annual mean outflow. The relationship is significantly negative ( $F\text{-ratio}=8.5567, p<0.05, df=1,16$ ). Data includes only stations sampled every year.

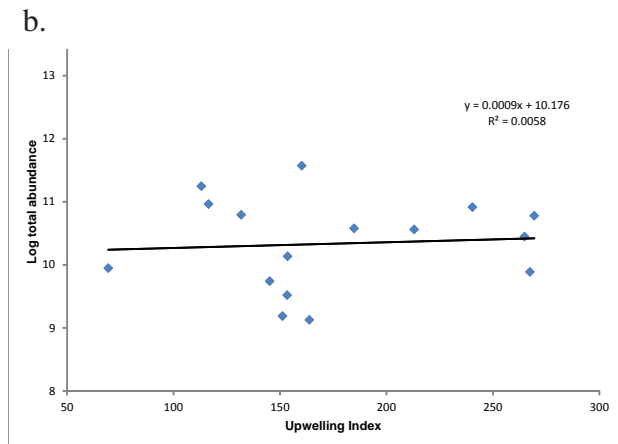
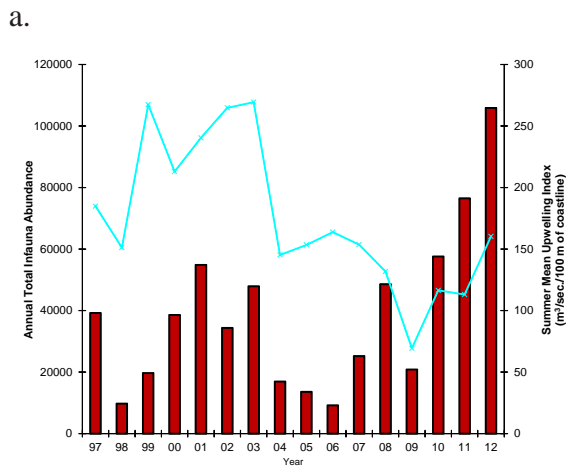


Figure 5-3a & b

Annual total abundance plotted with the mean summer upwelling index (NOAA 2014) is not significant ( $F\text{-ratio}=0.0816, p=0.7793, df=1,16$ ). Data includes only stations sampled every year.

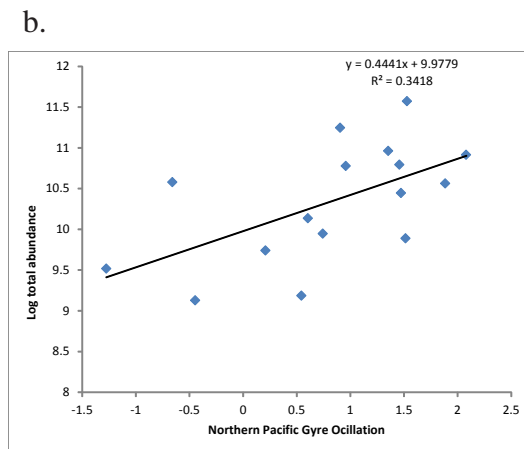
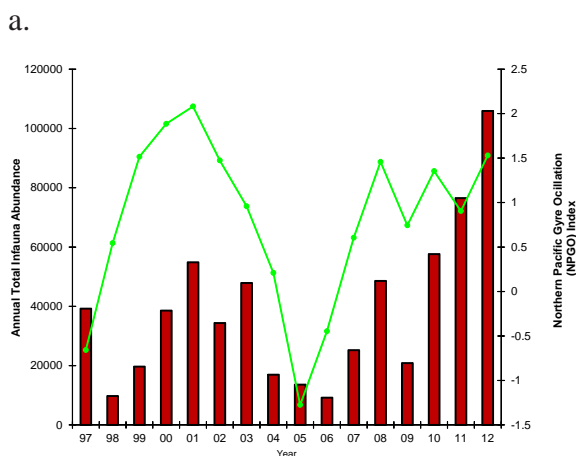


Figure 5-4a & b

Annual total abundance plotted with the North Pacific Gyre Oscillation is significant ( $F\text{-ratio}=7.268, p<0.05, df=1,16$ ). Data includes only stations sampled every year.

### 5.2.2. COMMUNITY PARAMETERS

Community measures of abundance, species richness (number of species), Shannon-Weiner diversity ( $H'$ ), and Pielou's evenness ( $J'$ ) were calculated and are shown in Figures 5-2 to 5-10. The values calculated for each station for the sixteen-year period are in Appendix E-3.

#### 5.2.2.1. Abundance

Annual infauna abundance measured at stations collected every year varied substantially and ranged from a low of 9,187 organisms in 2006 to a high of 105,883 organisms in 2012 (Figures 5-2a, 5-3a, 5-4a). During the last 3 years of the survey period infauna abundance offshore has increased greatly mostly due to the increase in abundance of the polychaete, *Spiophanes norrisi*. Over the study period, there has been an inverse correlation between the Corrected Delta Outflow Index (<http://www.water.ca.gov/dayflow/output/Output.cfm>) and infauna abundance (Figure 5-2a). Figure 5-2b shows

this relationship is significant (Linear regression F-ratio=8.5567,  $p=0.011$ ,  $df=1,16$ ), however statistical tests show that the relationship is not linear. The positive relationship between total abundance and the appearance and strength of the summer upwelling index documented in the previous summary report (NRLMD 2010a) did not persist with additional data (Figures 5-3a & b). The graphs also indicate that the infauna abundance is positively correlated with the Northern Pacific Gyre Oscillation Index (Figure 5-4a and b).

Abundance at individual stations varied annually and ranged from a low of 23 individuals at Station 41 in 1999 to a high of 13,007 individuals at Station 73 in 2012 (Figure 5-5). Overall, abundance was lower at stations in the Golden Gate group than at other stations in the study area. The greatest overall abundance was observed in 2012 at the southern reference stations (6, 60, 61, 62, 67, and 69), reef-effect stations (59, 73, 75, 77, 78), ancillary stations (4, 56) and outfall stations (01, 02, and 25).

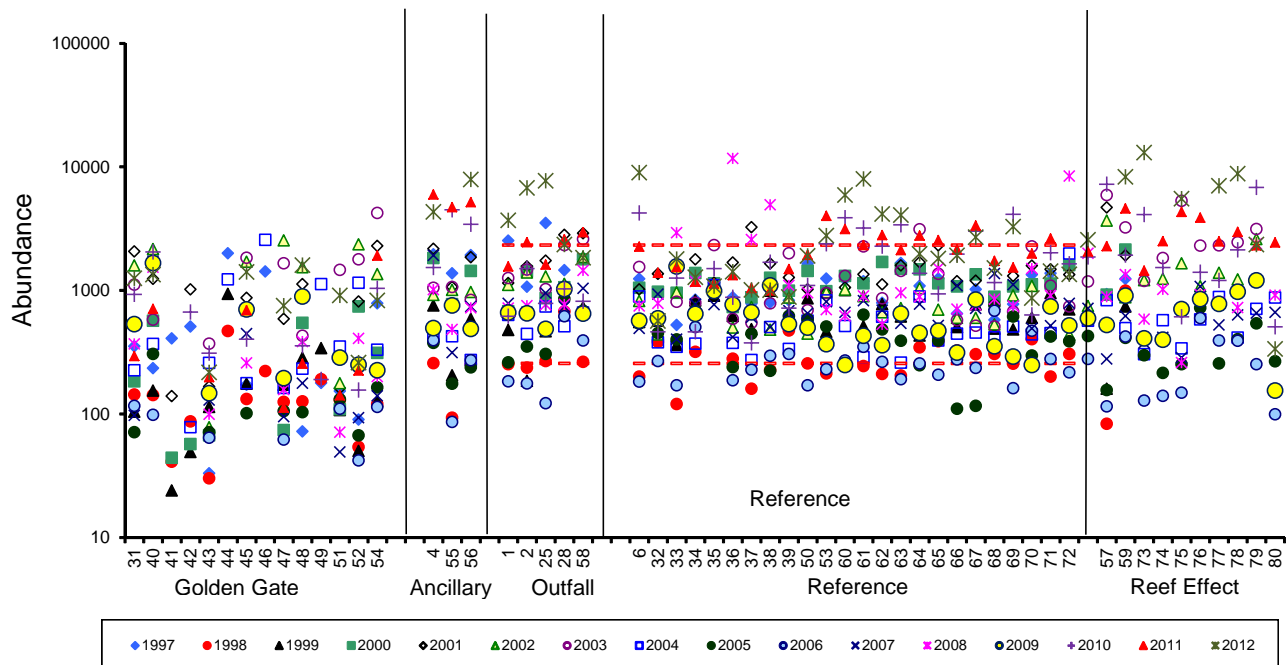


Figure 5-4

*Infaunal total abundance values plotted against a logarithmic axis by station groups. Year 2012 generally had the greatest abundance values per station, especially in the southern reference, reef effect, ancillary and outfall stations. Dashed red lines indicate the upper and lower tolerance interval bounds of the reference envelope analysis and are derived from 380 reference samples.*



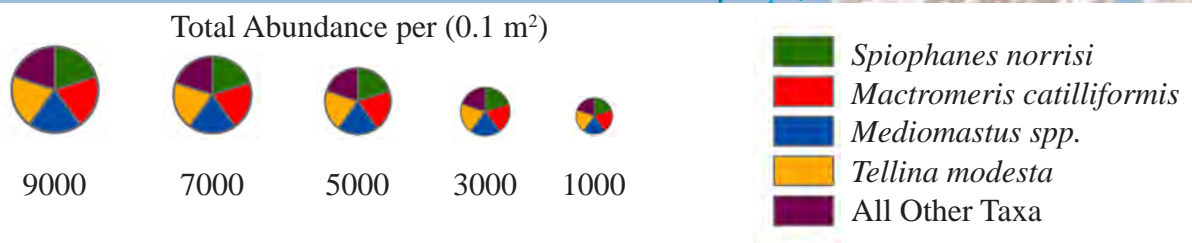
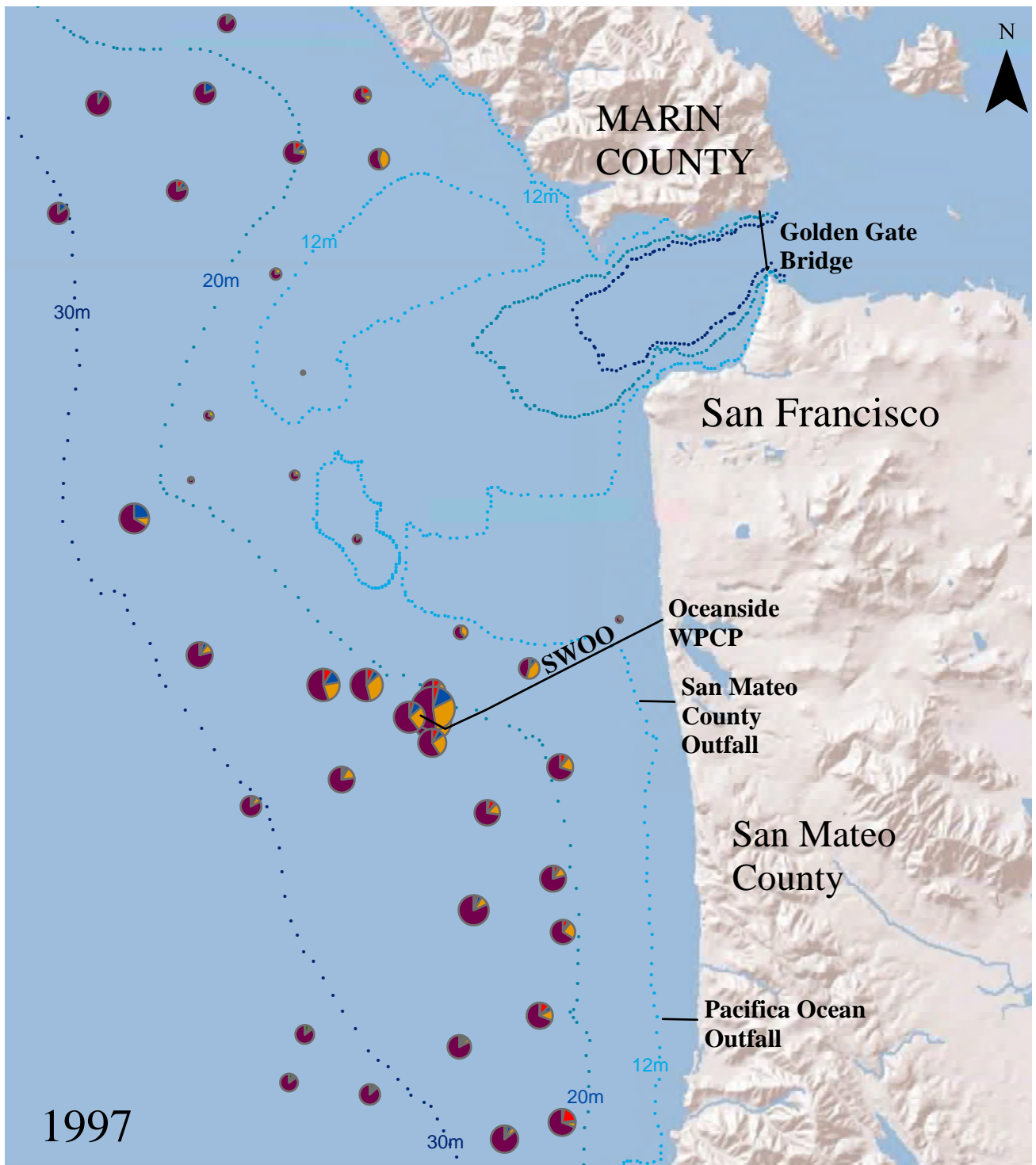
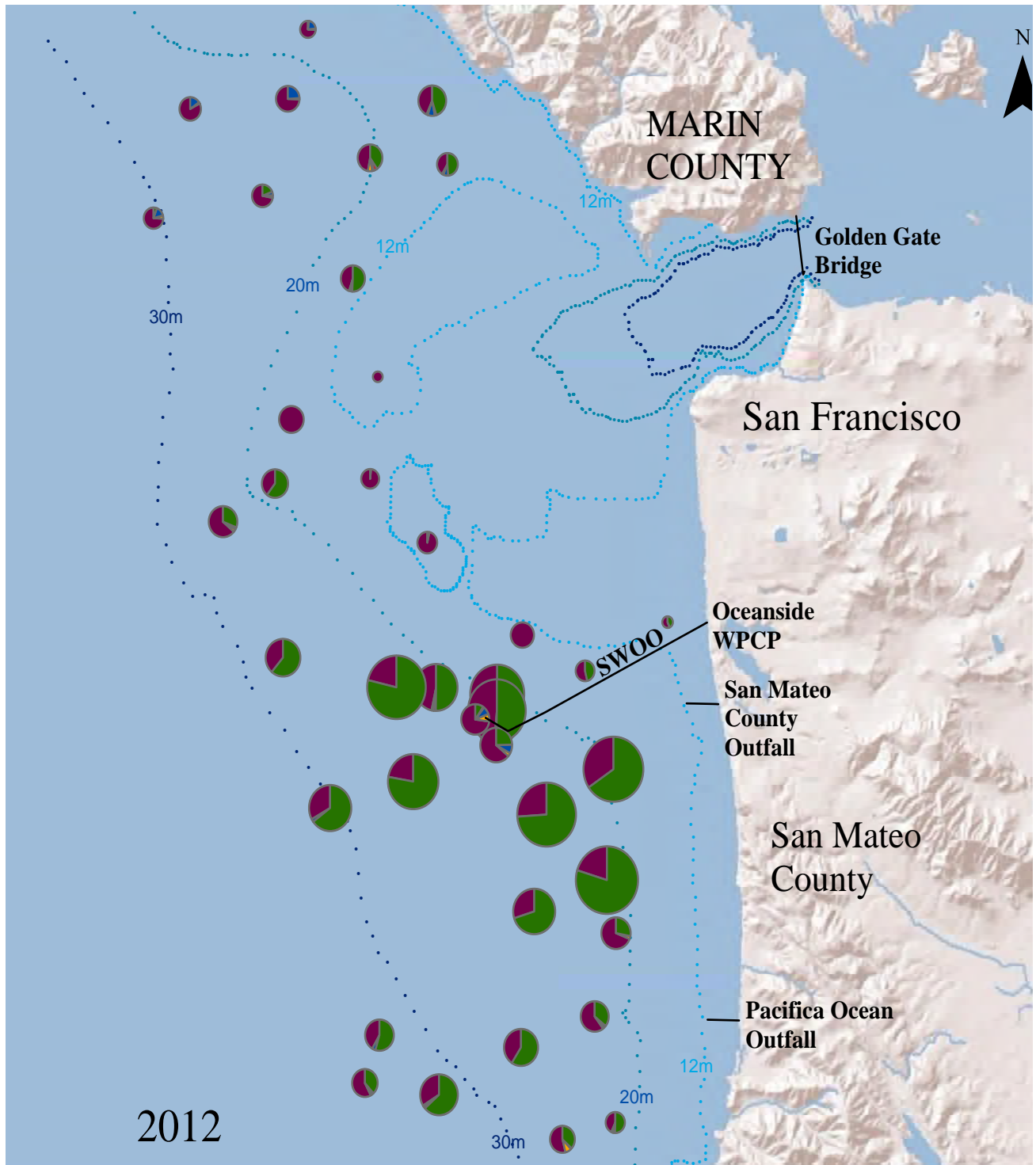


Figure 5-6a  
Percent composition and total abundance for the four most abundant species from 1997.



2012

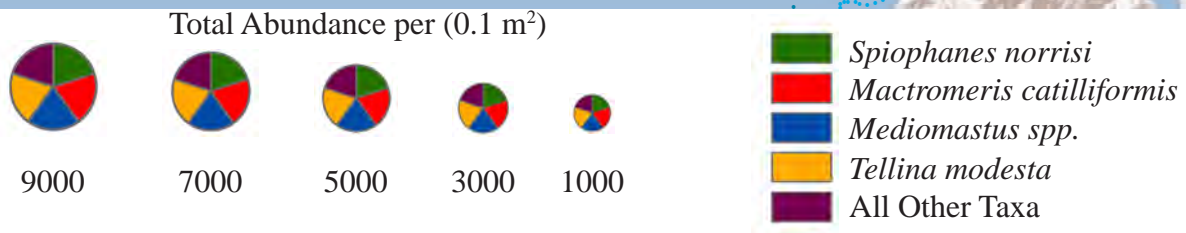


Figure 5-6b.

Percent composition and total abundance for the four most abundant species from 2012 (Does not include reef-effect stations).

This increase shows a trend of increased infaunal abundances and are largely due to the increase in the population of *Spiophanes norrisi*. This species was also responsible for the high abundances values at all of the station groups for 2010, 2011 and 2012. All of the station abundance values above 5,000 occurred during the 2003, 2008, 2010, 2011 and 2012 sampling years. In stations south of the outfall (particularly stations 59, 64 and 79), the polychaete, *Owenia collaris* was the most abundant species during 2003, with the exception of station 57, which had sustained high abundances of *Spiophanes norrisi*. In 2008, recruitment of the bivalve, *Mactromeris catilliformis* is responsible for the spikes in abundance at northern reference stations 36 and 38. Until 2009 no station was dominated by a single species for more than one year. Abundance distributions for 1997 and 2012

are shown geographically in Figure 5-6a and 5-6b respectively. These maps show the major changes in the abundance and dominant species.

#### 5.2.2.2. Diversity, Evenness, and Species Richness

The Shannon-Weiner diversity index primarily reflects both species richness and the evenness of the distribution of individuals among those species. Shannon-Weiner index values ranged from a low of 0.54 at ancillary station 55 in 2010 to a high of 3.91 at reference station 39 in 2000 (Figure 5-7). Similar to abundance values, species diversity was generally lower at Golden Gate stations in all years compared to other stations; however, low diversity values were apparent in all offshore stations (reference, outfall, reef effect and ancillary) during 2009, 2010, 2011 and 2012. The general pattern of diversity is shown in Figure 5-8. Stations with

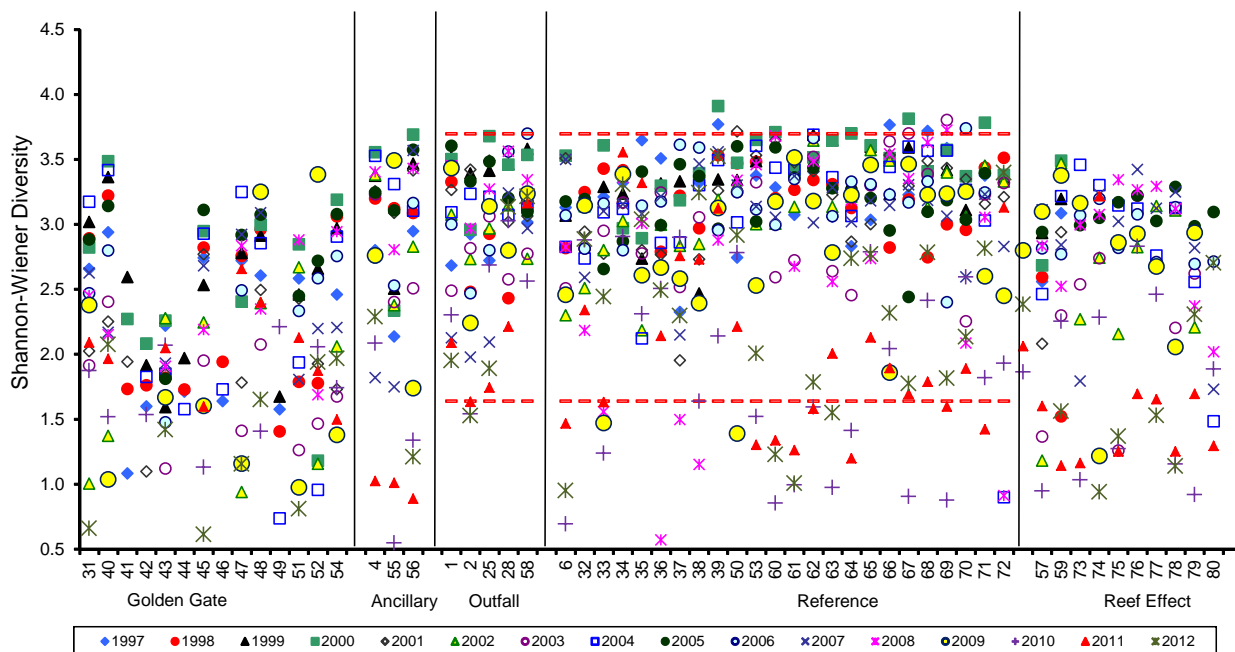


Figure 5-7

*Shannon-Weiner diversity index by station group for the trend period. During the first 16 years of the survey, diversity indices were lowest overall at the Golden Gate stations, however during 2009 through 2012 low values were seen at all stations. The offshore (ancillary, outfall, reference and reef effect) station's low values are caused by spikes in abundance of the polychaete, Spiophanes norrisi. Dashed red lines indicate the upper and lower tolerance interval bounds of the reference envelope analysis and are derived from 380 reference samples.*

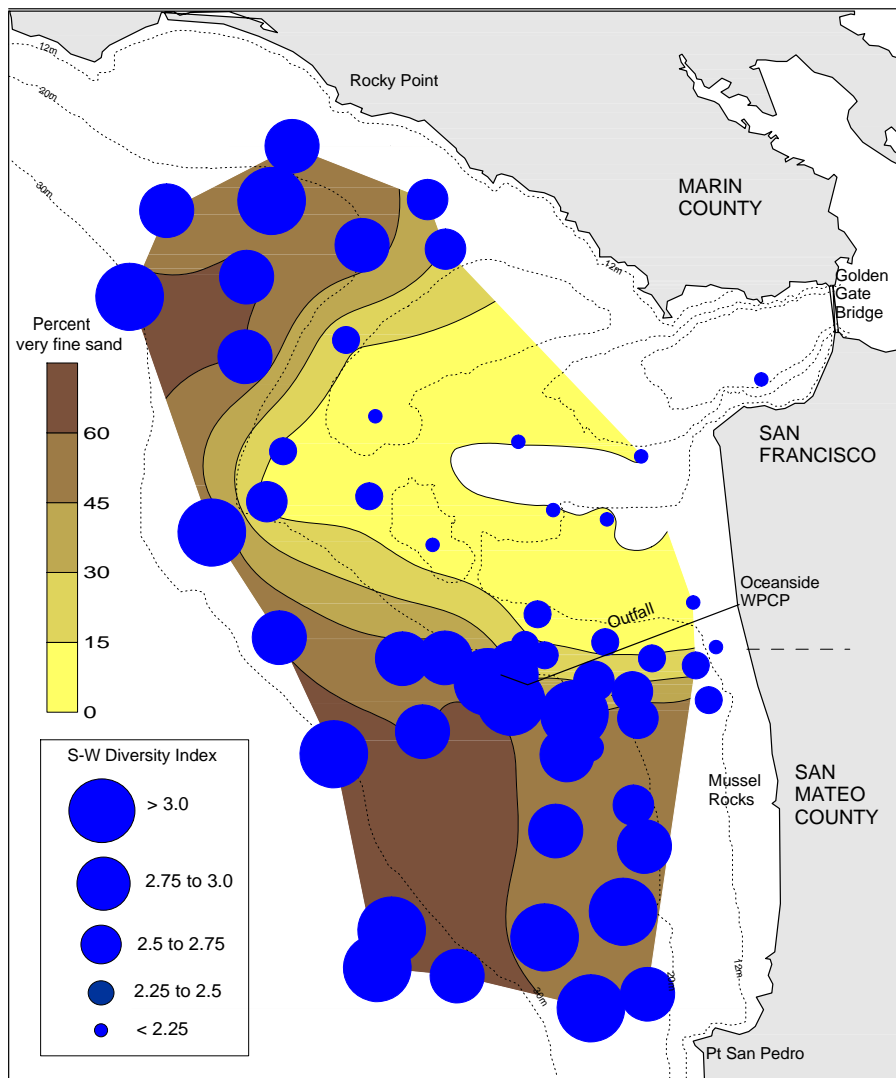


Figure 5-8

Geographic distribution of the average Shannon-Weiner Diversity index values over the trend period. Diversity distribution was not affected by relative annual abundance. Greatest diversity values were associated with very fine sands.

very fine sands usually have higher diversity values.

Pielou's evenness is a measure of the degree of equality in the distribution of individuals among species. As evenness approaches 1.0, the distribution of individuals among species is more equal. Evenness values ranged from a low of 0.12 at reference station 36 in 2008 (due to recruitment spike of *Mactromeris catilliformis*), to a high 0.94 at Golden Gate station 41 in 1999 (Figure 5-9). Low evenness values observed at stations 01, 02, 04, 25, 37, 55, and 73 were due to the dominance of a few species:

*Mactromeris catilliformis* and *Apoprionospio pymaea* which accounted for over half of the species observed at those stations in 2007. Station 80 also had low evenness values in 2007, however it was dominated by *Spiophanes norrisi*. These 3 species were the overall dominant taxa for 2007 as shown in Table 5-3. In general, low evenness values were observed at most stations in 2009, 2010, 2011 and 2012. Low diversity and poor evenness values are due primarily to the extraordinary numeric dominance of *Spiophanes norrisi* in 2010, 2011 and 2012.

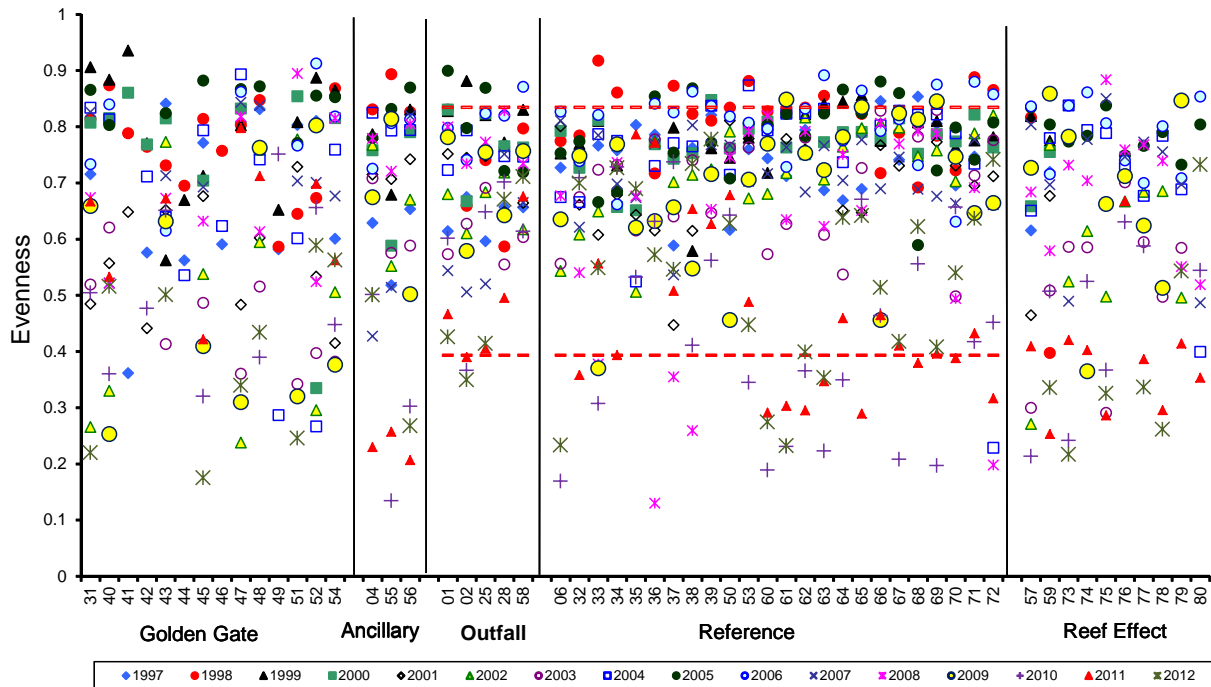


Figure 5-9

*Pielou's evenness values by station group. The pattern is similar to the S-W diversity plot. Dashed red lines indicate the upper and lower tolerance interval bounds of the reference envelope analysis and are derived from 380 reference samples.*

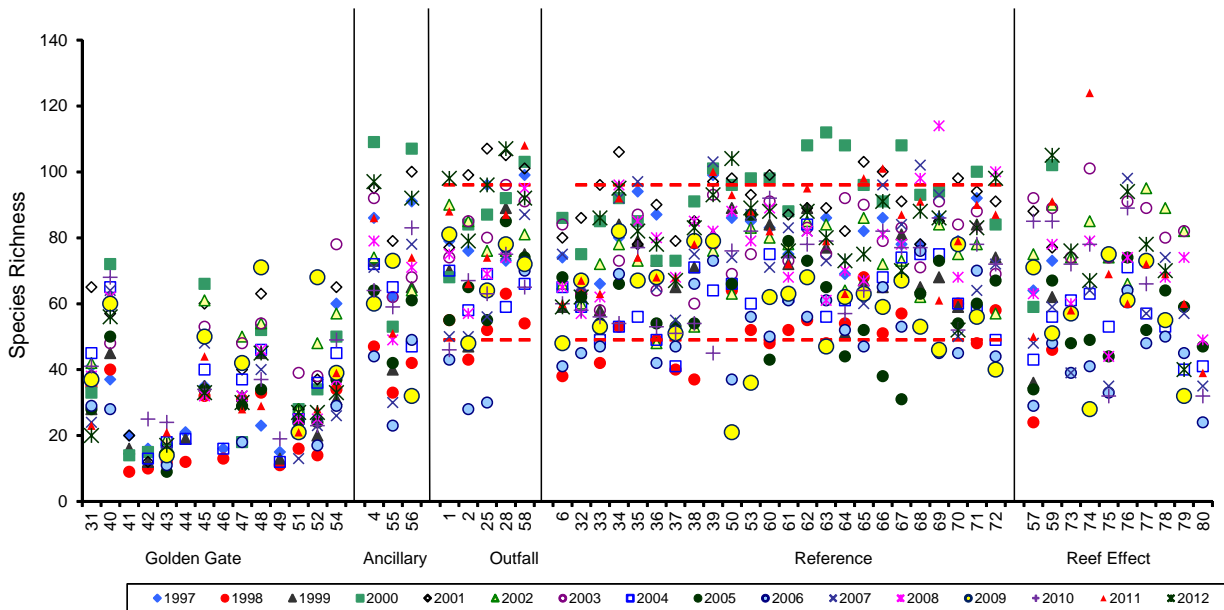


Figure 5-10

*Species richness (total number of taxa per station) values for the trend period. In general, Golden Gate stations had lower species richness values. Dashed red lines indicate the upper and lower tolerance interval bounds of the reference envelope analysis and are derived from 380 reference samples.*

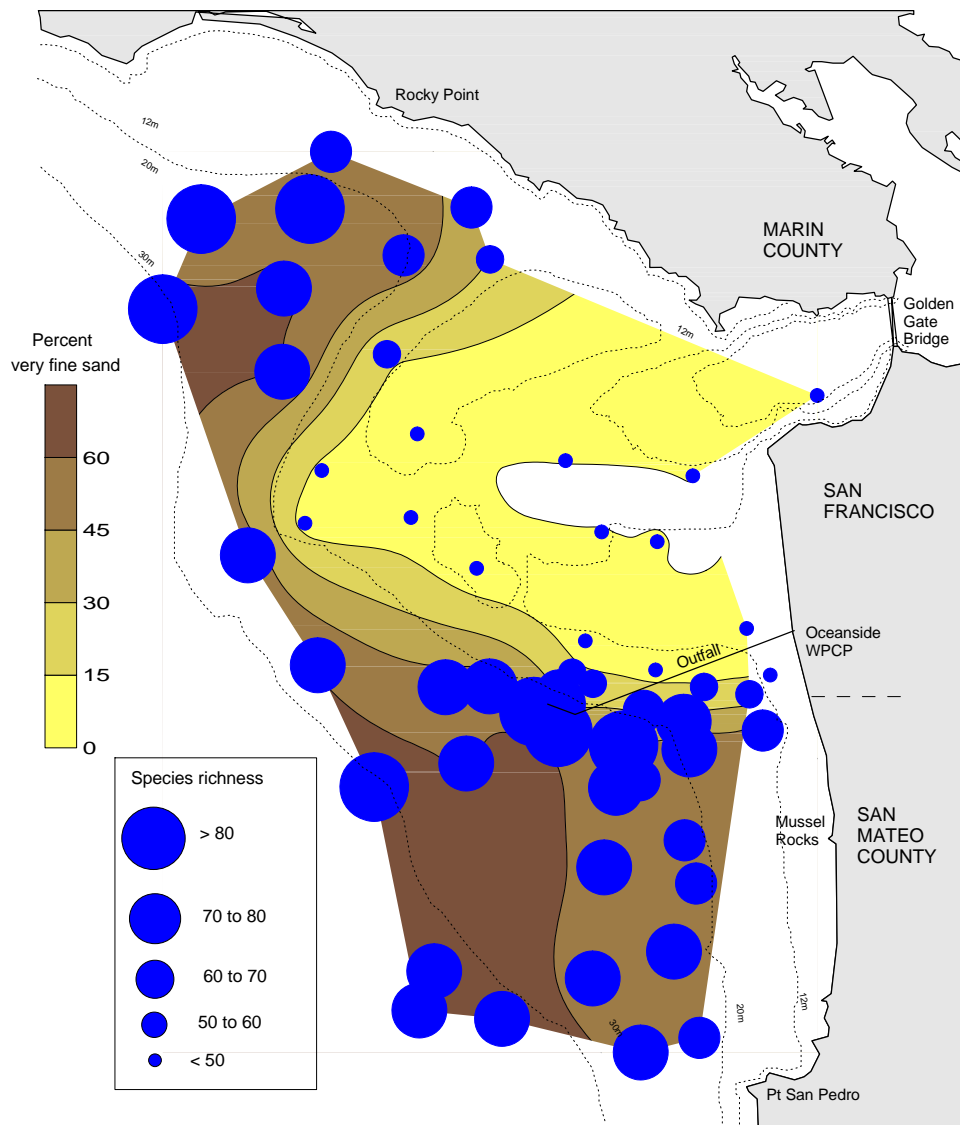


Figure 5-11

*Geographic distribution of the average Species richness values over the trend period. The distribution was not affected by annual richness values with the greatest species richness values associated with very fine sands.*

Species richness was variable between years and ranged from a low of 9 at Golden Gate station 41 in 1998, to a high of 124 at reef effect station 74 in 2011. As shown in Figure 5-10, the years with the overall lowest species richness are 1998 and 2006 corresponding to the years with the lowest overall abundance values. Generally, species richness values were lower at Golden Gate stations in all years. This general pattern for species richness has remained consistent each year with the fewest species collected at the Golden Gate stations, and the number of species

increasing with increasing percentages of very fine sand. Species richness follows a similar geographic pattern as diversity (Figure 5-8).

### 5.2.3. COMMUNITY PATTERNS

#### 5.2.3.1. Ordination Analysis

Ordination analysis can be used to assess differences in benthic invertebrate assemblage structure between outfall and reference conditions (Bernstein and Smith 1986). By examining assemblage structure at reference

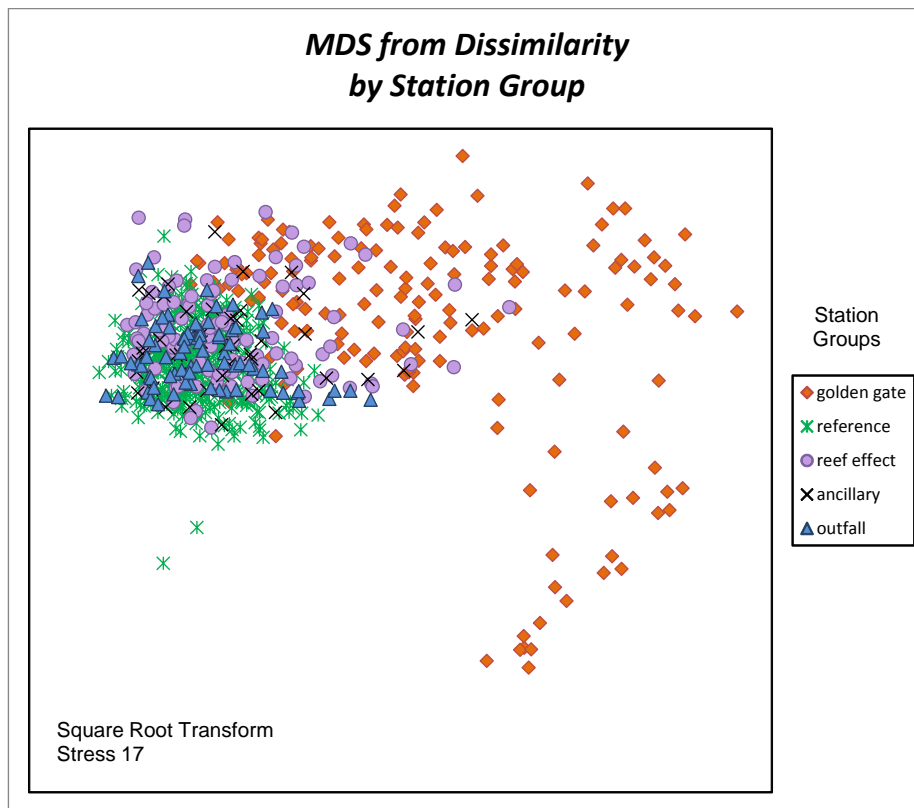


Figure 5-12

Two dimensional NMDS (non-metric multidimensional scaling) plot of the benthic infauna community. The plot shows the distribution of each station type. The plot shows that outfall stations (blue triangles) are distributed among reference (green squares) and reef effect stations (purple circles).

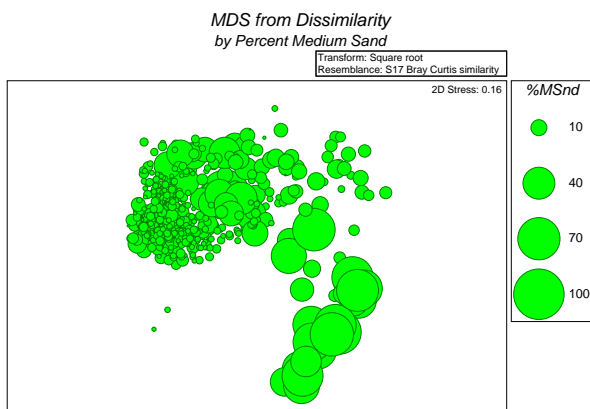


Figure 5-13

Sediment medium sand grain size data superimposed on the benthic infauna community two-dimensional NMDS plot (Figure 5-12). The medium sand distribution primarily corresponds to the Golden Gate stations and is opposite of the very fine sand grain size plot (Figure 5-14).

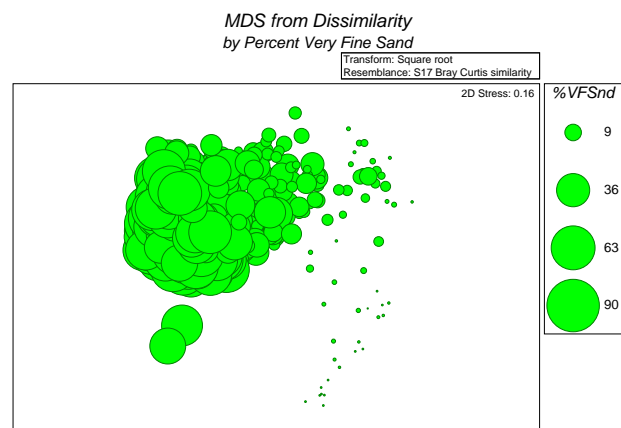


Figure 5-14

Sediment very fine sand grain size data superimposed on the benthic infauna community two-dimensional NMDS plot (Figure 5-12). The very fine sand distribution includes both reference and outfall stations and is opposite of the medium sized sand grainsize plot (Figure 5-13).

stations compared to those at outfall stations through time, a determination can be made on the departure or convergence of outfall and reference conditions (Bray and Curtis, 1957).

The ordination technique (Methods, 2.4.2. Multivariate Analyses, and Appendix B) employed non-metric multi-dimensional scaling

(NMDS), which defines gradients of community similarity by utilizing a matrix of values derived from the Bray-Curtis index of dissimilarity (see Appendix B.2.1.1.). The resulting 2-dimensional plot is shown in Figure 5-12.

When the sediment grain size data are overlaid onto the biotic NMDS bubble plots, they show that the distribution of the two parameters (medium sand and very fine sand) have opposite patterns (Figure 5-13 and Figure 5-14). The horizontal axis of the NMDS plot has a very strong relationship with the very fine sand sediment component while the vertical axis has a relationship with the medium sand content.

The NMDS plot did not correspond well with any of the other abiotic parameters including sediment total PAHs, organics or metals.

### 5.2.3.2. Cluster Analysis

Cluster analysis (Appendix B.2.1.3) produced a hierarchical dendrogram of sample relationships that, like the ordination analysis, was based upon Bray-Curtis similarity index scores. The distance between samples on the dendrogram reflects the degree of dissimilarity in species composition and abundance. Based upon examination of the dendrogram cluster groups were chosen such that samples within a cluster group were more similar to each other than they were to samples in other cluster groups.

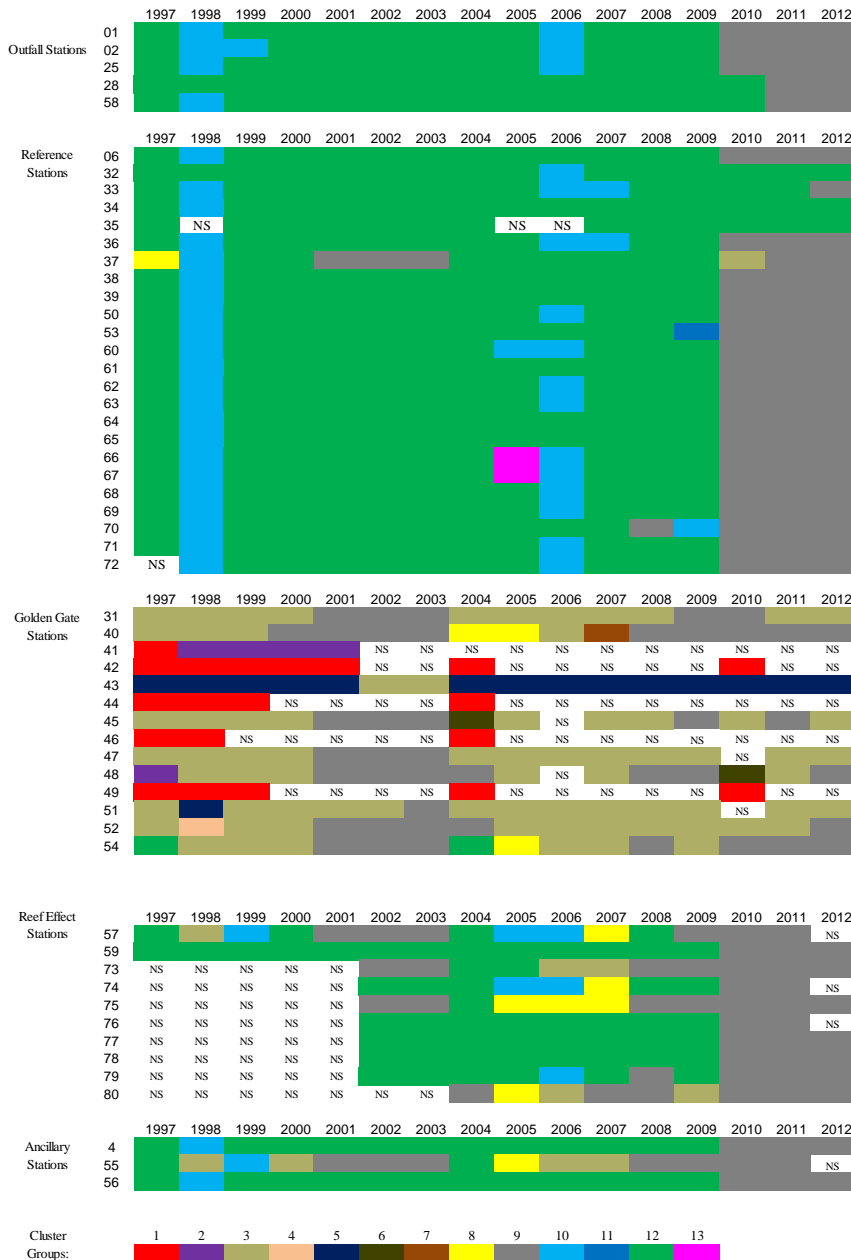


Figure 5-15

Station cluster group matrix showing the biological affinities of stations through time. Samples within a cluster group are more similar to each other biologically than they are to samples in other cluster groups.



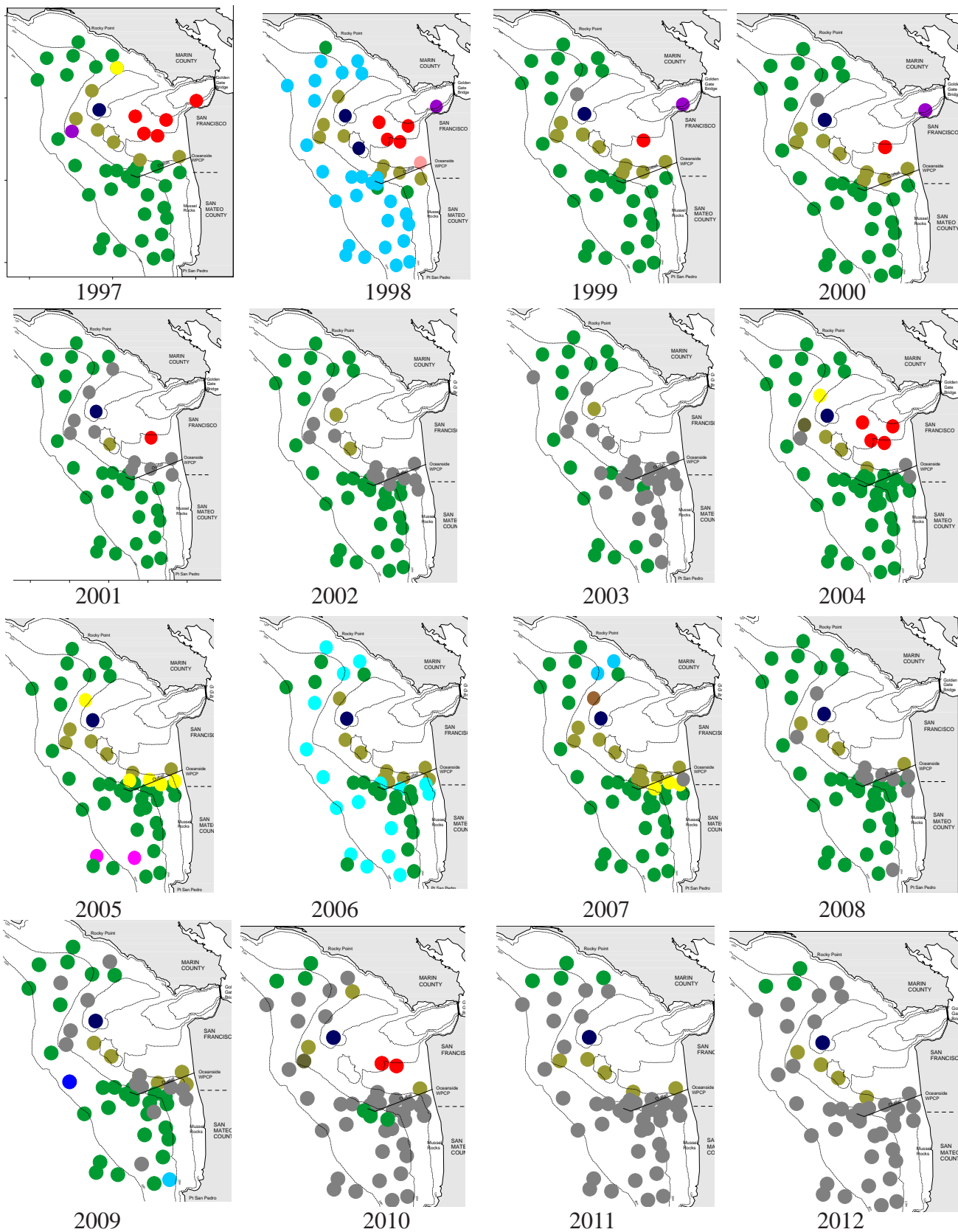


Figure 5-16

Geographic distribution of the benthic community cluster groups 1997-2012. See figure 5-4 for cluster group color codes



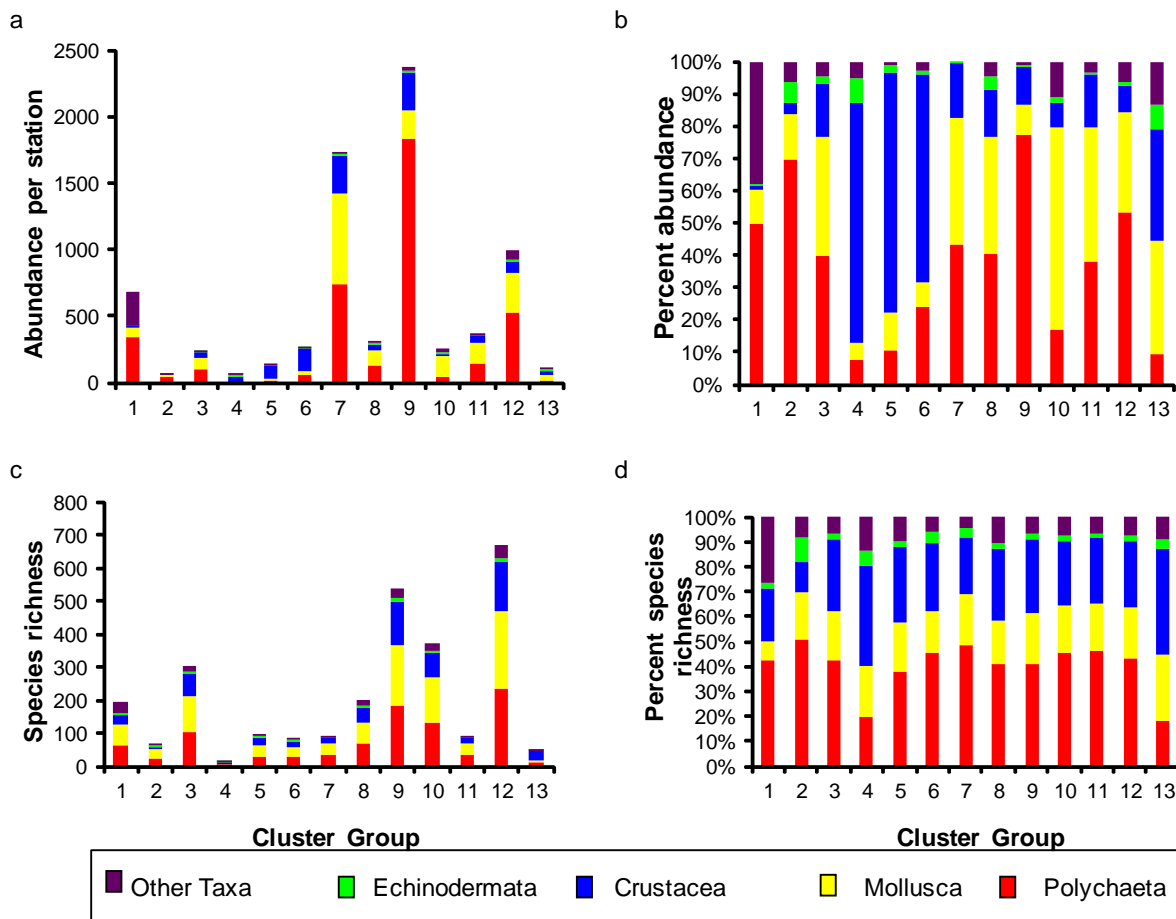


Figure 5-17

Graphical representation of abundance and species richness composition of major taxa in each cluster group (both actual amounts and percent composition)

by abundant polychaete species such as *Mediomastus* spp. and *Scoletoma luti* (Figure 5-17c and Table 5-4).

Cluster group 9 generally replaced cluster group 12 offshore in 2010 through 2012 except for a few northern reference stations. The taxa identified within cluster group 9 constitute a benthic infauna community associated with very fine sand and is numerically dominated by the polychaete, *Spiophanes norrisi*.

Cluster groups 2, 3 and 5 contained samples primarily from on or near the sand bars. These stations are the shallowest in the study area and are, therefore, subject to the most wave disturbance. The sediments at these stations are predominantly well-sorted fine sand, averaging

greater than 50% fine sand for the sixteen-year period (Section 4.2.1.). Cluster “groups” 4, 7 and 11 and were each composed of a single sample and were respectively dominated by a single species: the amphipod crustacean, *Americhelidium shoemakeri*, (Station 52 in 1998) a bivalve mollusk, *Mactromeris catilliformis* (Station 41 in 2007) and the gastropod mollusk, *Astrix gausapauda* (Station 53 in 2009)(Table 5-4). Of the three sandbar cluster groups, cluster group 2 had the lowest abundance, averaging 64 organisms per station (Figure 5-17a). Groups 4, 5 and 6 of the Golden Gate sandbar stations had substantial crustacean populations (Figure 5-17b, Table 5-4). The increased relative abundance of crustaceans at the sand bar

stations is consistent with a pattern of zonation at a high-energy subtidal beach in Monterey Bay described by Oliver et al. (1980) where crustaceans were more abundant in the areas of greatest wave disturbance.

Cluster group 1 consisted of stations 42, 44, 46, and 49 for all years sampled and station 41 in 1997. These stations had the coarsest grain size in the study area, averaging greater than 50% coarse and medium sands (Section 4.2.1). The taxa identified with cluster group 1 constitute a distinct benthic infauna community associated with coarse and medium sands and are numerically dominated by two small, interstitial-like polychaetes, *Hesionura coineaui difficilis* and *Heteropodarke heteromorpha* and the bivalve *Tellina nuculoides* (Table 5-4).

#### 5.2.4. REFERENCE ENVELOPE

Reference envelope analysis is a method by which indicators from potentially impacted sites (e.g. outfall stations) can be compared to a range of indicator values from reference sites (see Reference Envelope Analysis 2.4.2.3. and Appendix B 2.4.). The analysis involves computation of tolerance-interval bounds to define limits distinguishing reference from non-reference conditions (Smith 2002). Outfall station indicator values that fall outside the envelope, defined by the tolerance-interval bounds, represent a potential impact. Figures 5-4, 5-7, 5-9, and 5-10 show infauna indicators plotted with reference envelope tolerance-interval bounds for the reference and outfall station groups.

Because the Pearson/Rosenberg model predicts that abundance and diversity measures at sewage discharge sites increase and then decrease along an impact gradient, both upper and lower tolerance-interval bounds were plotted for species richness, diversity and abundance. Lower tolerance limits changed noticeably since the last trend report for diversity and evenness since reference stations were affected by the low values in 2010, 2011 and 2012.

Several outfall and reference stations fell outside reference conditions for each community

measure. High abundance at outfall stations 01 and 25 in 1997, relative to reference stations, was due to recruitment by the bivalve *Tellina modesta* (WQB 1998). High abundances at some outfall stations in 2001, 2003, 2010, 2011 and 2012 were matched by high abundances at several reference stations in those years. Similarly, low abundances at outfall stations in 2006 were matched by low abundances at several reference stations in 2006. Outfall station 02 and some southern reference, ancillary and reef effect stations had diversity and evenness values that fell below the lower tolerance-interval bound in 2010, 2011 and 2012. During those years, some of the extremely low values for diversity and evenness were at reference, ancillary and reef effect stations. Species richness was outside of reference conditions at outfall stations 02 and 25 and reference stations 06, 32, 36 and 50 in 2006. The changes in the communities in 2010 to 2012 did not seem to affect the species richness. None of the outfall stations differed greatly from reference conditions each year. There has not been a persistent pattern that would indicate outfall stations have altered community measures relative to reference stations. The reef effect stations fell outside of the reference conditions to a greater extent than the outfall stations, especially in 2002, 2003, 2006, 2007, 2009, 2010, 2011 and 2012. These deviations may be due to the effect of a region-wide recruitment of *Spiophanes norrisi* during those years in 2002, 2003, 2009, 2010, 2011 and 2012; and by various species in 2006 and 2007 (see Tables 5-3 and 5-4).

Ninetyfive percent of the time, ten percent of the reference population might exceed a tolerance-interval bound in the absence of an impact with the parameters used in this analysis (Methods Section 2.4.2 Multivariate Analyses), thus high abundance at some reference (and outfall) stations may occur naturally. The high abundances at some stations in 1997, 2000, 2001 and 2004, as indicated by the reference envelope analysis, most likely represent haphazard recruitment events as discussed previously under abundance (Section 5.2.2.1.). Sustained

high abundance values in 2009, 2010, 2011 and 2012 are due to increased *Spiophanes norrisi* populations.

### 5.2.5. BACIP ANALYSIS

The BACIP (Before-After-Control-Impact-Paired) statistical model (BACIP Analysis 2.4.2.3.) was used to test whether the relationship between outfall and reference station abundances has changed since the onset of treated wastewater discharge from the SWOO. The null hypothesis states that the mean differences between the outfall and reference station abundances are the same in pre-discharge and discharge periods. The test assumes that natural temporal changes in abundance over time will be reflected at both outfall and reference stations, thus the mean differences in abundances in the two areas should not change. On the other hand, if the wastewater discharge causes a change in abundance, the mean differences between outfall and reference stations will change. Stations 01 (outfall) and 06 (reference) were used in the model because they have the longest history of continuous collection. The results, shown in Table 5-5 indicate that the SWOO discharge has not affected infauna abundances in the study area. Figure 5-18 shows the pattern of infauna abundance at the outfall and reference stations over time.

### 5.2.6. PEARSON/ROSENBERG MODEL

The Pearson/Rosenberg model (Pearson & Rosenberg 1978) predicts a pattern of change in the infaunal community in response to an organic enrichment source. At the point source, there is a zone of low faunal abundance and species richness. Further from the source is a region of high abundance comprised of mostly opportunistic species. Species richness will also increase in this zone, but reach its maximum well outside the influence of the point source.

The SWOO benthic infauna data did not show results that the Pearson/Rosenberg model would predict for an outfall with a significant impact. Although some outfall samples in some

Table 5-5

*BACIP (Before and After Control Impact Paired) analysis of log<sub>10</sub> total abundance assesses the impact of a pollution source compared to baseline data collected prior to construction of the outfall. The null hypothesis is that the differences in outfall and reference means are equal in the pre-discharge and discharge periods tested. The null hypothesis is accepted (paired t=-1.61, df=28, p=0.12).*

	Pre-Discharge 1982 - 1986 n=4	Discharge 1986 - present n=26
Outfall Station 01	3.102	2.956
Reference Station 06	2.818	2.926
difference	0.284	0.029

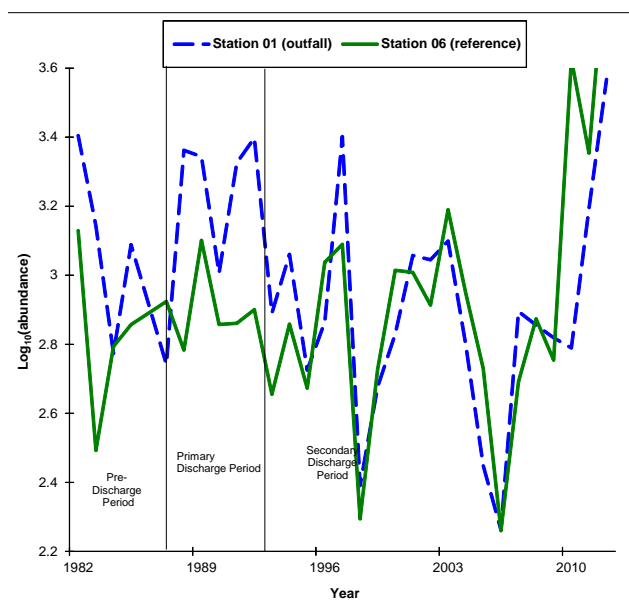


Figure 5-18

*Pattern of infauna abundance (log<sub>10</sub>) at outfall (station 01) and a reference station (station 6). After the start of secondary treatment in 1993, the distribution pattern between the two stations are very similar (average difference of log<sub>10</sub> abundance values = 0.05).*

years had higher abundances than reference stations, most outfall samples were within the range for reference conditions. The high abundances at outfall stations 01 and 25 in 1997 were associated with relatively low diversity and evenness (Figure 5-5), suggestive of an opportunist zone in the model, but that situation was due to massive recruitment of the bivalve *Tellina modesta* (Table 5-2) and did not persist. Similarly, low diversity and evenness values were observed at outfall stations 01, 02 and 25, in 2010, 2011 and 2012, but, as shown in Table 5-4, these were due to large abundances of *Spiophanes norrisi*. Neither species is known to be an indicator of enrichment or other impacts typically associated with wastewater discharges. High abundances at outfall stations in 2001, 2003, 2009, 2010, 2011 and 2012 were reflected in the reference stations (Figure 5-5), suggesting that the increase in abundance was region-wide. It is possible that the SWOO data do not fit the Pearson/Rosenberg model in a classic sense. This is probably due to the relatively low volume and high quality of effluent that is discharged and the high-energy regime of the receiving water environment.

### 5.3. SUMMARY AND CONCLUSIONS

The multi-year trend analysis shows a distinct change in the offshore community in the last 3 to 4 years of the survey period. This change is also reflected in the sediment data (see section 4.2.1.3) and seems to be region-wide, rather than due to a particular point source.

Ordination analysis of all stations shows that sediment grain size is the most important factor in structuring of infauna communities in the study area. Ordination and Cluster analysis indicate that stations in the reference, outfall, reef effect, and ancillary station groups are very similar to one another, with similar changes in community composition from one year to another.

Cluster analysis identified several distinct benthic infauna communities. The area inside the barrier sand bars, with predominantly

coarse and medium sands, has a distinct community of mostly interstitial organisms. Another community exists on or near the sand bars, which is dominated by Crustacea and by the polychaete, *Spiophanes norrisi* especially in 2002 and 2003. Composition of the communities may change from year to year depending on large scale climatic changes either within a year (as with the El Niño-La Niña years, 1997 to 1998 and periods of large Delta outflow in 1998 and 2006) or over a period of several years (as with the change to the *Spiophanes norrisi* dominated community observed from 2010 to 2012).

Reference envelope analysis shows that some outfall stations have been outside tolerance-interval bounds for some infauna community measures each year. Higher abundance values at outfall stations might be suggestive of enrichment; however similar high abundance values are found in reference stations. The species responsible for high abundance at the outfall are not known indicators of enrichment or any other impacts typically associated with wastewater discharges. A comparison of infauna abundance at an outfall and reference site spanning periods before and after effluent discharge demonstrated no statistically significant difference.

Examinations of long-term SWOO benthic infauna data (Kellogg et al 1998; NRLMD 2010a) have found that seasonality and oceanographic influences were major factors affecting the infauna community structure in the study area. This report supports the observations made by previous studies and has correlated some of these oceanographic influences with the observed patterns. The current monitoring supports the regional strategy for evaluation of point source impacts showing that the response in the benthic community near the outfall is reflective of the changes seen region-wide.



## **SECTION 6**

### **DEMERSAL FISH AND EPIBENTHIC INVERTEBRATES**

# DEMERSAL FISH AND EPIBENTHIC INVERTEBRATES

## 6.1 INTRODUCTION

The City and County of San Francisco has conducted trawl sampling in the Gulf of the Farallones over 26 years, from 1982-2008. Trawl sampling has been required in NPDES permits as a means of characterizing the resident fish and epibenthic invertebrate assemblages. NPDES permit requirements have included monitoring of these organisms to demonstrate that the community within the influence of the discharge is not degraded, and that a balanced indigenous population exists within and beyond the zone of initial dilution. Additionally, trawl sampling has been a means of collecting English sole (*Pleuronectes vetulus*) specimens in order to monitor tissue concentrations of pollutants (bioaccumulation) in demersal fish near the SWOO outfall as well as in reference areas.

The number, locations and seasonality of

trawl sampling have varied over the life of the monitoring program (Appendices F-1 and F-2, Figure 6-1), however, beginning in 2003, NPDES permit requirements reduced fishery sampling to single trawls at one outfall station (Station 01) and one reference station (Station 06). These two fixed locations have been sampled consistently, with at least one trawl each fall that sampling was conducted.

Under the adaptive management provisions of the NPDES permit Monitoring and Reporting Program, and with notification to the U.S. EPA, trawl sampling was curtailed in 2009 due to the listing of longfin smelt (*Spirinchus thaleichthys*) as a threatened species by the California Department of Fish and Wildlife (CDFW). Longfin smelt were commonly caught by the SWOO monitoring program during trawl sampling as by-catch (Table 6-1).

We argue to drop the trawl sampling requirement in light of the information gathered through trawl sampling to date, using the reported results of over two decades of monitoring. A discussion of each of these reasons for permanently discontinuing the trawl program is provided below:

Table 6-1

*Recorded occurrence of Osmeridae, including the threatened longfin smelt, in SWOO Regional Monitoring Program community trawls.*

Year	Month	Jack Smelt	Longfin Smelt	Night Smelt	Whitebait Smelt	Osmeridae
1982	10		408	148		
1983	6		600	430		
1983	10		294	222		
1984	2	1	2			
1985	6					1
1994	3		57	43		
1994	7		1	20		
1994	9		97	49		
1995	3		1	33		
1995	9		72	28		13
1996	3		183	172		
1996	11		377		75	



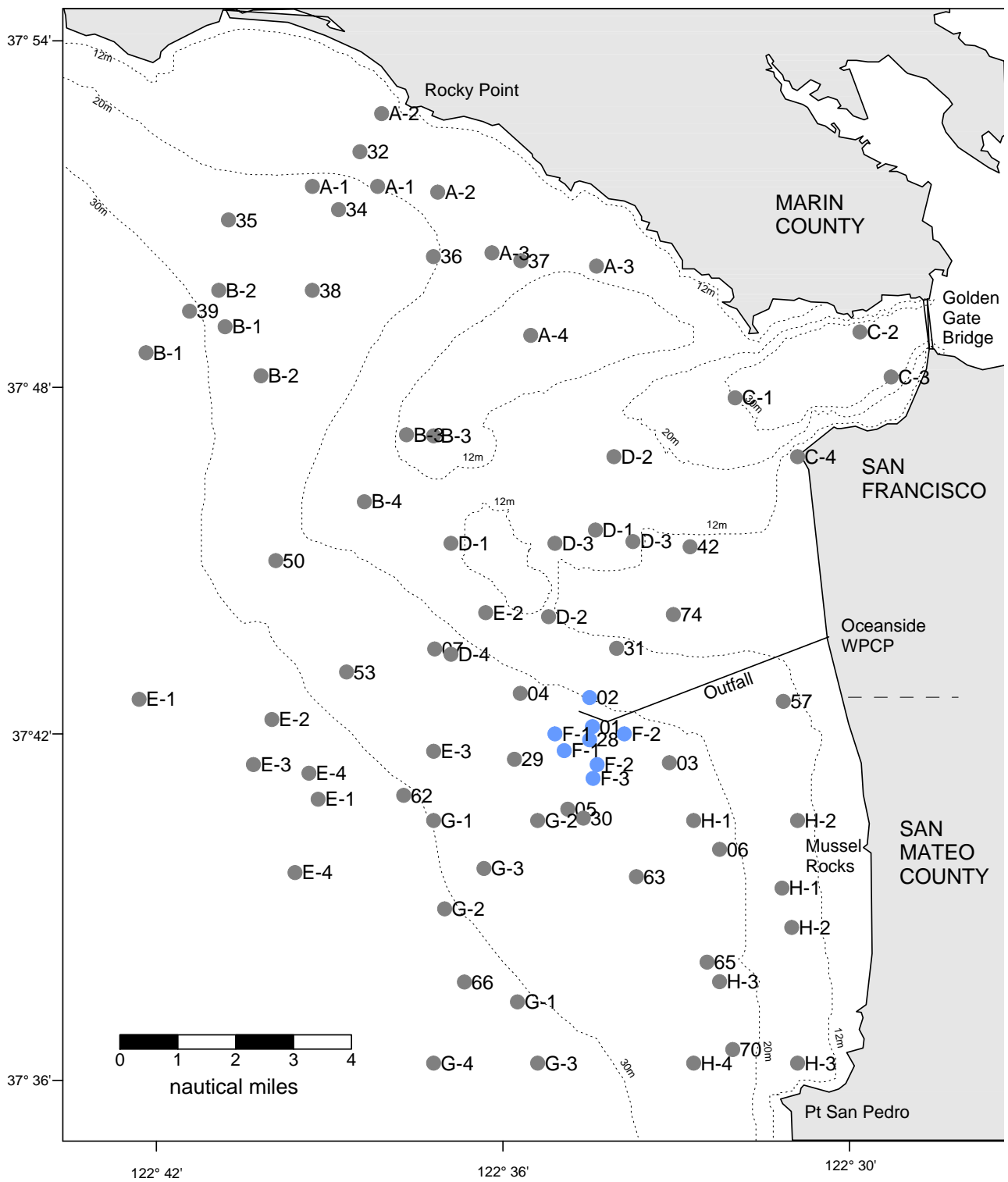


Figure 6-1  
 Coverage of the SWOO study area (1982-2008) by trawl samples used for community analyses.  
 Stations in blue were used to characterize outfall conditions

- 1) The trawl sampling has not revealed a significant difference between outfall area and reference area demersal fish and epibenthic invertebrate communities
- 2) The trawl sampling program is not suited to finding an outfall effect
- 3) The demersal fish specimens collected are not necessarily representative of contaminant exposure to consumers of local fishes or of body burdens obtained within the Gulf of Farallones
- 4) The trawl sampling results in significant and unnecessary mortality to demersal fish and epibenthic organisms including listed species
- 5) The trawl sampling destroys benthic habitat
- 6) Other new sources of high-quality environmental data are available
- 7) Given the absence of outfall effects demonstrated by data, the trawl program is excessively costly and burdensome to implement

## 6.2 DISCUSSION OF REASONS TO DROP THE TRAWL REQUIREMENT

### 6.2.1 THE TRAWL SAMPLING HAS NOT REVEALED A SIGNIFICANT DIFFERENCE IN OUTFALL AREA OR REFERENCE AREA DEMERSAL FISH OR EPIBENTHIC INVERTEBRATE COMMUNITIES

Twenty six years of monitoring have shown that there has been little difference between demersal fish and epibenthic invertebrate assemblages found at the outfall and in the reference areas. Measurements of community metrics (number of species, organism abundance, Shannon Weiner Diversity, and Pielou's Evenness) for demersal fish and epibenthic invertebrate assemblages show that no long-term trend of degradation

has been found at the outfall station (NRLMD 2010a). Likewise, the short- and long-term trends in community metrics for assemblages of demersal fish and epibenthic invertebrates from 1982-2008 show that community metrics values tended to be similar for the outfall and reference stations, despite substantial variability over the study period (NRLMD 2010a). There has also been a great deal of similarity between organism assemblage composition between reference and outfall locations since the SWOO went into operation in 1986 (NRLMD 2010a). The similarity in results from outfall and reference stations demonstrates that changes in the biota are representative of changes in the region as a whole, rather than an outfall-related effect.

### 6.2.2 THE TRAWL SAMPLING PROGRAM IS NOT SUITED TO FINDING AN OUTFALL EFFECT

There has been occasionally great variability in the trawl sampling record which results in very high variances in recorded assemblage composition even between replicate trawls at the same station. Much of the variability that has been observed is attributable to chance encounters with mobile species (e.g., bay shrimp *Crangon* spp. or market squid, *Loligo opalescens*). An analysis of seven years of SWOO trawl data (Niemi & Warheit 1989) found strong temporal and spatial effects associated with all variables and concluded that it would be statistically difficult to show any discharge effect, even if one existed. They further concluded "In effect, to properly control for effects other than sewage effluent, sampling would have to occur almost year-round, an unreasonable and costly suggestion. This problem is by no means unique to the data collected in this monitoring program, and has been discussed by other authors for other areas [Richkus 1980, Seger and Stamman 1986]."

### 6.2.3 THE DEMERSAL FISH SPECIMENS COLLECTED ARE NOT NECESSARILY REPRESENTATIVE OF CONTAMINANT EXPOSURE TO CONSUMERS OF LOCAL FISHES OR OF POLLUTANT BODY BURDENS OBTAINED WITHIN THE GULF OF THE FARALLONES

Tissue samples have been collected from two commercially important fisheries in the San Francisco Bay Area, English sole *Pleuronectes vetulus* and Dungeness crab *Metacarcinus magister* (= *Cancer magister*). English sole were collected from trawl samples taken at Stations 06 in the reference area (and additional stations when necessary) and from Stations 01, 02, and 28 in the outfall area. Dungeness crab were collected, using crab traps, at reference Station 06 and outfall Station 01.

Bioaccumulation of pollutants was measured in fish and crab muscle tissue as well as in fish liver and crab hepatopancreas but crab tissue seems to yield higher quality data. Analysis of bioaccumulation data collected by the SWOO Regional Monitoring Program is reported in detail in Section 7 of this report, but the general trends found to date are relevant here. In general, no persistent significant differences in contaminant concentrations have been found between tissues collected from outfall stations and reference stations. No discernible trends in the concentrations of contaminants in fish or crabs from reference or outfall areas have been observed over 14 years of bioaccumulation monitoring. To date, a broader suite of pollutants have been observed in crab tissue than fish tissue in the SWOO regional monitoring record and all contaminant compounds that have generally been observed in fish tissue have been observed in crab tissue as well (NRLMD 2010a). Additionally, most pollutants that have been observed are found at higher concentrations in crab tissue than in fish tissue (NRLMD 2010a), implying that crab tissue can give a better indication of trends of environmental pollution in the monitoring region.

Contaminant concentrations found in English sole tissue by the SWOO Regional Monitoring Program are not necessarily representative of contaminant exposure to consumers of local fishes or body burdens obtained within the Gulf of the Farallones. Historically, English sole specimens gathered in the trawl samples have been relatively small (30-250 mm), and English sole of “marketable size” (greater than 279 mm) have never been caught in a SWOO Regional Monitoring Program trawl sample (Figure 6-2). Fish used for bioaccumulation tissue samples should be at least 135 mm in length, though fish greater than 200 mm in length are preferred for bioaccumulation monitoring and these are typically a very small component of the catch (Figure 6-2). In recent years the mean standard length for English sole used for bioaccumulation monitoring has been as low as 86 mm, due to difficulty in finding fish of a more suitable size. Contaminants found in the tissues of such small fish are likely to be influenced by contaminant concentrations in their nursery habitat (the San Francisco Estuary) rather than offshore regions where they migrate as they mature (Lassuy 1989, Pearson et al. 2001). A special study of the San Francisco Estuary Regional Monitoring Program for Water Quality (RMP) (Melwani et al. 2009) attempted to compare fish bioaccumulation within San Francisco Bay and the adjacent open coast by comparing data from the RMP (for San Francisco Bay) and from the South West Ocean Outfall Regional Monitoring Program (open coast). They concluded that “Lower [mercury] concentrations for this species [English sole] may be due to their small size, wide foraging range, and earlier lifestage than other species and locations included in this study. Due to their size, comparison of these data to OEHHA’s fish consumption guidelines was not appropriate.” By contrast, the Dungeness crabs that are collected for bioaccumulation analysis generally meet the criteria for commercial catch (legal sized, male). The SFPUC is not arguing to discontinue bioaccumulation monitoring of Dungeness crab.

The SFPUC was able to substitute Pacific

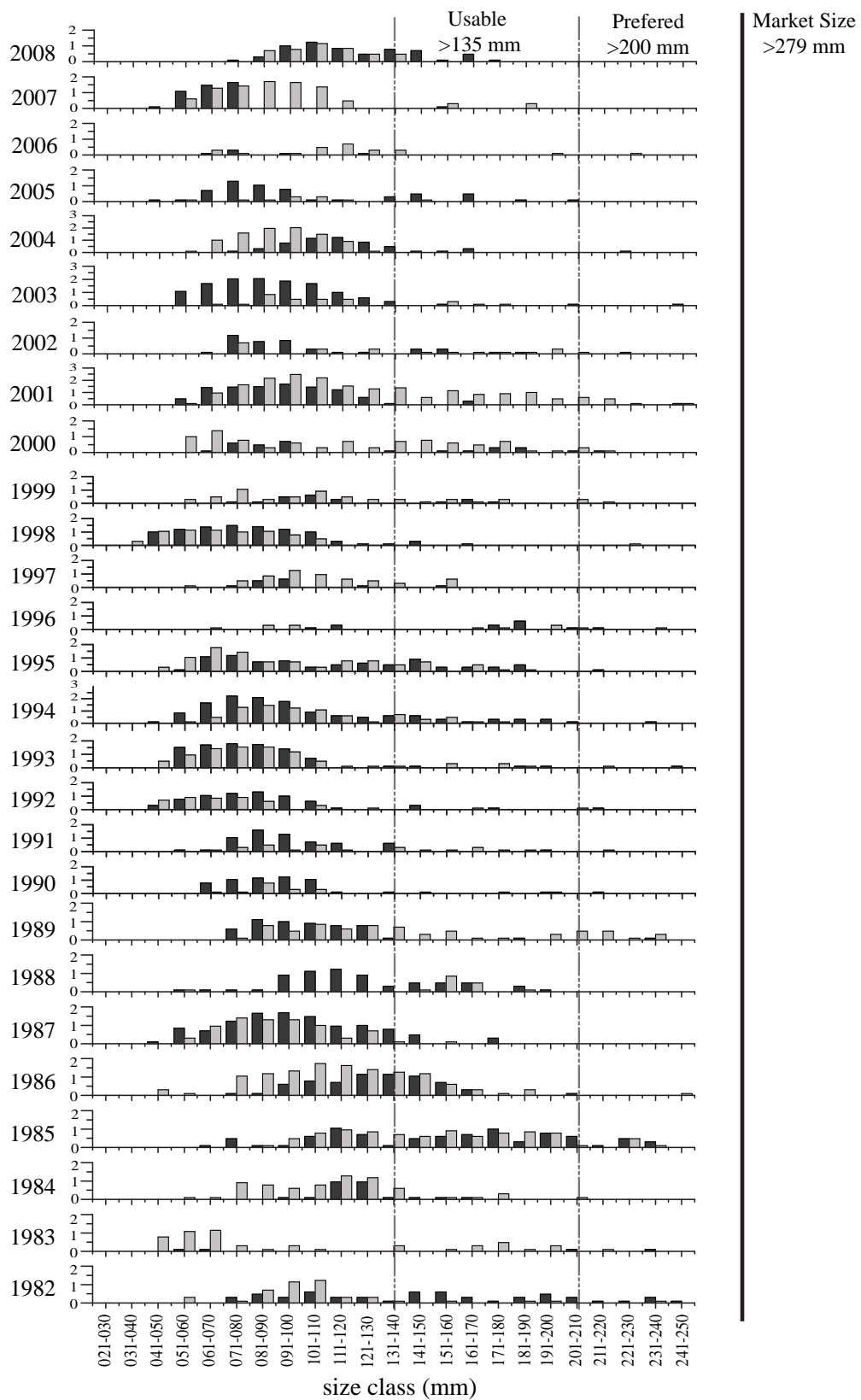


Figure 6-2  
 Size and log<sub>10</sub> abundance of English sole from outfall Station 01 (grey bars) and reference Station 06 (black bars) in community trawls, 1982-2008 by station and year.

sanddab (*Citharichthys sordidus*) for English sole one year (2006) when very few English sole were present, however, no fish species in our trawl samples consistently occurs of sufficient size or abundance to be useful for bioaccumulation monitoring.

#### 6.2.4 THE TRAWL SAMPLING RESULTS IN SIGNIFICANT AND UNNECESSARY MORTALITY TO DEMERSAL FISH AND EPIBENTHIC ORGANISMS INCLUDING LISTED SPECIES

The current mandated trawl sampling program includes one 10-minute trawl at depth at a speed of 2 knots at outfall station 01 and reference station 06 each year to assess assemblage composition. It is estimated that this activity alone resulted in approximately 3,000 fish being hauled to the surface for processing each year. While great care was taken to process quickly, salvage as many fish as possible and return them to the water alive, inevitably most did not survive the trawl. Additional trawls, 21 additional trawls in 2005 (Appendix F-1), were often conducted at each location (typically up to 30 minute duration at depth and at various speeds) in order to gather enough English sole to make the requisite tissue samples for bioaccumulation monitoring (Appendix F-1). The number of fish encountered by our net during the extra trawls conducted to collect fish tissue were not recorded, but were likely to be an order of magnitude more than those encountered (and recorded) during community trawls. A portion of the by-catch of the SWOO Regional Monitoring Program trawl sampling was longfin smelt (*Spirinchus thaleichthys*). Because the target of the trawl monitoring were demersal fish and epibenthic invertebrates, pelagic by-catch were not always identified or enumerated. Nonetheless, capture of longfin smelt was recorded in the monitoring data in a few instances (1982-85 and 1994-96). Smelt catch has not been reported in more recent years and fish encountered in extra trawls conducted to gather English sole for bioaccumulation have

not been recorded. The California Department of Fish and Wildlife listed the longfin smelt as threatened under the California Endangered Species Act (CESA) in 2009. CESA prohibits unpermitted taking of listed species and thus trawl sampling was curtailed by the SWOO Regional Monitoring Program in 2009 so that longfin smelt would not be harmed by the monitoring program.

#### 6.2.5 TRAWL SAMPLING DESTROYS BENTHIC HABITAT

Not only are fish populations affected by the trawling, bottom habitat is also disrupted by trawling activity. Otter boards and foot chains plough into the sea bed, disrupting sediment and benthic infaunal communities through direct physical disturbance as well as by creating a turbid cloud of fine sediment. The marine science literature is replete with accounts of the lasting damaging effects of trawling (e.g., NRC 2002) and the harmful effects of trawling on benthic habitat and the structure of benthic infaunal assemblages (e.g., Engel and Kvitek 1998).

#### 6.2.6 OTHER NEW SOURCES OF HIGH QUALITY ENVIRONMENTAL DATA ARE AVAILABLE

Since the inception of the SWOO Regional Monitoring Program other notable monitoring efforts have come on-line which provide biological data overlapping with our monitoring program.

The State Water Resources Control Board's Surface Water Ambient Monitoring Program (SWAMP) conducted a two-year, statewide biomonitoring screening survey of coastal fish 2009-2010, focusing on methylmercury and PCBs in fish tissue but evaluating several other contaminants as well. This study was exceptional in that it applied uniform methods to surveys of fish among 68 locations on the California coast, including ten locations in the SWOO monitoring program region and San Francisco Bay. The SWAMP Coast Survey is expected to be repeated

on a 5-year interval.

With regard to monitoring demersal fish and epibenthic species populations, there are studies that monitor in regions adjacent to the area currently sampled by the SWOO Regional Monitoring Program. The California Department of Fish and Wildlife, San Francisco Bay Study intensively samples in San Francisco Bay, including Central Bay, monthly using both otter trawl and midwater trawl. The National Oceanic and Atmospheric Administration (NOAA) U. S. West Coast Groundfish Bottom Trawl Survey samples regions adjacent to the SWOO Regional Monitoring Program region twice per year (summer and fall) to monitor the demersal fish and epibenthic species populations.

#### **6.2.7 GIVEN THE ABSENCE OF OUTFALL EFFECTS DEMONSTRATED BY THE DATA, THE TRAWL PROGRAM IS COSTLY AND BURDENSOME TO IMPLEMENT**

Trawl sampling requires significant additional ship time and cost to the SWOO Regional Monitoring Program because it is not compatible with the other monitoring elements of the program. Safe use of trawling gear requires that the decks of our monitoring vessel be cleared of all other bottom sampling gear. Because trawl sampling disturbs bottom habitat, it cannot be conducted until all sampling of bottom sediments, benthic infauna, and crab were completed. Setting and retrieving crab traps is, however, compatible with benthic sampling and those two monitoring activities are conducted concurrently. The SFPUC is not at this time advocating that crab monitoring be discontinued.

#### **6.3 SUMMARY AND CONCLUSIONS**

After reviewing 26 years of trawl sampling data we find no significant effect from the Southwest Ocean Outfall and no long-term degradation associated with the outfall. Bioaccumulation monitoring using tissue from fish caught by trawl and crab caught by traps in the reference and outfall areas have likewise not shown a significant outfall effect. Bioaccumulation

monitoring has shown that crab tissue has yielded a stronger, more persistent signal of environmental pollution in the region than tissue from demersal fish species. Trawl sampling is costly to biota and environment, endangering by-catch organisms, including the longfin smelt, which is listed as a threatened species. While no monitoring program exactly duplicates the efforts of the SWOO Regional Monitoring Program in the Gulf of the Farallones, several ongoing monitoring programs do exist which provide data of a quality and context that we are not able to match. In light of these discoveries we request that the trawl sampling component be removed from the SWOO Regional Monitoring Program.



## **SECTION 7**

### **BIOACCUMULATION AND PHYSICAL ANOMALIES**

# PHYSICAL ANOMALIES AND BIOACCUMULATION

## 7.1 INTRODUCTION

This section contains information on physical anomalies of individual organisms as well as organic and inorganic bioaccumulation data in the commercially important Dungeness crab *Cancer magister*. Information regarding the assessment of tissues of fish (primarily English sole *Pleuronectes vetulus* from 1997 through 2008) is in NRLMD 2010a.

Analysis of organic and inorganic compounds (trace metals) detected in organism tissues from reference and outfall areas can assist in assessing relative contamination of the study area. In addition, contamination of tissues in commercially important species may have public health implications. Comparing results from organisms collected at reference and outfall areas provides information on potential outfall effects, though caution should be used regarding those comparisons because of the mobility of these organisms.

This report references 2012 annual data and identifies trends from the 1997 through 2012 SWOO surveys (WQB 1998, 1999, 2000, 2001a, 2003a, 2003b, 2004; NRD 2006a, 2006b; NRLMD 2007, 2008, 2010a, 2010b, 2011, 2012). Comparisons of tissue pollutant levels with other agency data assist in the assessment of contaminant body burdens and human health concerns.

## 7.2 RESULTS AND DISCUSSION

### 7.2.1 PHYSICAL ANOMALIES

Dungeness crabs were examined for tumors and gross physical anomalies at the time of collection and taxonomic identification. Beginning in 2004, all Dungeness crab were more extensively assessed for physical anomalies (whether they were used for bioaccumulation analyses or not) than in previous years.

Adult Dungeness crab were examined for black necrotic disease (BND) or shell disease syndrome, a broad term describing the presence

of discrete areas of melanin production, necrosis, and exoskeleton erosion on crustaceans that may be a response to disease or injury and can lead to tissue death. The presence of BND is subject to molt cycles, and may be indicative of environmental pollution or naturally occurring degradative processes in sediment (Sawyer 1982, Comely and Ansell 1989, Noga et al. 2000, Vogán et al. 1999, 2002, 2008).

#### 7.2.1.1 Survey Year 2012

In 2012, 20% and 18% of the Dungeness crab collected in crab pots from the reference and outfall areas, respectively, had observed instances of BND (Table 7-1).

#### 7.2.1.2 Survey Years 1997 – 2003

Frequency of BND on crabs that were collected from the reference or outfall areas, and were used for bioaccumulation analyses, varied from 0 to 12% from 1997 to 2003.

#### 7.2.1.3 Survey Years 2004 – 2012

In 2004-2012, 7 – 36% of all the Dungeness crab collected in crab pots (and in community analysis and bioaccumulation trawls, when applicable) at both reference and outfall areas (regardless of whether they were used for analyses or not) had physical anomalies that were almost exclusively BND (Table 7-1).

The apparent increases from 1997-2003 to 2004-2012 in external physical anomalies are likely due to the more extensive external physical assessment of an increased number of organisms collected during part of the period 2004-2012 rather than an increase in the actual incidence of anomalies in fish and macro-invertebrates.

### 7.2.2 BIOACCUMULATION ORGANISMS

Prior to 2009, Dungeness crabs were collected by commercial crab-pot and trawl, primarily at reference station 06 and outfall station 01. Beginning in 2009, crab for bioaccumulation were collected using crab-pots, exclusively.

Within each survey year, organisms from



Table 7-1

*Abundance of Dungeness crab affected by external physical anomalies, collected from SWOO Reference and Outfall areas for community composition and bioaccumulation analyses, 1997 - 2012. The last survey year to include trawl-collected crab was 2008.*

Survey Year	REFERENCE						OUTFALL					
	No Anomaly Observed	Physical Anomaly Observed				Organisms Affected	No Anomaly Observed	Physical Anomaly Observed				Organisms Affected
		Deformity	Erosion	Tumor	BND*			Deformity	Erosion	Tumor	BND*	
2012	45	0	0	0	11	20%	62	0	0	0	14	18%
2011	20	0	0	0	2	9%	25	0	0	0	5	17%
2010	90	1	0	0	22	20%	70	0	0	0	35	33%
2009	37	0	0	0	12	24%	21	0	0	0	7	25%
2008	83	0	0	0	13	14%	59	0	0	0	5	8%
2007	24	0	0	0	17	10%	214	0	0	0	12	7%
2006	63	0	0	0	35	36%	91	0	0	0	19	17%
2005	67	0	0	1	25	28%	61	0	0	0	34	36%
2004	183	1	0	0	43	19%	180	0	0	0	42	19%
2003	119	0	0	0	9	7%	93	0	0	0	20	18%
2002	245	0	0	0	0	0%	76	0	0	0	1	1%
2001	131	0	0	0	3	2%	114	0	0	0	1	1%
2000	28	0	0	0	5	15%	26	0	0	0	4	13%
1999	ND	0	0	0	0	0%	ND	0	0	0	0	0%
1998	30	0	0	0	0	0%	29	0	0	0	1	2%
1997	ND	0	0	0	0	0%	ND	0	0	0	0	0%

\*BND = black necrotic disease  
ND = no data

both reference and outfall areas were of similar size and weight, with similar lipids content in both muscle and hepatopancreas tissues in organisms from both reference and outfall areas (Appendix G-1). Organic compounds and trace metals analyzed in tissues are listed in Appendices G-2 and G-4.

### 7.2.3 BIOACCUMULATIVE POLLUTANTS

#### 7.2.3.1 Sources

The source of bioaccumulated pollutants in tissue samples may be contaminated sediments from San Francisco Bay, since Dungeness crab rely heavily on estuarine environments during their juvenile stages and, as adults, are mobile predators that can range widely along latitudinal and onshore-offshore gradients (Hankin and Warner 2001; Pauley et al 1989, PSMFC 1996).

These SWOO data address the extent of bioaccumulated contaminants in these organisms; attempts to assess contaminant origins or environmental and public-health impacts of those pollutant body burdens should be made with caution.

#### 7.2.3.2 Organic Pollutants

Organic analyses of the tissues in this study confirm the lipid affinity of pesticides and PCBs, as evidenced by the increased number and concentrations of organic compounds detected in hepatopancreas tissue compared to muscle tissue (Table 7-2).

##### 7.2.3.2.1 Survey Year 2012

In 2012, one organochlorine pesticide (DDE), two PAHs, and eight PCBs were found above detection limits at both reference and outfall areas; concentrations of these compounds were generally low, at levels comparable to previous years (Table 7-2).

##### 7.2.3.2.2 Survey Years 1997 – 2012

Of the three DDTs, 18 PAHs and 53 PCB congeners assessed in tissues during the 16-year study period, one DDT (DDE), one PAH and 11 PCB congeners were detected in crab tissues in at least half of the survey years. There is only one instance of statistically significantly-

elevated level of organic compounds (total PCBs) in muscle tissue of crab from the outfall area compared to the reference area in 2003 (Table 7-2). In all other years, levels of organic compounds in tissues were not statistically different between reference and outfall areas, or were statistically significantly elevated in tissues of organisms from the reference area compared to the outfall area (PAHs in muscle in 2001 and in hepatopancreas in 2003, and DDTs in muscle in 2004).

7.2.3.2.2.1 *Organochlorine pesticides (DDTs)*

Low concentrations of the DDT break-down compound 4,4'DDE were frequently detected in

crab muscle tissue in all survey years (Appendix G-3), with 4,4'DDT and 4,4'DDD not detected in crab tissues since 2007.

7.2.3.2.2.2 *Polycyclic Aromatic Hydrocarbons (PAHs)*

Throughout the SWOO study in both reference and outfall areas, PAHs were detected in both tissue types at varying concentrations (Table 7-2).

7.2.3.2.2.3 *Polychlorinated biphenyls (PCBs)*

PCBs are environmental contaminants of concern due to their general resistance to metabolism and their tendency to biomagnify.

Table 7-2

*Mean concentrations (ppb, wet weight) of organic pollutants detected in tissues of Dungeness crab collected from SWOO Reference and Outfall areas 1997 - 2012, and other available study data (actual concentrations may be less than value indicated when one-or-more replicates were below detection limits).*

Crab Muscle																	
		Reference area															
CFCP		1997	1998	1999	2000	2001	2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012
6	Σ DDTs	-	13	5	15	-	8	11	<b>10</b>	11	17	7	10	6	5	9	3
	Σ PAHs	22	-	-	95	<b>324</b>	494	57	20	33	4	99	11	16	19	-	2
9	Σ PCBs	-	-	-	36	-	-	2	2	4	2	2	-	-	-	-	-
		Outfall area															
CFCP		1997	1998	1999	2000	2001	2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012
6	Σ DDTs	-	21	5	13	-	6	10	8	13	12	16	15	4	4	4	5
	Σ PAHs	25	-	-	69	46	281	9	18	31	10	75	5	18	19	4	4
9	Σ PCBs	-	-	-	3	-	-	<b>6</b>	-	4	2	-	2	-	-	-	-

Crab Hepatopancreas																	
		Reference area															
CFCP		1997	1998	1999	2000	2001	2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012
	Σ DDTs	231	473	522	270	136	155	121	86	57	46	118	67	101	4	71	32
	Σ PAHs	18	-	-	6	22	111	<b>168</b>	7	33	4	262	10	10	66	2	-
	Σ PCBs	56	162	366	127	198	194	95	61	19	32	101	71	70	28	41	28
		Outfall area															
CFCP		1997	1998	1999	2000	2001	2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012
	Σ DDTs	307	460	300	171	132	195	135	85	61	57	143	59	65	4	61	43
	Σ PAHs	-	48	-	9	14	86	98	39	24	5	272	8	11	52	10	-
	Σ PCBs	164	153	326	130	159	167	87	50	20	22	154	47	65	29	37	43

- = below detection limits

**Blue font** = statistically significantly higher than corresponding tissue at other area; one-tailed T-test, unequal variance, α = 0.05

CFCP - California Fish Contamination Program (RWQCB 2005, CEDEN 2010)

Organisms at the top of the food web (including humans who consume fish and shellfish) are, therefore, vulnerable to the effects of PCB exposure (U.S. EPA 1997, 1999), which can lead to toxic effects such as developmental abnormalities and growth suppression, disruption of the endocrine system, impaired immune function, and cancers (ATSDR 2000, 2010).

Levels of total PCBs in crab tissues were generally low (near detection limits, Appendix G-2), except in hepatopancreas of crab from both the reference and outfall areas (Table 7-2), as were concentrations of individual PCB-congeners (Appendix G-3).

PCBs of high concern were detected primarily in hepatopancreas of organisms from both the outfall and reference areas throughout the study years. The PCB congeners of most concern are PCB 77, PCB 126, and PCB 169, which closely mimic the potency of dioxin, one of the most toxic substances identified (SFEI 1999a, U.S. EPA 1996). PCB 126, which has the potency one-tenth that of dioxin, was detected in hepatopancreas of crab from both reference and outfall areas at levels generally near detection limits from 1999 to 2004, and not at all since 2004 (Appendix G-3).

#### 7.2.3.2.2.4 *Trends*

None of the regressions involving sediment and tissue concentrations of these organic compounds (total DDTs, total PAHs, total PCBs) are significant (regression analysis,  $\alpha = 0.05$ ). There appears to be a trend of decreasing PCB-levels in hepatopancreas of crabs from both reference and outfall areas over time (regression analysis,  $\alpha = 0.05$ ), but those levels are generally detected near-or-below detection limits, and conclusions about them should be made cautiously.

#### 7.2.3.2.2.5 *Other studies*

Comparisons of SWOO data with other local studies is limited due to differences in species composition, numbers, or organisms assessed, and the types of tissues and contaminants analyzed in each study. Body burdens of

organic pollutants in Dungeness crab from the San Mateo coast assessed in the California Fish Contamination Program (CFCP) (RWQCB 2005, CEDEN 2010) are generally similar to those levels found in the SWOO study data (Table 7-2).

#### 7.2.3.3 Inorganic Analyses – Trace Metals

##### 7.2.3.3.1 Survey Year 2012

In 2012, there were no statistically significant differences in trace metal concentrations in tissues of organisms collected from the reference and outfall areas, and individual analyte concentrations are similar to previous years (Table 7-3).

##### 7.2.3.3.2 Survey Years 1997 – 2012

For survey years 1997 – 2000 wet-weight data for trace metals are unavailable, making comparisons with data from subsequent years inappropriate; those dry-weight data and T-test results are in Appendix G-5.

Concentrations of trace metals analyzed 2001 – 2012 were similar, in all years for muscle and hepatopancreas of organisms from both reference and outfall areas (Table 7-3). Comparisons of metal concentrations indicated five (out of 312 comparisons) statistically significantly elevated trace-metal levels (iron, manganese, silver, nickel, and lead) in tissues of organisms from the outfall area compared to the reference area within sample years (Table 7-3), and none since 2004. Silver and zinc in muscle, and chromium in hepatopancreas, were detected at statistically higher levels in tissues of organisms from the reference area (compared to the outfall area) in 2003.

##### 7.2.3.3.2.1 *Trends*

Metals bioaccumulation in tissues of crab from both reference and outfall areas were similar for the entire study period, with no statistically significant trends and little pattern apparent in those concentrations (regression analysis,  $\alpha = 0.05$ ). The exception is iron, detected in decreasing concentrations in



#### 7.2.3.3.2.2 *Other studies*

Body burdens of trace metals in Dungeness crab collected from the San Mateo County coast assessed in the CFCP (RWQCB 2005, CEDEN 2010) are generally similar to those levels found in both the SWOO-study reference and outfall areas (Table 7-3).

### 7.3 **SUMMARY AND CONCLUSIONS**

Overall organism condition (based on frequency of observed anomalies) and concentrations of pollutants analyzed in tissues of organisms collected from the SWOO study area were similar between reference and outfall regions in both muscle and hepatopancreas. This similarity indicates that the SWOO discharge does not appear to affect organism body burdens of these pollutants. The bioaccumulation data demonstrate that concentrations of organic pollutants and trace metals are found in varying levels and tend to accumulate in the fatty hepatopancreas tissue, so this kind of monitoring may be important in the management of the fishery as well as for public health issues. Historically, crab hepatopancreas (crab ‘butter’) has been considered a delicacy and continues to be eaten by portions of the population; public awareness and education may be important to inform people that those tissues may not be suitable for consumption. Consequences of increased body burdens of organic pollutants (e.g. resulting in potential immunological and reproductive impairment) or trace metals (e.g. resulting in neurological damage) on the health of the organism populations are unknown at this time. Variable contaminant concentrations in tissues of organisms from reference and outfall regions, and between sampling years, may be attributable to species selection; these organisms are mobile and therefore not necessarily representative long-term residents of the locations from which they are collected. These data likely reflect the general concentration of bioaccumulated contaminants from the entire SWOO study area and this region of the California coast.

Of those pollutants detected at elevated

levels, most were detected in hepatopancreas tissues. These data provide further evidence of the importance in educating the population against eating whole organisms that include organ tissues, and recommending limiting consumption to muscle tissues.



**SECTION 8**  
**REFERENCES**

## REFERENCES CITED

- Agency for Toxic Substances and Disease Registry 2000. Toxicological profile for polycyclic aromatic hydrocarbons (PCBs). Atlanta, GA: U.S. Department of Health and Human Service, Public Health Service.
- 2010. Public health implications of exposure to polychlorinated biphenyls (PCBs), at <http://www.atsdr.cdc.gov/DT/pcb007.html> (1/30/2010).
- Anderson, J.W. and R.W. Gossett 1988. PAH contamination in sediments from coastal waters. *Coastal Water Research News* 1(3): 1.
- Anderson, J.W., R.G. Riley, S.L. Kiesser, B.L. Thomas, and G.W. Fallingham 1983. Natural weathering of oil in marine sediments: tissue contamination and growth of the littleneck clam, *Protothaca staminea*. *Canadian Journal of Fisheries and Aquatic Science* 40(Suppl. 2): 70-77.
- ATSDR see Agency for Toxic Substances and Disease Registry
- Barkley, R.A. 1972. Selectivity of towed-net samplers. *Fishery Bulletin* 70: 799-819.
- Barnard, P.L., Hanes, D.M., Rubin, D. M., and Kvitck, R. G., 2006. Giant sand waves at the mouth of San Francisco Bay: *Eos* (American Geophysical Union Transactions), v. 87, no. 29, p. 285, 289.
- Barnard, P.L., Schoellamer, D.H., Jaffe, B.E., McKee, L.J., 2013. Sediment transport in the San Francisco Bay Coastal System: An overview. *Marine Geology* 345 (2013) pp. 3-17.
- Barnard, P.L., Hansen, J.E., Erikson, L.H., 2012. Synthesis study of an erosion hot spot, Ocean Beach, California. *Journal of Coastal Research* 28(4), pp. 903-922
- Bernstein, B.B. and J. Zalinski 1983. An optimum sampling design and power tests for environmental biologists. *Journal of Environmental Management* 16:35-43.
- Bernstein, B.B. and R. Smith 1986. Community approaches to monitoring. *IEEE Oceans '86* Conference Proceedings pp. 934-939.
- Bilyard, G.R. 1987. The value of benthic infauna in marine pollution monitoring studies. *Marine Pollution Bulletin* 18 (10): 581-585.
- Boesch, D.F. and R. Rosenberg 1981. Response to stress in marine benthic communities. Pp. 179-299 in: G.W. Barrett and R. Rosenberg, editors. *Stress effects on natural ecosystems*. John Wiley, London.
- Bolin, R.L. and D.P. Abbott 1963. Studies on the marine climate and phytoplankton of the central coastal area of California, 1954-1960. *California Cooperative Oceanic Fisheries Investigations Reports Volume IX:23-45*.
- Bray, J. and J. Curtis 1957. An ordination of upland forest communities of southern Wisconsin. *Ecological Monographs* 27:325-349.
- Brown and Caldwell 1971a. A pre-design report on marine waste disposal. Vol. I. Oceanographic and ecological base data acquisition and evaluation of alternative locations. Prepared for the City and County of San Francisco.
- Brown and Caldwell 1971b. A pre-design report on marine waste disposal. Vol. II. Data supplement. Oceanographic and ecological base data acquisition and evaluation of alternative locations. Prepared for the City and County of San Francisco.
- Bryan, G.W. 1971. The effects of heavy metals (other than mercury) on marine and estuarine organisms. *Proceedings of the Royal Society of London* 177:389-410.
- Bureau of Water Pollution Control (BWPC) 1984. Ocean outfall monitoring program 1982-83 annual report. City and County of San Francisco, Department of Public Works. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 1985. Ocean outfall monitoring program 1983-84 annual report. City and County of San Francisco, Department of Public Works. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 1988. Ocean outfall monitoring program

- 1987 annual report. City and County of San Francisco, Department of Public Works. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 1989. Ocean outfall monitoring program 1988 annual report. City and County of San Francisco, Department of Public Works. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 1990. Ocean outfall monitoring program 1989 annual report. City and County of San Francisco, Department of Public Works. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 1992a. Ocean outfall monitoring program 1990 annual report. City and County of San Francisco, Department of Public Works. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 1992b. Ocean outfall monitoring program 1991 annual report. City and County of San Francisco, Department of Public Works. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 1993. Ocean outfall monitoring program 1992 annual report. City and County of San Francisco, Department of Public Works. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 1994. Ocean outfall monitoring program 1993 annual report. City and County of San Francisco, Department of Public Works. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 1995. Ocean outfall monitoring program 1994 annual report. City and County of San Francisco, Department of Public Works. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- BWPC see Bureau of Water Pollution Control
- California Department of Public Health 2006. Draft guidance for saltwater beaches. [http://www.cdph.ca.gov/HealthInfo/ environhealth/water/Documents/Beaches/ DraftGuidanceforSaltWaterBeaches.pdf](http://www.cdph.ca.gov/HealthInfo/environhealth/water/Documents/Beaches/DraftGuidanceforSaltWaterBeaches.pdf) [accessed 25 January 2010]
- Casteel, M.J. 2005. A review of the state of the science for detecting *Toxoplasma gondii*, *Sarcocystis neurona*, and morbilliviruses – potential pathogens for marine mammals – in wastewater and stormwater. City and County of San Francisco, Public Utilities Commission, Water Quality Bureau. Submitted to the U.S. EPA Region 9 and the San Francisco Bay Regional Water Quality Control Board.
- Carlson, P.R. and D.S. McCulloch 1974. Aerial observations of suspended-sediment plumes in San Francisco Bay and the adjacent Pacific Ocean. U.S. Geological Survey, Journal of Research 2(5):519-526.
- CDPH see California Department of Public Health
- CH<sub>2</sub>M Hill 1983. Southwest Ocean Outfall, City and County of San Francisco Clean Water Program. Predischarge Oceanographic Study, Phase 2 Final Report to PBQ and D, Inc., September 1983
- Chapman, P.M., R.N. Dexter, S.F. Cross, and D.G. Mitchell 1986. A field trial of the sediment quality triad in San Francisco Bay. NOAA Technical Memorandum NOS OMA 25; 134 pp.
- Chow, S. and P. Rodgers. 2005. Applet for drawing 3 set area-proportional Venn diagrams. <http://www.cs.kent.ac.uk/people/staff/pjr/EulerVennCircles/EulerVennApplet.html> [accessed 25 January 2010]
- Clarke KR and Warwick RM. 1994. Similarity-based testing for community pattern: the 2-way layout with no replication. Marine Biology 118:167-176.
- Clarke, K. R. and R. N. Gorley. 2006. Primer v6: User Manual / Tutorial. Primer-E Ltd, Plymouth Marine Laboratory. 1-190.
- Comely, C.A. and A.D. Ansell 1989. The occurrence of black necrotic disease in crab species from the west of Scotland. Ophelia 30(2): 95-112.
- Conomos, J.T. 1979. Properties and circulation of San Francisco Bay waters. Pp. 47-84 in: Conomos, J.T., Ed. San Francisco Bay: the



- urbanized estuary. American Association for the Advancement of Science, Pacific Division, San Francisco, CA; 493 pp.
- Cooper, A. 1973. Structure of the Continental Shelf West of San Francisco, California. United States Department of the Interior Geological Survey, Open File Report. 65 pp.
- Dean, W.E. and Gardener, J.V. 1995. Geochemistry of surface sediments in the Gulf of the Farallones. United States Geological Survey Open-File Report 95-527: 8 pp.
- De Lappe, B.W., W.R. Sistek, and R.W. Risebrough, 1980. Predischarge Studies: San Francisco Southwest Ocean Outfall Project: The Distribution of Higher-Molecular Weight Hydrocarbons in the Coastal Environment. The Bodega Marine Laboratory. A report to CH2M Hill, 2200 Powell Street, Emeryville, California 94608.
- DHS see California Department of Health Services
- Diener, D.R. and S.C. Fuller 1995. Infaunal patterns in the vicinity of a small coastal wastewater outfall and the lack of infaunal community response to secondary treatment. *Bulletin of the Southern California Academy of Sciences* 94(1): 5-20.
- Dillon, W.R., and M. Goldstein. 1984. *Multivariate Analysis Methods and Applications*. John Wiley & Sons, New York. 587 pp.
- Dorsey, J.H., C.A. Phillips, A. Dalkey, J.D. Roney, and G.B. Deets 1995. Changes in assemblages of infaunal organisms around wastewater outfalls in Santa Monica Bay, California. *Bulletin of the Southern California Academy of Sciences* 94(1): 46-64.
- Di Lorenzo, E., N. Schneider, K.M. Cobb, P.J.S. Franks, K. Chhak, A.J. Miller, J.C. McWilliams, S.J. Bograd, H. Arango, E. Curchitser, T.M. Powell, and P. Riviere 2008. North Pacific Gyre Oscillation links ocean climate and ecosystem change. *Geophysical Research Letters*, Volume 35, L08607; 6 pp.
- Eaton, A.D., Clesceri, L.S., and A.E. Greenberg 2005. *Standard methods for the examination of water and wastewater*, 21<sup>st</sup> edition. Washington, D.C.
- Engel, J. and Kvitek, R. (1998), Effects of Otter Trawling on a Benthic Community in Monterey Bay National Marine Sanctuary. *Conservation Biology*, 12: 1204 -1214. doi: 10.1046/j.1523-1739.1998.0120061204.x
- Folk, R.L. and W.C. Ward 1957. Brazos River Bar: a study in the significance of grain size parameters. *Journal of Sedimentary Petrology* 27 (1): 3-26.
- Furse, M.T., D. Moss, J.F. Wright, and P.D. Armitage 1984. The influence of seasonal and taxonomic factors on the ordination and classification of running water sites in Great Britain and on the prediction of their macroinvertebrate communities. *Freshwater Biology* 14: 257-280.
- Gobas, F.A.P.C., J.R. McCorquodale, and G.D. Haffner. 1993. Intestinal absorption and biomagnification of organochlorines. *Environmental Toxicology and Chemistry* 14(5): 801-807.
- Grassle, J.F., R. Elmgren and J.P. Grassle. 1981. Response of benthic communities in MERL experimental ecosystems to low level chronic additions of No. 2 fuel oil. *Marine Environmental Research* 4: 279-297.
- Hankin, D. and R.W. Warner 2001. Dungeness crab, pp. 107-111 in: W.S. Leet, et al. Eds. *California's Living Marine Resources: A Status Report*. The Resources Agency, California Department of Fish and Game and the University of California, Sea Grant Program; 592 pp. [also available at: <http://www.dfg.ca.gov/mrd/status/index.html>]
- Hurlbert, S.H. 1984. Pseudoreplication and the design of ecological field experiments. *Ecological Monographs* 54: 187-211.
- Hutzinger, O. 1982. *The handbook of environmental chemistry*. Volume 3, Part B. Anthropogenic compounds. Springer-Verlag. New York. 210 pp.

- IEP see Interagency Ecological Program
- Keller, E.A. 1976. Environmental Geology, University of California, Santa Barbara.
- Kellogg, M.G., L. Riege, A. Navarret, and R.W. Smith 1998. San Francisco ocean monitoring program: analysis of long term data, pp. 1606-1618 in: Magoon, O.T., *et al.*, editors. California and the world ocean '97: taking a look at California's ocean resources: an agenda for the future. Conference Proceedings, Vol. 2, American Society of Civil Engineers, Reston, VA; 1756 pp.
- Khan, S.V. 1980. Role of humic substances in predicting fate and transport of pollutants in the environment. Pp. 215-230 in: R. Hague, ed. Dynamics, exposure, and hazard assessment of toxic chemicals. Ann Arbor, Science Ann Arbor, MI.
- Knox, G.A. 1977. The role of polychaetes in benthic soft-bottom communities. Pp. 547-605 in: D. J. Reish, and K. Fauchald, editors. Essays on polychaetous annelids in memory of Dr. Olga Hartman. Allan Hancock Foundation, University of Southern California, Los Angeles.
- LACSD see Los Angeles County Sanitation Districts
- Lassuy, D.F. 1989. Species profiles: life histories and environmental requirements of coastal fishes and invertebrates (Pacific Northwest) – English sole. U.S. Fish Wildl. Serv. Biol. Rep. 82(11.101). U.S. Army Corps of Engineers, TR EL-82-4. 17pp.
- Law, R.J., and J.L. Biscaya, 1994. Polycyclic Aromatic Hydrocarbons (PAH) – Problems and Progress in Sampling, Analysis and Interpretation. Marine Bulletin, Vol. 29, NOS 4-5, pp 235 – 241.
- Lenat, D.R. and M.T. Barbour 1994. Using benthic macroinvertebrate community structure for rapid, cost-effective, water quality monitoring: rapid bioassessment. Pp. 187-215 in: S.L. Loeb and A. Spacie, editors. Biological monitoring of aquatic systems. Lewis Publishers, Boca Raton, Florida.
- Levensen, H. and D. Barnard 1988. Wastes in Marine Environments. U.S. Congress, Office of Technology Assessment, Washington D.C.; 312 pp.
- Levinton, J. 1972. Stability and trophic structure in deposit-feeding and suspension-feeding communities. American Naturalist 106(950): 472-486.
- Long, E.R., D.D. MacDonald, S.L. Smith, and F.D. Calder. 1995. Incidence of adverse biological effects within ranges of chemical concentrations in marine and estuarine sediments. Env. Mgmt. 19:81 – 97.
- Los Angeles County Sanitation Districts (LACSD) 1981. Ocean monitoring and research: Annual report 1980-1981. 384 pp. + Appendices.
- Malins, D.C. 1982. Alterations in the cellular and subcellular structure of marine teleosts and invertebrates exposed to petroleum in the laboratory and field: a critical review. Canadian Journal of Fisheries and Aquatic Sciences 39:877-889.
- B.B. McCain, D.W. Brown, A.K. Sparks, and H.O. Hodgins 1980. Chemical contaminants and biological abnormalities in Central and Southern Puget Sound. NOAA Technical Memorandum OMPA-2. 295 pp.
- M.M. Krahn, D.W. Brown, L.D. Rhodes, M.S. Myers, B.B. McCain, and S.-L. Chan. 1985a. Toxic chemicals in marine sediment and biota from Mukilteo, Washington: relationships with hepatic neoplasms and other hepatic lesions in English sole (*Parophrys vetulus*). Journal of the National Cancer Institute 74(2):487-494.
- M.M. Krahn, M.S. Meyers, L.D. Rhodes, D.W. Brown, C.A. Krone, B.B. McCain, and S.-L. Chan. 1985b. Toxic chemicals in sediments and biota from a creosote-polluted harbor: relationships with hepatic neoplasms and other hepatic lesions in English sole (*Parophrys vetulus*). Carcinogenesis 6(10):146-469.
- U. Varanasi, D.W. Brown, M.M. Krahn, and S.-L. Chan 1986. Biological transport of contaminants in marine environments: bioavailability and biotransformations.

- Rapports et Proces-verbaux des Reunions, International Council for the Exploration of the Seas 186: 442-448.
- McCave, I.N. 1981. Location of coastal accumulations of fine sediments around the southern North Sea. Rapports et Proces-verbaux des Reunions, International Council for the Exploration of the Seas 181: 15-27.
- McDermott, D.J., G.V. Alexander, D.R. Young and A.J. Mearns. 1976. Metal contamination of flatfish around a large submarine outfall. *Journal of the Water Pollution Control Federation* 48:1915-1918.
- Melwani, A.R, B.K. Greenfield, L. Targgart, and M. Kellogg 2009. Patterns in mercury and trace organic contamination of sportfish and sediments in San Francisco Bay compared to the offshore coast. San Francisco Estuary Institute, Regional Monitoring Program for Water Quality, Contribution 597; 31 pp.
- Mudroch, A. and J. M. Azcue, 1995. *Manual of Aquatic Sediment Sampling*. Lewis Publishers, Boca Raton 219p.
- National Oceanic and Atmospheric Administration (NOAA) 1992. Monterey Bay National Marine Sanctuary. Final Environmental Impact Statement/ Management Plan. Volume II: Appendices. U.S. Department of Commerce: National Oceanic and Atmospheric Administration, Sanctuaries and Reserves Division.
- 1999a "El Niño theme page." [<http://www.pmel.noaa.gov/toga-tao/el-nino/nino-home.html>]
- 2005. <ftp://orpheus.pfeg.noaa.gov/outgoing/upwell/monthly/upindex.mon> (downloaded 7/22/2005).
- 2014. [http://www.pfeg.noaa.gov/products/PFEL/modeled/indices/upwelling/NA/upwell\\_menu\\_NA.html](http://www.pfeg.noaa.gov/products/PFEL/modeled/indices/upwelling/NA/upwell_menu_NA.html) (downloaded 2/6/2014).
- National Research Council (NRC) (2002) *Effects of Trawling and Dredging on seafloor Habitat*. Washington, DC: The National Academies Press, 2002
- Natural Resources Division (NRD) 2006a. Southwest ocean outfall regional monitoring program eight-year summary report 1997-2004. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA Region 9 and S.F. Bay Regional Water Quality Control Board.
- 2006b. Southwest ocean outfall regional monitoring program 2005 data report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA Region 9 and S.F. Bay Regional Water Quality Control Board.
- Natural Resources and Lands Management Division (NRLMD) 2007. Southwest ocean outfall regional monitoring program 2006 data report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA Region 9 and S.F. Bay Regional Water Quality Control Board.
- 2008. Southwest ocean outfall regional monitoring program 2007 data report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA Region 9 and S.F. Bay Regional Water Quality Control Board.
- 2010a. Southwest ocean outfall regional monitoring program twelve-year summary report 1997-2008. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA Region 9 and S.F. Bay Regional Water Quality Control Board.
- 2010b. Southwest ocean outfall regional monitoring program 2009 data report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA Region 9 and S.F. Bay Regional Water Quality Control Board.
- 2011. Southwest ocean outfall regional monitoring program 2010 data report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA Region 9 and S.F. Bay Regional Water Quality Control Board.
- 2012. Southwest ocean outfall regional monitoring program 2011 data report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA Region 9 and S.F. Bay Regional Water Quality Control Board.

- Control Board.
- Niemi, C.A and K.I. Warheit 1989. Variation in species diversity and physical environment associated with primary treated sewage effluent. Pp. 635-640 in: Oceans '89: an international conference addressing methods for understanding the global ocean. Conference Proceedings, Vol. 2, Institute of Electrical and Electronics Engineers.
- NOAA see National Oceanic and Atmospheric Administration
- Nobel, M. and G. Gelfenbaum 1988. A proposal for a pilot study of sediment transport and currents in the Gulf of the Farallones region. U.S. Geological Survey, Open File Report.
- Noga, E.J., R. Smolowitz and L.H. Khoo 2000. Pathology of shell disease in the blue crab, *Callinectes sapidus* Rathbun, (Decapoda: Portunidae). Journal of Fish Diseases 23: 389-399.
- North Pacific Gyre Oscillation (NPGO). At <http://eros.eas.gatech.edu/npgo/> (1/30/2010)
- NRD see Natural Resources Division
- NRLMD see Natural Resources and Lands Management Division
- Null, J. 2001. Golden Gate Weather Services. San Francisco Rainfall Season in Review. [<http://ggweather.com/sf/climate.html>]
- Null, J. 1995. Climate of San Francisco, 3<sup>rd</sup> Rev. NOAA Technical Memorandum NWS WR-126. [http://www.wrh.noaa.gov/mtr/sfd\\_sjc\\_climate/sfd/SFD\\_CLIMATE3.php](http://www.wrh.noaa.gov/mtr/sfd_sjc_climate/sfd/SFD_CLIMATE3.php) [accessed 24 Dec 2009]
- NWS see National Weather Service
- Oliver, J.S., P.N. Slattery, L.W. Hulberg, and J.W. Nybakken. 1980. Relationships between wave disturbance and zonation of benthic invertebrate communities along a subtidal high-energy beach in Monterey Bay, California. Fishery Bulletin 78(2):437-454.
- Overton, W.S., D. White, and D.L. Stevens, Jr. 1990. Design report for EMAP environmental monitoring and assessment program. U.S. Environmental Protection Agency, U.S. EPA/600/3-91/053; 41 pp.
- Pacific States Marine Fisheries Commission (PSMFC) 1996. Habitat Program, English Sole factsheet. URL: <[http://www.psmfc.org/habitat/edu\\_ole\\_fact.html](http://www.psmfc.org/habitat/edu_ole_fact.html)>. (3/31/2010)
- 1996. Habitat Program, Dungeness Crab factsheet. URL: <[http://www.psmfc.org/habitat/edu\\_crab\\_fact.html](http://www.psmfc.org/habitat/edu_crab_fact.html)>. (3/31/2010).
- Parker, W.R. and K. Lee 1981. The behaviour of the fine sediment relevant to the dispersal of pollutants. Rapports et Proces-verbaux des Reunions, International Council for the Exploration of the Seas 181: 28-34.
- Parsons, T.R. and M. Takahashi 1984. Benthic communities. In: Biological oceanographic processes, 3rd edition. Pergamon Press. Elmsford, New York.
- Pauley, G.G., D.A. Armstrong, R. VanCitter, and G.L. Thomas 1989. Species profiles: life histories and environmental requirements of coastal fishes and invertebrates (Pacific Southwest) – Dungeness crab. U.S. Fish Wildl. Serv. Biol. Rep. 82(11.121). U.S. Army Corps of Engineers, TR EL-82-4. 20pp.
- Pavlova, Y.V. 1966. Seasonal variations of the California current. Oceanology 6:806-814.
- Pearson, D.E., S.L. Owen, and D. Thomas 2001. English sole, pp. 384-385 in: W.S. Leet, et al. Eds. California's Living Marine Resources: A Status Report. The Resources Agency, California Department of Fish and Game and the University of California, Sea Grant Program; 592 pp. [also available at: <http://www.dfg.ca.gov/mrd/status/index.html> ]
- Pearson, T.H. and R. Rosenberg 1978. Macrobenthic succession in relation to organic enrichment and pollution of the marine environment. Oceanography and Marine Biology Annual Review 16: 229-311.
- Plumb, R.J., Jr. 1981. Procedures for handling and chemical analysis of sediment and water samples. U.S. Environmental Protection Agency Technical Report U.S. EPA/CE-81-1. 403 pp.
- PSMFC see Pacific States Marine Fisheries Commission
- Reed, M.K.J., D. French, M.L. Spaulding, T. Isaji, and J. Rosen 1986. A model system

- for estimating fate and effects of pollutants in marine ecosystems: applications and sensitivity analyses. *Rapports et Proces-verbaux des Reunions, International Council for the Exploration of the Seas* 186: 80-103.
- Reish, D.J. 1980. Effect of domestic wastes on the benthic marine communities of southern California. *Helgolander Meeresunters* 33: 377-383.
- Reish, D.J. 1983. Survey of the marine benthic infauna collected from the United States Radioactive waste disposal sites off the Farallon Islands, California. U.S. EPA 520/1-83-006. Office of Radiation Programs, Washington, D.C.; 54 pp.
- Rhodes, L., E. Casillas, B. McKnight, W. Gronlund, M. Myerts, O.P. Olson, and B. McCain. 1985. Interactive effects of cadmium, polychlorinated biphenyls, and fuel oil on experimentally exposed English sole (*Parophrys vetulus*). *Canadian Journal of Fisheries and Aquatic Sciences* 43:1870-1880.
- Rice, E.W., R.B. Baird, A.D. Eaton, and L.S. Clesceri. 2012. *Standard Methods for the Examination of water and Wastewater*, 22<sup>nd</sup> edition. Washington, D.C.
- Rosenberg, D.M., H.V. Danks, and D.M. Lehmkuhl 1986. Importance of insects in environmental impact assessment. *Environmental Management* 10: 773-783.
- RWQCB see San Francisco Bay Regional Water Quality Control Board
- Rygg, B. 1985. Distribution of species along pollution-induced diversity gradients in benthic communities in Norwegian Fjords. *Marine Pollution Bulletin* 16(12): 469-474.
- San Francisco Bay Regional Water Quality Control Board (RWQCB) 2005. Chemical concentrations of fish tissues from selected reservoirs and coastal areas in the San Francisco Bay region. Final Report. Surface Waters Ambient Monitoring Program.
- San Francisco Estuary Institute (SFEI) 1996. San Francisco Estuary regional monitoring program for trace substances: 1994 Annual Report. San Francisco Estuary Institute, Richmond, CA; 338 pp.
- 1999a. Contaminant Concentrations in Fish from San Francisco Bay 1997. A technical report of the San Francisco Estuary Regional Monitoring Program for Trace Substances. RMP Contribution #35.
- 2000. The Pulse of the Estuary. Monitoring and Managing Contamination in the San Francisco Estuary 1993-99. [<http://www.sfei.org/rmp/1999/1999Pulse.pdf>]
- 2002. Pulse of the Estuary: Monitoring and managing contamination in the San Francisco Estuary: 2000 update. San Francisco Estuary Institute, Richmond, CA. 28 pp.
- 2004. The Pulse of the Estuary: Monitoring and Managing Water Quality in the San Francisco Estuary. SFEI Contribution 78. San Francisco Estuary Institute, Oakland, CA.
- 2013. The Pulse of the Bay: Contaminants of Emerging Concern. SFEI Contribution 701. San Francisco Estuary Institute, Richmond, CA.
- San Francisco Estuary Project (SFEP) 1993. San Francisco Bay-Delta Estuary. Information Brochure, San Francisco Estuary Project, P.O. Box 2050, Oakland, CA 94604; 4 pp.
- Sawyer, T.K. 1982. Distribution and seasonal incidence of "black gill" in the rock crab, *Cancer irroratus*. In, *Ecological Stress and the New York Bight: Science and Management*. G.F. Mayers (ed.). Estur. Res. Fed., S. Carolina, 715 pp.
- Schwartzlose, R.A. and J.L. Reid 1972. Nearshore circulation in the California current. *California Cooperative Oceanographic Fisheries Investigations, Reports Volume 16*:57-65.
- Segar, D.A. and E. Stamman 1986. Fundamentals of marine pollution monitoring program design. *Marine Pollution Bulletin* 17(5): 194-200.
- SFEI see San Francisco Estuary Institute
- SFEP see San Francisco Estuary Project
- Sinderman, C.J. 1979. Pollution-associated diseases and abnormalities of fish and

- shellfish: a review. *Fishery Bulletin* 76:717-749.
- Smith, R.W. 1995. The reference envelope approach to impact monitoring. Submitted to U.S. Environmental Protection Agency, Region IX. Grant No. X-009904-01-0.
- Smith, R.W. 1998. Analysis of 1994 Orange County outfall benthic data in a regional context. Submitted to County Sanitation Districts of Orange County, Fountain Valley, CA; 71 pp.
- Spies, R. 1984. Benthic pelagic coupling in sewage-affected marine ecosystems. *Marine Environmental Research* 13: 195-230.
- State Water Resources Control Board (SWRCB) 1997. California Ocean Plan. Water Quality Control Plan for Ocean Waters of California.
- State Water Resources Control Board 2012. California Ocean Plan. [http://www.swrcb.ca.gov/water\\_issues/programs/ocean/docs/oplans/oceanplan2012.pdf](http://www.swrcb.ca.gov/water_issues/programs/ocean/docs/oplans/oceanplan2012.pdf) [accessed Jan. 2, 2014]
- Stevenson, A.F. 2000. Metal concentrations in marine sediments around Scotland: a baseline for environmental studies. *Continental Shelf Research* 21 (2001) p. 879 – 897.
- Stewart-Oaten, A., W.W. Murdoch, and K.R. Parker. 1986. Environmental impact assessment: “pseudoreplication” in time? *Ecology* 67:929-940.
- Stull, J.K. 1995. Two decades of marine biological monitoring, Palos Verdes, California, 1972 to 1992. *Bulletin of the Southern California Academy of Science* 94(1): 21-45.
- Suedel, B.C., J.A. Boraczek, R.K. Peddicord, P.A. Clifford, and T.M. Dillon. 1994. Trophic transfer and biomagnification potential of contaminants in aquatic ecosystems. *Rev. Environmental Contamination and Toxicology* 136:21-89.
- Swartz, R.C., F.A. Cole, D.W. Schults and W.A. DeBen 1986. Ecological changes in the Southern California Bight near a large sewage outfall: Benthic conditions in 1980 and 1983. *Marine Ecology Progress Series* 31: 1-13.
- Swartz, R.C., D.W. Schults, G.R. Ditworth and W.A. Deben 1984. Toxicity of sewage sludge to *Rhepoxynius abronius*, a marine benthic amphipod. *Arch. Environ. Toxicol.* 13: 207-216.
- SWRCB see California State Water Resources Control Board
- Tetra Tech, Inc. 1982. Design of 301(h) monitoring programs for municipal wastewater discharges to marine waters. U.S. Environmental Protection Agency. U.S. EPA Contract No. 430-9-82-010. Tetra Tech, Inc., Bellevue, WA.
- 1985a. A summary of U.S. EPA approved methods, standard methods, and other guidance for 301(h) monitoring variables. Final program document prepared for the Marine Operations Division, Office of Marine and Estuarine Protection, U.S. Environmental Protection Agency. U.S. EPA Contract No. 68-01-6938. Tetra Tech, Inc., Bellevue, WA; 18 pp.
- 1985b. Bioaccumulation monitoring guidance: 1. Estimating the potential for bioaccumulation of priority pollutants and 301(h) pesticides discharged into marine and estuarine waters. Final program document prepared for the Marine Operations Division, Office of Marine and Estuarine Protection, U.S. Environmental Protection Agency. U.S. EPA Contract No. 68-01-6938. Tetra Tech, Inc., Bellevue, WA; 69 pp.
- 1985c. Bioaccumulation monitoring guidance: 2. Selection of target species and review of available bioaccumulation data Vol. I. Final program document prepared for the Marine Operations Division, Office of Marine and Estuarine Protection, U.S. Environmental Protection Agency. U.S. EPA Contract No. 68-01-6938. Tetra Tech, Inc., Bellevue, WA. 52 pp.
- 1985d. Bioaccumulation monitoring guidance: 2. Selection of target species and review of available bioaccumulation data Vol. II. Final program document prepared for the Marine Operations Division, Office

- of Marine and Estuarine Protection, U.S. Environmental Protection Agency. U.S. EPA Contract No. 68-01-6938. Tetra Tech, Inc., Bellevue, WA Appendices Final Program Document. 345 pp.
- 1985e. Bioaccumulation monitoring guidance: 3. Recommended analytical detection limits. Final program document prepared for the Marine Operations Division, Office of Marine and Estuarine Protection, U.S. Environmental Protection Agency. U.S. EPA Contract No. 68-01-6938. Tetra Tech, Inc., Bellevue, WA; 23 pp.
- 1985f. Recommended biological studies for 301(h) monitoring program. Final program document prepared for the Marine Operations Division, Office of Marine and Estuarine Protection, U.S. Environmental Protection Agency. U.S. EPA Contract No. 68-01-6938. Tetra Tech, Inc., Bellevue, WA; 17 pp.
- 1986a. Quality assurance and quality control (QA/QC) for 301(h) monitoring programs: guidance on field and laboratory methods. Final program document prepared for the Marine Operations Division, Office of Marine and Estuarine Protection, U.S. Environmental Protection Agency. U.S. EPA Contract No. 68-01-6938. Tetra Tech, Inc., Bellevue, WA; 267 pp.
- 1986b. Bioaccumulation monitoring guidance: 4. Analytical methods for U.S. EPA priority pollutants and 301(H) pesticides in tissues from estuarine and marine organisms. Final program document prepared for the Marine Operations Division, Office of Marine and Estuarine Protection, U.S. Environmental Protection Agency. U.S. EPA Contract No. 68-01-6938. Tetra Tech, Inc., Bellevue, WA.
- 1986c. Analytical methods for U.S. EPA priority pollutants and 301(H) pesticides in tissues from estuarine and marine sediments. Final program document prepared for the Marine Operations Division, Office of Marine and Estuarine Protection, U.S. Environmental Protection Agency. U.S. EPA Contract No. 68-01-6938. Tetra Tech, Inc., Bellevue, WA.
- 1987. Quality assurance / quality control (QA /QC) for 301(h) monitoring programs: guidance on field and laboratory methods. U.S. EPA 430/9-86-004; 267 pp. + appendices.
- 1993. Method 1625: Semi-volatile organic compounds by isotope dilutions GC/MS. Prepared by the Office of Water Regulations and Standards, Industrial Technology Division, Office of Water.
- 1996. PCBs: Cancer Dose-Response Assessment and Application to Environmental Mixtures. National Center for Environmental Assessment, Office of Research and Development. EPA/600/P-96/001F.
- 1997. Integrated Risk Information System (IRIS) 1997. [<http://www.epa.gov/ngispgm3/iris/sugst/0294.htm>]
- 1999. Polychlorinated Biphenyls (PCBs) Update: Impact on fish advisories. Office of Water, Washington, D.C. EPA-823-F-99-019.
- 2002. Implementation Guidance for Ambient Water Quality Criteria for bacteriology (Draft). EPA 823-B-02-003. [<http://www.epa.gov/ost/standards/bacteria/bacteria.pdf>]
- and San Francisco Bay Regional Water Quality Control Board (RWQCB) 1997. NPDES requirements and waste discharge
- Contract No. 68-01-6938. Tetra Tech, Inc., Bellevue, WA.
- Thompson, B., S. Lowe, and M.G. Kellogg 2000. Results of the benthic pilot study 1994-1997: Part 1 – Macrobenthic assemblages of the San Francisco Bay-Delta, and their responses to abiotic factors. San Francisco Estuary Regional Monitoring Program for Trace Substances, Technical Report 39: 40 pp.
- U.S. Environmental Protection Agency (U.S. EPA) 1983. Methods for chemical analysis of water and wastes. 3<sup>rd</sup> Ed. U.S. EPA Environmental Monitoring and Support Laboratory, Cincinnati, OH. U.S. EPA 600/4-79-030.

- requirements for City and County of San Francisco's Oceanside Water Pollution Control Plant and the Westside wet weather combined sewer system. NPDES permit #CA0037681.
- and San Francisco Bay Regional Water Quality Control Board (RWQCB) 2003. Final Order No. R2-2003-0073, NPDES permit no. CA 0037681 for the Oceanside Treatment Plant, Southwest Ocean Outfall, and Westside wet weather facilities, City and County of San Francisco.
- 2006. Water quality standards for coastal recreation waters: using single sample maximum values in state water quality standards. Office of Water, 4305T, EPA-823-F-06-013. <http://www.epa.gov/waterscience/beaches/files/SSM.pdf> [accessed 25 January 2010]
- U.S. EPA see U.S. Environmental Protection Agency
- U.S. Geological Survey (USGS) 1999. El Nino sea-level rise wreaks havoc in California's San Francisco Bay region. USGS Fact Sheet 175-99: 4 pp. [<http://geopubs.wr.usgs.gov/fact-sheet/fs175-99/>]
- USGS see U.S. Geological Survey
- Vogan, C.L., P.J. Llewellyn and A.F. Rowley 1999. Epidemiology and dynamics of shell disease in the edible crab *Cancer pagurus*: a preliminary study of Langland Bay, Swansea, UK. *Dis. Aquat. Org.* 35: 81-87.
- Vogan, C.L., C. Costa-Ramos and A.F. Rowley 2002. Shell disease syndrome in the edible crab, *Cancer pagurus* – isolation, characterization and pathogenicity of chitinolytic bacteria. *Microbiology* 148: 743-754.
- Vogan, C.L., A. Powell and A.F. Rowley 2008. Shell disease in crustaceans – just chitin recycling gone wrong? *Environmental Microbiology* 10(4): 826-835.
- Waldichuk, M. 1985. Biological availability of metals to marine organisms. *Marine Pollution Bulletin* 16:7-11.
- Water Environment Federation 2005. Parking Lot Sealants Examined as Source of Pollution. [<http://www.wef.org/PolicyAction/USGovernmentAffairs/TWIW/TWIW12092005.htm>] Downloaded 12/27/2005.
- Water Quality Bureau (WQB). 1997a. Ocean outfall monitoring program 1995 annual report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 1997b. Ocean outfall monitoring program 1996 annual report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 1998. Southwest ocean outfall regional monitoring 1997 annual report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 1999. Southwest ocean outfall regional monitoring 1998 annual report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 2000. Southwest ocean outfall regional monitoring 1999 annual report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 2001a. Southwest ocean outfall regional monitoring program 2000 annual report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 2001b. Ocean Beach Recreational use survey, September 1998 to October 2000. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 2003a. Southwest ocean outfall regional



- monitoring program Five-Year Summary Report 1997-2001. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 2003b. Southwest ocean outfall regional monitoring report program 2002 data report. City and County of San Francisco. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- 2004. Southwest ocean outfall regional monitoring report program 2003 data report. City and County of San Francisco, Public Utilities Commission. Submitted to U.S. EPA, Region 9 and S.F. Bay Regional Water Quality Control Board.
- WEF see Water Environment Federation.
- Western Regional Climate Center (WRCC)  
1998. "El Nino, La Nina, and the western U.S., Alaska and Hawaii." [<http://www.wrcc.dri.edu/enso/ensofaq.html>]
- 1999. "Monthly total precipitation at Mission Dolores, San Francisco, California." [<http://www.wrcc.dri.edu/cgi-bin/cliMAIN.pl?casfod+sfo>]
- Western Regional Climate Center 2013. San Francisco Richmond [District], California (047767) period of record monthly climate summary period of record: 7/ 1/1948 to 2/28/2013. <http://www.wrcc.dri.edu/cgi-bin/cliMAIN.pl?ca7767> [accessed: 20 March 2014]
- Westernhagen H.V. and V. Dethlefsen. 1975. Combined effects of cadmium and salinity on development and survival of flounder eggs. *Journal of the Marine Biological Association of the United Kingdom* 55:945-957.
- White, D., A.J. Kimerling, and W.S. Overton 1992. Cartographic and geometric components of a global sampling design for environmental monitoring. *Cartography and Geographic Information Systems* 19(1):5-22.
- Word, J.Q. 1978. The infaunal trophic index. Pp. 19-40 in: W. Bascom, ed. Annual Report for the Year 1978. Southern California Coastal Water Research Project, Long Beach, CA; 251 pp.
- and A.J. Mearns 1978. The sixty-meter control survey. pp. 41-56 in: Southern California Coastal Water Research Project: 1978 Annual Report.
- B.L. Myers, and A.J. Mearns 1977. Animals that are indicators of marine pollution. Pp. 199-206 in: W. Bascom, ed. Annual Report for the Year 1977. Southern California Coastal Water Research Project, Long Beach, CA.
- WQB see Water Quality Bureau
- WRCC see Western Regional Climate Center
- Zmarzly, D.L., T.D. Stebbins, D. Pasko, R.M. Duggan, and K.L. Barwick 1994. Spatial patterns and temporal succession in soft-bottom macroinvertebrate assemblages surrounding an ocean outfall on the southern San Diego shelf: relation to anthropogenic and natural events. *Marine Biology* 118: 293-307.



**APPENDIX A  
FACILITIES AND SETTING**

# APPENDIX A FACILITIES, MONITORING HISTORY, AND SETTING

## A.1 FACILITIES AND SYSTEM

### A.1.1 COMBINED SEWER SYSTEM

The City and County of San Francisco (City) has a combined sewer system that collects domestic sanitary flow, industrial effluents, and urban stormwater runoff. Components of the system include a network of sewer pipes, catch basins and transport structures. Prior to 1982, the collected wastewater was treated at three primary treatment plants, which had a combined wet weather capacity of approximately 225 million gallons per day

(MGD). The three treatment plants, shown in Figure A-1, included the Southeast Water Pollution Control Plant (WPCP) that served the central and southeastern sections of San Francisco, the North Point WPCP that served the central and northeastern sections of San Francisco, and the Richmond-Sunset WPCP that served the western section of San Francisco. When wet weather caused combined flows to exceed the hydraulic capacity of the combined system, the untreated wastewater was bypassed into San Francisco Bay or the Pacific Ocean. Untreated overflows typically occurred whenever rainfall exceeded 0.02 inches per hour, and contributed to elevated coliform bacteria levels in near shore waters that exceeded public health standards. Swimming beaches were posted throughout the winter, shell fishing was banned, and the aesthetic appeal of the shoreline and coastal areas was greatly diminished due to the presence of sewage-derived floatable materials.



Figure A-1  
Components of the San Francisco combined sewer system

In order to reduce overflows, the City prepared a Master Plan for Wastewater Management (DPW 1971). The plan was subsequently updated in 1974 to address requirements in the Federal Water Pollution Act of 1972 that wastewater treatment plants be upgraded to provide full secondary treatment. The main elements of the Wastewater Management Master Plan were (1) upgrade sewage treatment from primary to secondary, (2) construct transport/storage sewage collection systems to reduce combined sewer overflows into receiving waters, and (3) provide for the eventual discharge of all wastewater into the Pacific Ocean via an offshore outfall. A massive construction program was put in effect to meet the goals of the Master Plan.

#### A.1.1.1 Westside Core System

The Richmond-Sunset WPCP was built in 1939 to provide primary wastewater treatment for predominantly domestic sanitary flow and to reduce raw sewage overflows onto Ocean Beach. The Richmond-Sunset WPCP had a wet weather capacity of 40 MGD. The Westside Core System (Figure A-1) including the Westside Pump Station and the Southwest Ocean Outfall (SWOO) became operational in 1986. Effluent from the Richmond-Sunset WPCP was discharged near Lands End through the Mile Rock outfall (Figure A-1) until September 1986 when it was diverted to the SWOO.

The SWOO had the capacity to discharge both dry and wet weather flow from the Richmond-Sunset WPCP as well as wet weather flow intercepted by the Westside Transport. The Westside Transport went on line in January 1987 and provided an extra 48 million gallons of storage capacity. The Lake Merced Transport was completed in July 1993 and provided an extra 11 million gallons of storage capacity.

The Oceanside WPCP was built as a secondary treatment facility to replace the Richmond-Sunset WPCP and was completed in September 1993. The plant became operational on September 18, 1993 with primary treatment, and went to full secondary treatment on September 27, 1993. The treatment plant has a wet weather capacity of 65

MGD, of which 43 MGD receive full secondary and the remaining 22 MGD receive primary treatment that is blended with secondary flow. The final component of the system, the Richmond Transport, was completed in January 1997, with a storage capacity of 10 million gallons.

#### A.1.1.2 Bayside Core System

Construction of new sewage facilities in the Bayside Core System (Figure A-1), including the Southeast WPCP, the Channel and Northshore Pump Stations, and conversion of the North Point WPCP to the North Point Wet Weather Facility was completed in 1982. The Southeast WPCP was upgraded to full secondary with a wet weather capacity of 210 MGD, of which 145 MGD were provided with full secondary treatment and the remaining 65 MGD received primary treatment that was blended with secondary flow. Flow from the northeast section of San Francisco during dry weather periods is pumped from the Northshore Pump Station to the Channel Pump Station and then on to the Southeast WPCP for secondary treatment. The North Point facility was converted from a full time primary treatment plant to a wet weather facility providing primary treatment and operating only during wet weather periods. During wet weather, when the hydraulic capacity of the Southeast facilities is maximized, the North Point Wet Weather Facility can provide primary treatment for up to 150 MGD of combined storm water and wastewater.

Other Southeast area improvements included the completion of the Yosemite Facilities and the Griffith Pump Station in 1989, the Sunnydale Facilities in 1991 and the Mariposa Facilities in 1992. The Islais Creek Transport was completed in the summer of 1997. In order to address concerns regarding treated effluent discharged into Islais Creek during wet weather, improvements to the Southeast WPCP were designed and added to the Islais Creek Transport project, which extended the original estimated completion date of 1996. Islais Creek is a shallow-water, dead end channel. During wet weather periods, when the pumping capacity to the Southeast WPCP deep-water outfall to San Francisco Bay was exceeded (100 MGD), a

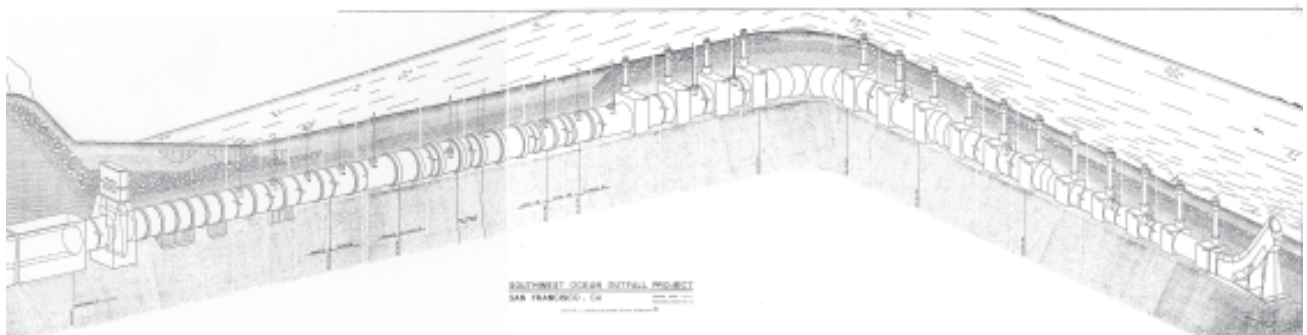
combination of primary and secondary treated effluent was discharged into Islais Creek. Such waste discharges that receive less than 10:1 initial dilution are prohibited in California surface waters and are in violation of requirements in the San Francisco Bay Basin Plan (RWQCB 1986, 1995). Discharges to the Creek occurred approximately 600 hours per year during wet weather. Because of the infrequency of the discharge, the City requested an exception to the Basin Plan from the San Francisco Bay Regional Water Quality Control Board (RWQCB). The exception was granted contingent upon improvements to the Southeast WPCP completed in 1997, which increased the wet weather capacity from 210 MGD to 250 MGD and guaranteed any discharges into Islais Creek would undergo full secondary treatment. Further improvements to the Booster Pump Station at Islais Creek in 2001 increased the pumping capacity to the deep-water outfall from 100 MGD to 110 MGD.

#### A.1.1.3 Southwest Ocean Outfall (SWOO)

The SWOO is a steel reinforced concrete pipe that extends approximately 7 km (3.75 miles) west-southwest, offshore of the Oceanside WPCP terminating at an approximate depth of 24 meters (mean lower low water) in the near shore Gulf of the Farallones. The end of the outfall consists of a

diffuser section approximately 900 meters in length and 3.5 meters in diameter, with risers located every 11 meters that discharge effluent (Figure A-2). Each riser is constructed with eight discharge ports that permit uniform flow to leave the diffuser and maximize dilution by the receiving water.

The hydraulic capacity of the SWOO pipeline with all risers operational is 575 MGD. Because the average daily dry weather flow through the SWOO is only 18 MGD, the circulation of saltwater through the diffusers would impede the discharge of treated effluent under normal dry weather operations if all the diffusers were open. Maintenance of the SWOO diffuser system requires manual manipulation under “hard-hat” diving conditions, making efforts to open or close the diffuser ports as needed to accommodate changes between dry and wet weather flow rates prohibitive. Therefore, a hydraulic analysis was conducted to determine optimum conditions to maintain an adequate port velocity to allow discharge through the ports during both dry and wet weather flow conditions. Results of the analysis indicated average dry weather flow is adequately dispersed to the environment with 21 of the 85 risers operational, and a maximum wet weather flow of 175 MGD can be effectively discharged offshore. Alternate risers located along the outer 460 meters of the diffuser section are active. The Southwest Ocean Outfall became operational in 1986.



*Figure A-2  
Southwest ocean outfall pipeline and diffusers*

## A.2 MONITORING STUDIES

### A.2.1 PRE-DISCHARGE STUDIES

Outfall construction was preceded by pre-design physical, chemical, and biological investigations to determine the optimum outfall location. Biological studies included an analysis of the plankton (Brown & Caldwell 1971a, 1971b), the intertidal zone (Brown & Caldwell 1973), the benthic infauna community, demersal fish and epibenthic invertebrate communities (Brown & Caldwell 1975a, 1975b), and bacterial die-off rates (CH<sub>2</sub>M Hill and Woodward-Clyde Consultants 1978). Pre-design physical and chemical data collections were conducted from 1977 to 1978 and included oceanographic and water quality measurements, plume behavior studies and a geophysical assessment (CH<sub>2</sub>M Hill and Woodward-Clyde Consultants 1978).

During outfall design and construction, pre-discharge oceanographic studies were conducted to further define the baseline conditions at the proposed discharge site. Seasonal sampling was conducted at permanent stations in the near shore Gulf of the Farallones (CH<sub>2</sub>M Hill 1980). Specifications for the sampling program were developed in conjunction with the State Water Resources Control Board (SWRCB) and the Marine-Estuarine Technical Committee (METC). Phase I (June 1978 to September 1979) included a continuation of the pre-design receiving water quality program, a benthic infauna community assessment, an assessment of demersal fish and epibenthic invertebrate communities, an analysis of trace metal and organic pollutants in the water, sediments, and fauna, and a review of existing biological and chemical literature and data. Phase II (October 1979 to November 1980) was a continuation of the Phase I program and also included life history studies of four selected benthic infauna species (CH<sub>2</sub>M Hill 1983).

In 1982, the City undertook the Pre-discharge Monitoring Program. Monitoring was conducted in the vicinity of the future discharge site following the same study design used previously by CH<sub>2</sub>M Hill.

To ensure the continuation of a comparative pre-discharge database, every effort was made to preserve procedures and methods between the two studies. This long-term pre-discharge monitoring program focused on physical and chemical water quality and sediment measurements, an assessment of benthic infauna, demersal fish, and epibenthic invertebrate communities, and an analysis of trace metal and organic pollutants in the sediment and fauna. Surveys were conducted three times per year and are summarized for the years 1982-83 and 1983-84 in monitoring program annual reports (BWPC 1984, 1985). Data from October 1984 through 1986 have not been compiled in report format.

### A.2.2 POST-DISCHARGE STUDIES

In 1979 the City requested a variance from secondary treatment requirements for the Richmond-Sunset WPCP as allowed under the Clean Water Act, Section 301(h). A tentative decision, which granted the discharge of less than secondary-treated sewage to the Pacific Ocean, was issued under an Administrative Order in September of 1986 (U.S. EPA and RWQCB 1986). The order allowed for diversion of wastewater from the Richmond-Sunset WPCP to the newly constructed SWOO for discharge into the Pacific Ocean approximately 3.75 miles offshore beginning in September 1986.

#### A.2.2.1 Temporary Monitoring Requirements (1987 to 1989)

A revised temporary monitoring plan for the new discharge site was required in the Administrative Order (U.S. EPA 1986). The new monitoring plan was an expanded version of the pre-discharge plan and included a greater emphasis on water quality monitoring. Shoreline bacteria measurements increased to three times per week year round. The frequency of offshore water quality monitoring increased to monthly surveys. Sampling frequencies for the biological community and pollutant studies remained the same. Summaries of surveys conducted under the Administrative Order during

1987, 1988, and 1989 are presented in annual reports (BWPC 1988, 1989, 1990).

#### A.2.2.2 Wastefield Transport and Bacteriological Studies

In 1987 the City, jointly with CH<sub>2</sub>M Hill and under the guidance of the U.S. EPA, conducted wastefield transport and bacteriological compliance studies to determine movement of the effluent plume (CH<sub>2</sub>M Hill 1989). Rhodamine dye was injected into the effluent at the Westside Pump Station. Concentrations of dye were measured in the receiving water with a fluorometer. Once dye concentrations were found in the receiving water, drogues were deployed to evaluate effluent plume transport. Results from these studies determined that the minimum initial dilution of the effluent plume, calculated using dye study data was 100:1, and that the effluent plume moves in a path toward and away from San Francisco Bay, influenced by flooding and ebbing tides. These studies determined that the effluent plume never reached Seal Rocks at Point Lobos outside of the Golden Gate, nor did it move in an onshore direction toward Ocean Beach. Based upon these studies, the U.S. EPA and RWQCB determined that chlorination and subsequent de-chlorination of the effluent was unnecessary for the protection of public health. Shoreline bacteria monitoring conducted year round continues to document that the effluent plume does not reach the shoreline.

#### A.2.2.3 1990 NPDES Permit Monitoring Program (1990 to 1996)

The City withdrew its request for a variance from secondary treatment requirements in 1989 and began measures to design the Oceanside WPCP as a full secondary dry weather treatment facility. Based on that decision, and after reviewing results from the existing monitoring program and the wastefield and bacteriological compliance studies, the temporary monitoring plan issued under the Administrative Order was revised to a long term National Pollutant Discharge Elimination System (NPDES) monitoring program. The NPDES permit

was issued in July 1990 and the receiving water monitoring program was implemented immediately.

The program was designed using a site-specific monitoring strategy to determine whether the outfall contributed to environmental impacts either through physical disturbances or pollutant loading. Stations were located to characterize the outfall zone of initial dilution (ZID), near field and far field areas around the outfall, and a reference site (Figure A-3). This traditional monitoring approach compares impact site characteristics with a reference site, and has the expectation of a gradient of impact between the ZID and the reference site.

The site-specific monitoring program under the NPDES permit differed from the temporary monitoring program implemented under the Administrative Order primarily by reduction of monitoring frequency, elimination of certain analytical parameters, and relocation of sampling stations. Water quality surveys were reduced from monthly to

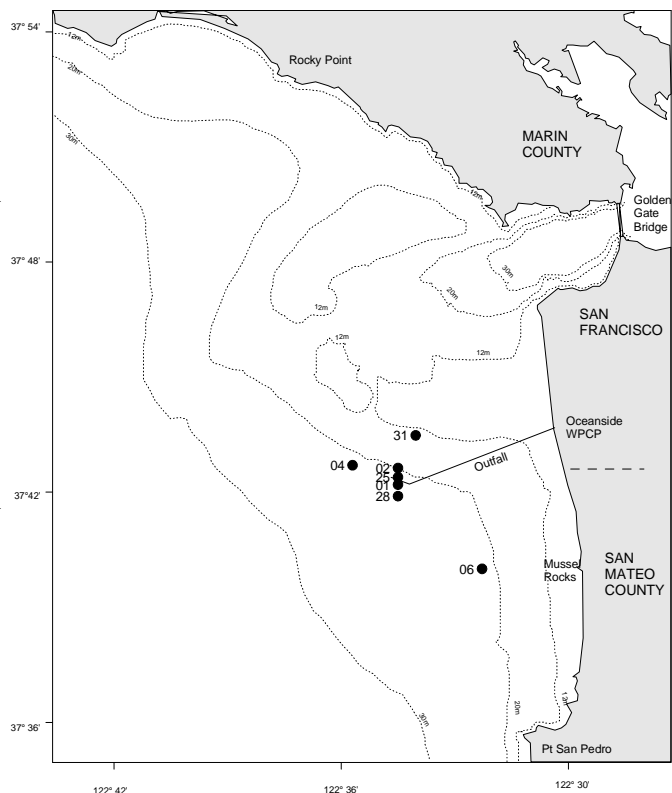


Figure A-3  
Site specific monitoring program

quarterly. Demersal fish and epibenthic invertebrate studies, benthic infauna studies, and marine sediments studies were reduced from tri-annual to semiannual, and trace metal and organic contaminant studies for organisms and sediment were conducted annually. Shoreline bacteriological sampling remained unchanged. Offshore stations were relocated to be in alignment with the movement of the effluent plume as determined by the wastefield transport studies (CH<sub>2</sub>M Hill 1989). Surveys conducted under the site-specific monitoring permit during 1990, 1991, 1992, 1993, 1994, 1995 and 1996 are summarized in annual reports (BWPC 1992a, 1992b, 1993, 1994, 1995; WQB 1997a, 1997b).

The specific components and objectives of the site-specific program included water quality studies to determine compliance with the California Ocean Plan water quality standards (SWRCB 1997), the location and extent of the wastewater plume, stratification in the water column, and onshore transport of bacteria found in the sewage effluent. Shoreline bacteria levels were monitored to warn the public about bacteria contaminated waters resulting from combined sewer overflows or other sources. Benthic monitoring included sediment studies to evaluate the spatial distribution of physical and chemical sediment characteristics, and to determine the accumulation of organic and inorganic contaminants in sediments in the vicinity of the ZID; and infauna community analyses to determine the presence or absence of a balanced indigenous population. Demersal fish and epibenthic invertebrate studies included community analyses to determine the presence or absence of a balanced indigenous assemblage of species; and bioaccumulation monitoring was conducted to determine the accumulation of organic and inorganic contaminants in the tissues of commercially important species and the potential for transfer to higher trophic levels.

Analyses of ten years of site-specific monitoring data indicated that seasonality was the predominant factor affecting differences between sites in the study area (BWPC 1988, 1989, 1990, 1992a, 1992b, 1993, 1994, 1995; WQB 1997a, 1997b). No evidence of the effluent plume in the water column

was detected. Near shore bacteria monitoring provided useful public health information and indicated that no onshore transport of the effluent plume occurred. Sediment chemistry and bioaccumulation studies indicated contaminant concentrations were not unusually elevated. El Niño events contributed to between-year fluctuations in community measures (Kellogg, et al. 1998). A single reference station in the site-specific monitoring program did not provide sufficient characterization of reference conditions. With the large amount of natural variability that existed in sediment and biota, the determination of differences between stations was difficult.

### **A.3 SETTING**

#### **A.3.1 GULF OF THE FARALLONES**

Monitoring locations in the SWOO study area lie on the continental shelf within the area known as the Gulf of the Farallones, bordered by Point Reyes on the north and Point San Pedro on the south and extending about 26 nautical miles west of the Golden Gate, to the Farallon Islands. The primary influences on the near shore water quality and sediment characteristics within the Gulf include the broad changes in wind and current conditions that define the oceanographic seasons, tidal ebb, flood, outflow from San Francisco Bay (Brown and Caldwell 1971a,b), and the complex topography of the shelf in this area (Noble and Gelfenbaum 1988). The San Francisco Estuary is a major supplier of fine sediments to the Gulf (Noble and Gelfenbaum 1988), with the magnitude of the effects depending on the season and amount of freshwater outflow from the Sacramento-San Joaquin Rivers. Sediment transport from the estuary also has the potential to transport nutrients and contaminants into the study area. The bathymetry of the continental shelf in the Gulf of the Farallones, with its broad shoaling region east of the Farallon Islands and the northern barrier formed by the Point Reyes Peninsula, may have a major effect on the along-shelf and cross-shelf flows, blocking upwelling and affecting sediment transport in the central regions of the Gulf (Noble and Gelfenbaum 1988).



## A.3.2 OCEANOGRAPHIC SEASONS

The California near shore marine climate consists of two major seasons: the California Current season during which the principal near shore current flow is southerly; and the Davidson Current season during which the principal near shore current direction is northerly. The California Current season comprises an upwelling and an oceanic period (Bolin and Abbott 1963, Pavlova 1966, Schwartzlose and Reid 1972).

### A.3.2.1 California Current Season

The California current season usually occurs between February or March and November and is divided into an upwelling period and an oceanic period. The current originates near the Canadian border and initially contains water characteristic of the Subarctic current and North Pacific current. As this water moves southward along the Pacific coast, the surface characteristics are modified by solar heating and by the effects of river inflow and exchange with estuaries and embayments.

#### A.3.2.1.1 Upwelling Period

During the upwelling period, usually beginning in February or March and extending to August or September, stationary high pressure systems offshore produce reasonably persistent north and northwest winds. Due to Coriolis force, these persistent winds along the California coast cause the surface waters to move westward (offshore). Colder, nutrient-rich waters from depth replace the westward moving surface waters. The persistence of winds determines the depth from which upwelling water is derived, as well as the duration of the upwelling. Weather systems are seldom stationary, and thus upwelling may occur sporadically during this period.

#### A.3.2.1.2 Oceanic Period

In the late summer and fall (August or September to November), the north and northwest

winds subside and upwelling ceases. That portion of the California Current season between the cessation of upwelling and the start of the Davidson Current season is identified as the oceanic period. During the oceanic period both ocean surface temperatures and salinities are at maxima.

### A.3.2.2 Davidson Current Season

From approximately November to February or March the northward flowing Davidson Current displaces the California Current offshore. During the rainy season, low-pressure systems offshore produce south and southwest winds along the central California coast. Through Coriolis force these winds produce onshore surface water movements. These onshore currents are blocked by the northwest trending coast and gain a northerly direction that generates the Davidson Current. Because the low pressure systems do not remain stationary, the Davidson Current does not occur at all times and the end of the Davidson Current period can be diffuse and difficult to pinpoint (Bolin and Abbot 1963).

## A.3.3 EL NIÑO AND LA NIÑA EVENTS

The intermittent oceanographic phenomena known as El Niño and La Niña have global weather consequences and may significantly impact water quality and sediment transport in the Gulf of the Farallones by altering normal seasonal climate patterns. El Niño events are characterized by warmer than normal sea-surface temperatures in the equatorial Pacific Ocean. La Niña events are characterized by colder than normal sea-surface temperatures in the equatorial Pacific Ocean. Both types of events can vary in strength and local effects are difficult to predict. Typical El Niño winters are wetter in the southwest United States from southern California eastward through Arizona, southern Nevada and Utah, New Mexico, and into Texas; but drier in the northwest including Washington, Oregon, and the mountainous portions of Idaho, western Montana, and northwest Wyoming. Northern and central California lie in a zone between these two areas and can experience either effect

(WRCC 1998). Generally, in all these regions, La Niña climate effects are approximately, but not exactly, opposite to El Niño climate effects (WRCC 1998). Locally, El Niño winters have included both greater than normal precipitation and drought. Oceanographically, the primary local effects of wet El Niño events are intensified storms and sustained southwest winds that reduce upwelling and result in higher than normal sea surface temperatures (USGS 1999). An unusually strong El Niño event occurred during in 1997-1998 (NOAA 1999a), with over two times the normal annual rainfall recorded in San Francisco (WRCC 1999). This event was followed by a La Niña that caused unusually strong upwelling of cold, nutrient-rich waters off the northern California coast (USGS 1999).

#### A.3.4 NATIONAL MARINE SANCTUARIES

Three national marine sanctuaries lie partially within or adjacent to the Gulf of the Farallones. Data collected from the SWOO regional monitoring program provide important information relevant to the marine habitat management goals of these marine sanctuaries.

##### A.3.4.1 Monterey Bay National Marine Sanctuary

The SWOO is surrounded on three sides by the boundary of the Monterey Bay National Marine Sanctuary (MBNMS) (Figure A-4), created in 1992. The Sanctuary includes the waters of Monterey Bay and the Pacific Ocean extending from southern Marin County southward to Cambria in San Luis Obispo County. The MBNMS is 348 nautical miles north to south and extends an average of 30 nautical miles offshore. An exclusion zone which extends off the north coast of San Mateo County and the City and County of San Francisco between Point Bonita and Point San Pedro was originally created to encompass the SWOO, the shipping channel providing access to and from San Francisco Bay, and the Golden Gate dredged material disposal site associated with the shipping channel (NOAA 1992). The sanctuary is managed to balance recreational and commercial uses with protection of natural resources, water quality,

habitats, and its bountiful resident and migratory marine life. Major resource management issues in the sanctuary are vessel traffic and its potential impact on living resources and water quality, disposal of dredged material, land-based sources of water pollution, direct and indirect fishing impacts, and impacts of non-native, invasive species (NOAA 1992, 1999a). Ten stations of the SWOO Regional Monitoring Program lie within the MBNMS.

##### A.3.4.2 Gulf of the Farallones National Marine Sanctuary

Adjacent to the northwest MBNMS boundary, the Gulf of the Farallones National Marine Sanctuary (GFNMS) (Figure A-5), designated in 1981, encompasses 948 square nautical miles including the Farallon Islands on the western edge of the Gulf and near shore tidal flats, rocky intertidal areas, wetlands, subtidal reefs, and coastal beaches north and west of San Francisco. Southeast Farallon Island, 26 nautical miles west of the Golden Gate Bridge in the south central part of the sanctuary, is a national wildlife refuge, with resting and breeding sites for marine mammals and seabirds which benefit from the nutrient-rich waters in the area. The sanctuary has thousands of seals and sea



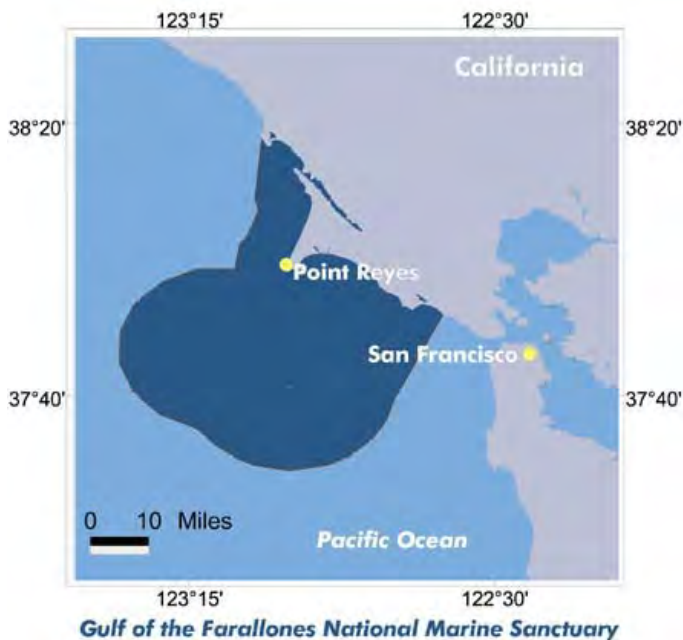
Figure A-4  
Monterey Bay National Marine Sanctuary  
(NOAA 1999a)

lions, and is home to the largest concentration of breeding seabirds in the contiguous United States. The resources are protected by managing human activities that may damage habitat and species, supporting restoration projects to revitalize disturbed areas, and conducting monitoring programs to assess changes from natural and human disturbance. Major resource management issues in the sanctuary are oil spills, sewage, toxic chemicals, petroleum products, pesticides, and urban runoff that threaten sanctuary waters (NOAA 1999b). Five stations of the SWOO Regional Monitoring Program lie within the GFNMS.

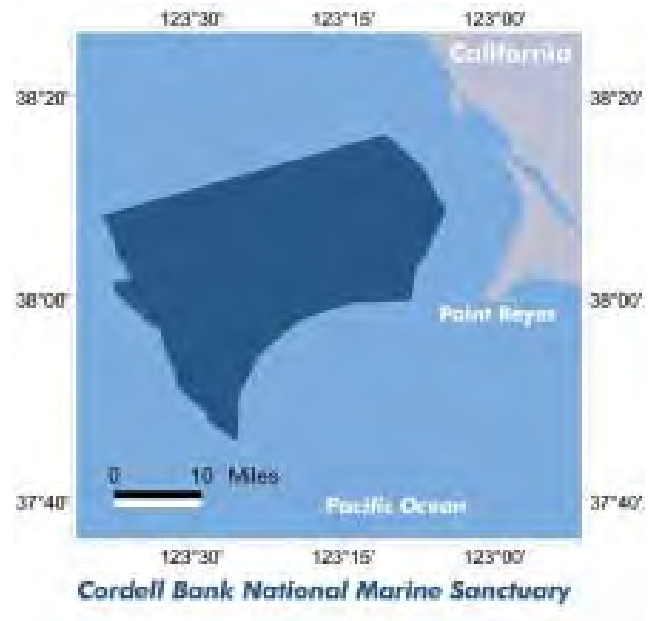
#### A.3.4.3 Cordell Bank National Marine Sanctuary

Cordell Bank National Marine Sanctuary (Figure A-6) is an offshore sanctuary about 43 nautical miles northwest of the Golden Gate Bridge. Near the edge of the continental shelf, Cordell Bank

rises from the sea floor as the northern most expression of the Farallon Ridge. The water is about 200 feet deep over most of the bank. Along a few of its ridges and pinnacles, this submerged island rises to within 120 feet of the ocean surface. Upwelling of nutrient rich ocean waters and the bank's topography create one of the most biologically productive areas on the West Coast. Cordell Bank was designated a marine sanctuary in 1989 in recognition of the significant value of this marine habitat. The sanctuary covers 397 square nautical miles of Pacific Ocean including and surrounding Cordell Bank. Algae, fish, and invertebrates proliferate, and the site is a lush feeding ground for many marine mammals and seabirds. Endangered humpback whales, Dall's porpoises, albatross, shearwaters, and countless other marine species flourish in this rich marine environment (NOAA 1999c).



*Figure A-5*  
*Gulf of the Farallones National Marine Sanctuary (NOAA 1999b)*



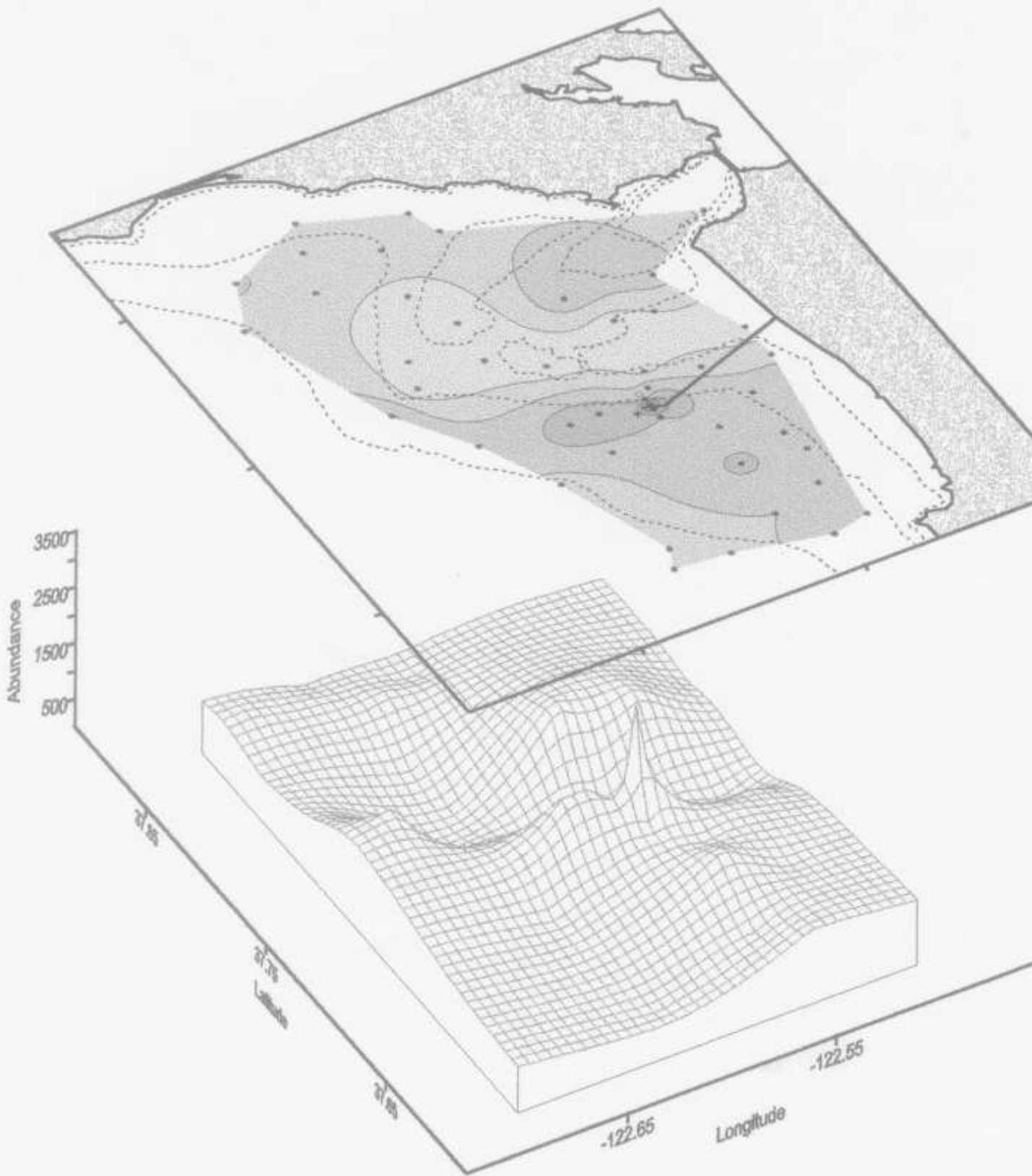
*Figure A-6*  
*Cordell Bank National Marine Sanctuary (NOAA 1999c)*

## A.4 REFERENCES CITED

- Bolin, R.L. and D.P. Abbott 1963. Studies on the marine climate and phytoplankton of the central coastal area of California, 1954-1960. California Cooperative Oceanic Fisheries Investigations, Reports Volume IX:23-45.
- Brown and Caldwell 1971a. A pre-design report on marine waste disposal. Vol. I. Oceanographic and ecological base data acquisition and evaluation of alternative locations. Prepared for the City and County of San Francisco, 214 pp.
- 1971b. A pre-design report on marine waste disposal. Vol. II. Data supplement. Oceanographic and ecological base data acquisition and evaluation of alternative locations. Prepared for the City and County of San Francisco.
- 1973. A pre-design report on marine waste disposal. Vol. III. Supplementary ecological investigations. Prepared for the City and County of San Francisco, 33 pp.
- 1975a. A pre-design report on marine waste disposal. Vol. IV. 1973-1974, Investigations and preliminary design. Prepared for the City and County of San Francisco, 198 pp.
- 1975b. A pre-design report on marine waste disposal. Vol. V. Data supplement. 1973-1974, Investigations and preliminary design. Prepared for the City and County of San Francisco.
- Bureau of Water Pollution Control (BWPC) 1984. Southwest ocean outfall monitoring program 1982-83 annual report. City and County of San Francisco, Department of Public Works.
- 1985. Southwest ocean outfall monitoring program 1983-84 annual report. City and County of San Francisco, Department of Public Works.
- 1988. Southwest ocean outfall monitoring program 1987 annual report. City and County of San Francisco, Department of Public Works.
- 1989. Southwest ocean outfall monitoring program 1988 annual report. City and County of San Francisco, Department of Public Works.
- 1990. Southwest ocean outfall monitoring program 1989 annual report. City and County of San Francisco, Department of Public Works.
- 1992a. Southwest ocean. Ocean outfall monitoring program 1990 annual report. City and County of San Francisco, Department of Public Works.
- 1992b. Southwest ocean outfall monitoring program 1991 annual report. City and County of San Francisco, Department of Public Works.
- 1993. Southwest ocean outfall monitoring program 1992 annual report. City and County of San Francisco, Department of Public Works.
- 1994. Southwest ocean outfall monitoring program 1993 annual report. City and County of San Francisco, Department of Public Works.
- 1995. Southwest ocean outfall monitoring program 1994 annual report. City and County of San Francisco, Department of Public Works.
- BWPC see Bureau of Water Pollution Control California State Water Resources Control Board (SWRCB) 1983. Draft environmental impact report for the water quality plan for ocean waters of California.
- 1997. California Ocean Plan - amendment to the water quality control plan for ocean waters of California. California Environmental Protection Agency. Resolution # 97-026.
- CH<sub>2</sub>M Hill 1980. Pre-discharge oceanographic monitoring, southwest ocean outfall project, Phase I - final report.
- 1983. Pre-discharge oceanographic monitoring, southwest ocean outfall project, Phase II - final report.
- 1989. Wastefield transport and bacteriological compliance studies of the San Francisco ocean outfall.
- CH<sub>2</sub>M Hill and Woodward-Clyde Consultants 1978. Pre-design oceanographic studies.

- DPW see Department of Public Works
- Department of Public Works (DPW) 1971. San Francisco master plan for wastewater management. City and County of San Francisco.
- U.S. EPA see U.S. Environmental Protection Agency
- Kellogg, M.G., L. Riege, A. Navarret, and R.W. Smith 1998. San Francisco ocean monitoring program: analysis of long term data. Pp. 1606-1618 *in*: Magoon, O.T., *et al.*, Eds. California and the world ocean '97: Taking a look at California's ocean resources: an agenda for the future. Conference Proceedings, Vol. 2, American Society of Civil Engineers, Reston, VA; 1756 pp.
- National Oceanic and Atmospheric Administration 1992. Monterey Bay National Marine Sanctuary. Final Environmental Impact Statement/Management Plan. Volume II: Appendices. U.S. Department of Commerce National Oceanic and Atmospheric Administration Sanctuaries and Reserves Division.
- 1999a. Monterey Bay National Marine Sanctuary. Internet address: <http://www.sanctuaries.nos.noaa.gov/oms/omsfarallones/omsfarallones.html>
- 1999b. Gulf of the Farallones National Marine Sanctuary. Internet address: <http://www.sanctuaries.nos.noaa.gov/oms/omsfarallones/omsfarallones.html>
- 1999c. Cordell Bank National Marine Sanctuary. Internet address: <http://www.sanctuaries.nos.noaa.gov/oms/omscordell/omscordell.html>
- NOAA see National Oceanic and Atmospheric Administration
- Nobel, M. and G. Gelfenbaum 1988. A proposal for a pilot study of sediment transport and currents in the Gulf of the Farallones region. U.S. Geological Survey, Open File Report.
- Pavlova, Y.V. 1966. Seasonal variations of the California current. *Oceanology* 6:806-814.
- RWQCB see San Francisco Bay Regional Water Quality Control Board
- San Francisco Bay Regional Water Quality Control Board (RWQCB) 1986. Water quality control plan for the San Francisco Bay Basin Region (2) 1986.
- San Francisco Bay Regional Water Quality Control Board (RWQCB) 1995. Amendments to the San Francisco Bay basin plan adopted July 20, 1995.
- Schwartzlose, R.A. and J.L. Reid 1972. Nearshore circulation in the California current. California Cooperative Oceanographic Fisheries Investigations, Reports Volume 16:57-65.
- SWRCB see California State Water Resources Control Board
- U.S. Environmental Protection Agency (U.S. EPA) 1986. Findings of violation and order for compliance in the matter of the City and County of San Francisco, Richmond-Sunset Water Pollution Control Plant. Docket No. IX-FY87-7 U.S. EPA Region 9.
- U.S. Environmental Protection Agency (U.S. EPA) and San Francisco Bay Regional Water Quality Control Board (RWQCB) 1997. NPDES requirements and waste discharge requirements for City and County of San Francisco's Oceanside Water Pollution Control Plant and the Westside wet weather combined sewer system. NPDES permit no. CA0037681.
- United States Geological Survey (USGS) 1999. El Nino sea-level rise wreaks havoc in California's San Francisco Bay region. USGS Fact Sheet 175-99: 4 pp. [<http://geopubs.wr.usgs.gov/factsheet/fs175-99/>]
- U.S. EPA see U.S. Environmental Protection Agency
- USGS see United States Geological Survey
- Water Quality Bureau (WQB) 1997a. Ocean outfall monitoring program 1995 annual report. City and County of San Francisco, Public Utilities Commission.
- 1997b. Ocean outfall monitoring program 1996 annual report. City and County of San Francisco, Public Utilities Commission.
- Western Regional Climate Center (WRCC) 1998. "El Nino, La Nina, and the western U.S., Alaska and Hawaii." [<http://www.wrcc.dri.edu/enso/ensofaq.html>]

————— 1999. “Monthly total precipitation at  
Mission Dolores, San Francisco, California.”  
[[http://www.wrcc.dri.edu/cgi-bin/  
cliMAIN.pl?casfod+sfo](http://www.wrcc.dri.edu/cgi-bin/cliMAIN.pl?casfod+sfo)]  
WQB see Water Quality Bureau  
WRCC see Western Regional Climate Center



## APPENDIX B

## STATISTICS METHODS

# APPENDIX B

## STATISTICAL METHODS

by

Michael L. Johnson, Dorothy Norris and Robert W. Smith<sup>1</sup>

### B.1. SUMMARIZATION OF SEDIMENT GRAIN SIZE DATA

Using the fractions of sediment in the different measured size classes, the standard sediment-size distribution parameters of mean, standard deviation, skewness, and kurtosis are computed for each sediment sample (Folk and Ward 1957). The sediment-size intervals are expressed in phi ( $\phi$ ) units, which are computed as  $-\log_2(\text{size in } m)$ . Higher values in phi units indicate finer sediments.

The mean of the sediment size distribution is a measure of the overall sediment size. The standard deviation (SD) is a measure of the sorting of the sediment. A sediment sample with a relatively large SD contains a wide range of sediment sizes, and is poorly sorted. Thus, the SD and the degree of sorting are inversely proportional. Skewness is a measure of the asymmetry of the sediment-size distribution. Sediment distributions with positive skewness have an elongated tail toward the finer sizes (higher phi intervals), and distributions with negative skewness have an elongated tail toward the coarser sediments. Kurtosis is a measure of the peakedness of the sediment distribution in relation to the tails distribution. A sediment sample with a relatively larger proportion of sediment in the middle of the distribution (compared to the tails) has a higher value of kurtosis.

The sediment grain size summary statistics and percentages in the individual grain size categories were used in a Principal Components Analysis to understand the patterns of variation in grain size in the sediment data. Correlations between the summary statistics and the percentages in the phi categories were performed as part of the PCA. The mean phi value was positively correlated with phi values between 3 and 4 and significantly negatively correlated with phi values between 0 to 1 and 1 to 2 indicating the prevalence of small grained sediments across the majority of the sample sites.

There are two rules of thumb for retaining

eigenvalues and their corresponding eigenvectors in a PCA. The first rule is to retain all values with eigenvalues greater than 1.0 as this is the value at which their explanatory power exceeds the explanatory value of an individual variable. The second rule involves using a scree plot; plotting the value of the eigenvalue from largest to smallest and determining the location where the values “level off”, i.e., the percent of the variation in the dataset that is explained by the axes does not vary substantially and is at some arbitrarily low value.

Nine of the 13 eigenvalues had values larger than 1.0 and the scree plot suggested that all 9 eigenvalues provided some explanatory value in the analyses. However, the first four axes accounted for approximately a 50% greater amount of the variance than the next 5 axes. Overall, the amount of variation explained by the first four axes was 60.8% compared to the 92.7% explained by the 9 eigenvalues. Interestingly, there was no one major principal axis that explained a majority of the variation in the data set. This suggests that there is significant variation across the monitoring locations in their grain size characteristics.

### B.2. COMMUNITY PATTERN AND CORRELATIONAL ANALYSES

Multivariate techniques are first used to elucidate community patterns in the biological data. Subsequently, correlation analyses are performed to assist in forming hypotheses as to possible environmental causes of these community patterns. This approach is detailed in Smith et al. (1988) and Clarke and Warwick (2001).

#### B.2.1. DEFINING COMMUNITY PATTERNS

##### B.2.1.1. Spatial Patterns of Biological Communities

The spatial patterns of the biotic communities in the survey area are explored with ordination and cluster analysis. Ordination and cluster methods are used to distinguish groups of entities (e.g. stations) according to similarity or dissimilarity of attributes (e.g. species) (Tetra Tech 1982). The Bray-Curtis index of dissimilarity (Bray and Curtis 1957) is used to compute a similarity matrix between all possible station pairs. The Bray-Curtis similarity index is based on the abundance and distribution of species within and between samples and ranges from 0.0, with complete similarity, to 1.0, with complete dissimilarity

<sup>1</sup> Deceased 2005.



between stations. The index is calculated as:

$$S_{jk} = 100 \left[ 1 - \frac{\sum_i |X_{ij} - X_{ik}|}{\sum_i (X_{ij} + X_{ik})} \right]$$

where  $X_{ij}$  is the abundance of species  $i$  at station  $j$ , and  $X_{ik}$  is the abundance of species  $i$  at station  $k$ . The species abundances are transformed by a square root prior to computation of  $S_{jk}$ . In some cases, dissimilarity indices are calculated to find differences in communities using the following formula:  $D_{jk} = 100 - (S_{jk} * 100)$ .

#### B.2.1.2. Ordination Analysis

Non-metric Multi-Dimensional Scaling (NMDS) analysis using Bray-Curtis similarity values described above was used for the benthic and trawl data (Clarke and Gorley 2006). NMDS does not use axes, but constructs a configuration of samples in a specified number of dimensions based on rank similarities (or dissimilarities). Individual abiotic variables can be tested against the distribution of biotic data as described below in Section B.2.1.3.1.

#### B.2.1.3. Cluster Analysis

Cluster analysis defines groups of stations with similar species composition and abundance. The results are displayed in a hierarchical tree-like structure called a dendrogram. On the dendrogram, two groups are first defined, and within these groups subgroups are defined. Subsequently, subgroups within the subgroups are defined. This process is continued until all stations are a separate subgroup. The hierarchical nature of the dendrogram allows the analyst to choose groups of stations that represent a scale of relevant community differences.

Cluster analysis is also used to define groups of species that tend to have similar distributional patterns among stations.

##### B.2.1.3.1. SIMPER program

The composition of the cluster groups of samples was explored using similarity percentage analysis (SIMPER; Clarke 1993, Clarke and Gorley 2006). SIMPER, a module of the Primer v6 software, identifies species that account for the Bray-Curtis dissimilarities observed within and between groups

of samples. The SIMPER routine computes mean Bray-Curtis similarity values found within user-defined groups of samples as well as between-groups dissimilarity. The SIMPER routine also ascribes the relative contribution of each species to within-group similarity, and between-groups dissimilarity, as well as reporting the average abundance and variability of each species within each group.

#### B.2.2. CORRELATIONS BETWEEN BIOLOGICAL COMMUNITY PATTERNS AND ENVIRONMENTAL PATTERNS

To evaluate possible causes of community patterns, as defined by the ordination and cluster analyses, further analyses correlating community patterns with sediment chemistry and size measurements are performed.

#### B.2.3. COMMUNITY PARAMETERS

The conceptual model of Pearson and Rosenberg (1978) predicts patterns of change for certain community parameters in response to organic enrichment. The following community parameters are computed for comparison with the model predictions.

Community parameters of total abundance and three diversity indices are computed. The Shannon-Weiner diversity index ( $H'$ ) is computed as

$$H' = -\sum_{i=1}^s p_i \cdot \ln(p_i)$$

where  $p_i$  is the proportion of species  $i$  in the sample,  $s$  is the number of species, and  $\ln$  is the natural logarithm (Pielou 1969). Shannon and Weaver (1949) devised this formula to represent the amount of information present per symbol of a code composed of  $s$  discrete symbols whose individual probabilities are  $p_i$ . In an ecological context,  $H'$  measures the diversity per individual in a multi-species community, and is a measure of the uncertainty associated with knowing the species of a randomly drawn individual from the community. This uncertainty is greater when there are more species present and when individuals are distributed more evenly among the species. Thus,  $H'$  is sensitive to both the richness and the evenness components of diversity.

An index of evenness ( $J'$ ) diversity is computed as

$$J' = \frac{H'}{\ln(s)}$$

When individuals are completely evenly distributed among the species,  $H'$  is equal to  $\ln(s)$ . Thus, the ratio  $J'$  indicates how close the community is to this maximum evenness value, with a  $J'$  value of 1 representing perfect evenness.

#### B.2.4. REFERENCE ENVELOPE

To investigate whether the outfall has caused a change in any community metric for either the benthic infauna or the demersal fish, several metrics were calculated for each biological endpoint. The comparisons of interest were each community measure from stations near the outfall with the same community measure from the reference stations. Traditional inferential statistics would apply a test such as an analysis of variance to these data but an alternative approach is to calculate tolerance envelope values for the reference stations and determine if there were systematic exceedances of the outfall stations relative to the tolerance envelope values (Smith 1995, 2001).

The analysis assumes that the reference sites are a random sample from all possible reference locations. Any metric calculated from this sample represents the distribution of values of that metric calculated from the entire population. The comparison of interest is the metric from the sample collected from the outfall population with the distribution of values from the reference sites.

The specific comparison of interest is the percentiles in the outer extremes of the reference distribution. If the value of the metric from an outfall site exceeds an extreme percentile of the reference distribution, for example the 90<sup>th</sup> percentile, it is likely that the outfall has affected the metric at the outfall sample location, i.e., it is unlikely that the value of the metric from the outfall site is from the same distribution as the reference sites.

The tolerance interval bounds are confidence intervals for a percentile of the reference distribution. In this case, the null hypothesis is that the outfall and the reference samples are collected from the same population, i.e. the outfall sites are no different from the reference sites with respect to the measured parameters. The alternative hypothesis is that the outfall sites are different from the reference sites with respect to the measured parameters. The difference could be a result of either the outfall sites having a much larger value of the parameter of interest (greater

than the 90<sup>th</sup> percentile) relative to the reference sites, or the outfall sites having a much smaller value of the parameter of interest (less than the 10<sup>th</sup> percentile) relative to the reference sites.

To complete the analyses, it is necessary to determine the percentile of interest  $P$ , i.e. the percentile value of the distribution that is considered to be an extreme value. Values too near the mean, e.g. the 60<sup>th</sup> percentile, increase the risk of concluding that the outfall has caused an affect on the metrics at the outfall sites. Values distant from the mean, e.g. the 99<sup>th</sup> percentile, increase the risk that any real effect of the outfall on the fauna would be missed. The 90<sup>th</sup> percentile was selected as the upper extreme value and the 10<sup>th</sup> percentile was selected as the lower extreme value. These percentiles were subjectively chosen to balance between environmental protection (sensitivity to impacts) and avoidance of false indications of impact (Smith 1998). It is also necessary to assign a value  $\alpha$ , for the confidence estimate. In this analysis,  $\alpha$  was selected as 0.05 which provides a 95% confidence interval for the percentile. The tolerance interval is the confidence interval calculated for the percentile of interest. In this analysis, a 95% tolerance interval was calculated for the 90<sup>th</sup> percentile and the 10<sup>th</sup> percentile of the reference distribution.

The tolerance interval is calculated differently depending on whether the reference sample is normally distributed or not. If the reference sample is not normally distributed, the data could be transformed to bring them into normality, or a non-parametric method used for the calculation. To facilitate interpretation of the data, the data were not transformed and the non-parametric method for calculating the tolerance envelope was used. A second assumption of the analysis is that the samples are independent. Samples collected yearly at the same location may not be independent. Autocorrelation analyses were used to determine if the samples from consecutive years were similar. The autocorrelation analysis calculates a correlation coefficient between values of any metric of each year and the following year (lag of one year), for each year and two years later (lag of two years), for each year and three years later (lag of three years), etc. If there are 20 years of data, the one year lag correlation value is calculated from the 19 consecutive year pairs, the two year lag correlation is calculated from the 18 two-year pairs, etc.

The data were checked for normality using the

Shapiro-Wilk W test statistic. The Shapiro-Wilk test is one of the most commonly used statistics to test the null hypothesis that the sample data are taken from a normally distributed population. It is not a definitive test and there exist additional tests for normality. However, the Shapiro-Wilk test is generally considered to be one of the best tests for small to medium sized samples.

Fish data (1982-2008) were analyzed by autocorrelation analysis to determine if there were any significant correlations between adjacent years. A significant correlation would indicate a lack of independence in the fish fauna across years. None of the variables had any significant autocorrelations at any time lag indicating independence across years in the samples of the fish fauna. Also, there was no indication of any significant periodicity in the data that could violate the assumptions of independence of samples.

All data were analyzed for normality. For those data that were determined to be normal, the 90% upper and lower tolerance envelope values were generated using the method described in Hahn and Meeker (1991). Briefly, for normally distributed data, the two sided tolerance interval bound was generated using the formula

$$b = \bar{x} \pm g_{(1-\alpha/2, p; n)} s$$

Where  $b$  are the upper and lower tolerance bounds,  $\bar{x}$  is the mean,  $s$  is the standard deviation of the sample,  $p$  is the percentile of interest,  $g$  is the constant appropriate for the value of  $\alpha/2$  or  $\alpha$  and selected confidence interval of the percentile which is calculated as  $100(1 - \alpha/2)$  (or  $100 - \alpha$ ). The value of  $\alpha/2$  is appropriate for the case where both the upper and lower tolerance envelope values are desired and the probability of an exceedance, in either direction, is set at the level  $\alpha$ . For this analysis, the confidence intervals were selected to be 90%. For normally distributed samples, values of  $k$  were taken from Table A.11b in Hahn and Meeker (1991).

For variables that were not normally distributed, non-parametric tolerance envelope values were calculated according to the method described in Chapter 5 of Hahn and Meeker (1991). Briefly, the values are ordered from smallest to largest and the number of extreme observations from each end of the distribution are found in Table A.16 of Hahn and Meeker (1991) that enclose 90% of the population at

an  $\alpha$  of 0.05. The number of values is matched to the numeric value in the sequence of numbers and those values become the upper and lower tolerance values.

### B.3. BIOACCUMULATION

Three analyses were performed: 1) determine if there are trends in the concentrations of metals in fish and crab tissues at the reference and outfall sites over time, 2) determine if there are differences in the concentrations of each metal from individual tissues between reference and outfall sites, and 3) determine if there are trends in the concentrations of metals in tissues at the reference and outfall sites (bioaccumulation of metals) over time.

The first analysis was performed using a linear regression of concentration against time. The slope of the regression line provides a measure of trend, and the significance of the regression slope indicates whether the trend is statistically significant. The null hypothesis for the regression slope analysis is that the slope is equal to 0 ( $H_0: b_1 = 0$ ), and the alternative hypothesis is that the slope is unequal to 0 ( $H_A: b_1 <> 0$ ). The residuals of all regressions were examined to determine if the assumption of linearity was reasonable, or alternatively, if the residuals indicated that a nonlinear model would be a better fit. In no instances were the residuals suggestive of a nonlinear fit indicating a linear regression was an adequate model for the data.

The second analysis was performed using either a one-tailed paired t-test with unequal variances or a Mann-Whitney U test for differences in the median concentration. Prior to performing each of these analyses, all data were checked for normality using a variety of test statistics. If any of the test statistics indicated that the data were normally distributed, the t-test was used. If any of the test statistics indicated the data were non-normal, the Mann-Whitney test was performed to determine if there were differences in concentration between reference and outfall sites.

The third analysis was performed using a linear regression of concentration in the tissue against concentration in the sediment. Again, the slope of the regression line provides a measure of trend and the significance of the regression slope indicates whether the trend is statistically significant. The null hypothesis is that there is no relationship between the concentration in the tissue and the concentration in the sediment ( $H_0: b_1 = 0$ ), and the alternative is that there is a significant relationship between sediment

and tissue concentration ( $H_A: b_1 \neq 0$ ). As in the first analysis, the alternative is a two-tailed test which means the tissue concentration could be either lesser or greater than the sediment concentration of any metal, i.e. as the concentration of the metal in the sediment increases, the concentration of the metal in the tissue could either increase or decrease. A slope of 1.0 indicates that the concentration of metal in the tissue and sediment increase at the same rate. A slope different from 1.0 indicates that the rate of increase in the tissue concentration is either higher or lower than the increase in the concentration of the metal in the sediment.

In all of these analyses, the large number of metals, organisms, and tissues, results in an equally large number of tests performed. In this case, there is a danger of inflating the Type I error rate when evaluating the hypotheses. When performing 100 tests for differences in the concentration of metals between two locations (e.g. reference and outfall) using an  $\alpha = 0.05$  indicates that there is a 5% chance that one will incorrectly reject the null hypothesis of no differences between sites and accept the alternative hypothesis that the sites differ in the concentration of the metals when no difference exists. In the second analysis, 52 tests (either one-tailed t-tests or Mann-Whitney U tests) were performed to determine if the concentration of metals differed between the reference and outfall sites. In this case, with  $\alpha = 0.05$ , 2.6 (3) tests would be expected to falsely reject the null hypothesis.

There are two general methods to correct this problem that are commonly used in statistical analyses: Bonferroni corrections, or the use of False Discovery Rate techniques. The Bonferroni corrections involve dividing the  $\alpha$  value by the number of tests either at the beginning of the evaluation, or sequentially during the evaluation of the hypotheses. In the case of the second analysis above, the  $\alpha = 0.05$  value would be divided by either 104 at the beginning of the analysis making  $\alpha = 0.00048$  the p value necessary for a statistically significant difference in each individual test (Bonferroni correction), or divided sequentially by the number of significant tests (sequential Bonferroni). Consequently, for the first significance test, the value of  $\alpha = 0.05$  (p value) is used. For the second significance test,  $\alpha = 0.025$  (p value) is necessary to reject the null hypothesis, for the third significance test  $\alpha = 0.0125$  is necessary to reject the null hypothesis, and so on.

Both the Bonferroni and sequential Bonferroni

corrections are generally considered too conservative and result in very few statistically significant differences. The alternative is to use recently developed False Discovery Rate (FDR) corrections. These techniques were developed in response to performing thousands of significance tests on data from molecular analyses and generally do not perform well when the number of tests is small. Fortunately, there was only one statistically significant difference in concentration between the reference and outfall sites (mercury in fish muscle, see below). A correction for the number of tests was not necessary in this analysis although it should be noted that given the large number of significance tests performed, there is the possibility that the significant difference in concentration of mercury in fish muscle does not exist. However, the formal analysis did not correct for the number of tests performed. The same caveat is true for the first and third analyses which also involved a large number of significance tests (tests of the slope of the regression line being equal to 0). There were a larger number of significant tests but neither the overly conservative Bonferroni corrections nor the FDR corrections for extremely large numbers of tests were applied. In these cases, the statistical test results should be viewed in the context of the biological and chemical mechanisms involved in the deposition and bioaccumulation of these metals.

#### **B.4. BACIP ANALYSIS**

The general BACIP (Before-After-Control-Impact Paired) experimental design involves sampling at predetermined “control” and “impact” areas before and after the onset of the potentially impacting activity (Bernstein and Zalinski 1983, Stewart-Oaten et al. 1986). A change in indicator values at a potentially impacted location after the onset of the impacting activity does not necessarily indicate that an impact has occurred, since indicator values can change naturally over time. With this statistical design, it is assumed that large-scale environmental factors causing natural temporal changes in indicator values will have a similar effect in both the impact and control areas. Thus, the test for impact is a test for changes in the after-impact period that do *not* take place in both the control and impact areas. The null hypothesis of the BACIP statistical test is that the average differences between impact and control sites will be the same in the before- and after-operational periods. The details of the sampling design can vary,

although all credible designs should involve multiple sampling periods both before and after the impact. As a paired test, the comparison involves a single impact and a single control location. This is because a point source impact (such as an outfall) will create gradients of change in the vicinity of the impact, and the severity of the impact at different locations on the gradient is of interest rather than the impact to the larger area.

An assumption of the BACIP test is that the differences within each group are normally distributed. When the data are positively skewed as total abundances usually are, log transformation will make these differences more normal. Using the log is equivalent to testing for the same ratio of abundances before and after impact. Student's T test was used to evaluate the differences in the log values prior to discharge and after discharge.

The City and County of San Francisco began pre-discharge benthic infauna studies in 1982. Since that time, one outfall station (station 01) and one reference station (station 06) consistently remained part of the sampling program. A BACIP analysis of infauna abundance at these two stations was performed to provide some information on the degree to which the wastewater discharge may have affected organism abundance.

## **B.5. REFERENCES CITED**

- Bray, J. R. and J. T. Curtis. 1957. An ordination of the upland forest communities of southern Wisconsin. *Ecol. Monogr.* 27:325-349.
- Clarke KR. 1993. Non-parametric multivariate analysis of changes in community structure. *Australian Journal of Ecology.* 18:117-143.
- Clarke, K. R. and R. N. Gorley. 2006. *Primer v6: User Manual / Tutorial.* Primer-E Ltd, Plymouth Marine Laboratory. 1-190.
- Clarke, K. R. and R.M. Warwick. 2001. Change in marine communities: An approach to statistical analysis and interpretation. *Primer-E Ltd, Plymouth Marine Laboratory, UK* 1-180.
- Folk, R.L. and W.C. Ward 1957. Brazos River Bar: a study in the significance of grain size parameters. *Journal of Sedimentary Petrology* 27(1): 3-26.
- Hahn, G.J. and W.Q. Meeker. 1991. *Statistical Intervals. A Guide for Practitioners.* A Wiley-Interscience Publication. John Wiley & Sons, Inc. New York. 392 pp.
- Pearson, T.H. and R. Rosenberg. 1978. Macrobenthic succession in relation to organic enrichment and pollution of the marine environment. *Oceanogr. Mar. Biol. Ann. Rev.* 16: 229-311.
- Pielou, E.C. 1969. *An Introduction to Mathematical Ecology.* Wiley-Interscience. New York: 286 pp.
- Shannon, C.E. and W. Weaver. 1949. *The Mathematical Theory of Communication.* University of Illinois Press, Urbana.
- Smith, R.W. 1995. The reference envelope approach to impact monitoring. A report to U.S. EPA, Region IX. Grant # X-009904-01-0.
- Smith, R.W. 1998. Analysis of 1994 Orange County outfall benthic data in a regional context. Submitted to County Sanitation Districts of Orange County, Fountain Valley, CA; 71 pp.
- Smith, R.W. 2001. The use of random-model tolerance intervals in environmental monitoring and regulation. *Journal of Agricultural, Biological, and Environmental Statistics.* 7(1): 74-94.
- Smith, R.W., B.B. Bernstein, and R.L. Cimberg. 1988. Community-Environmental Relationships in the Benthos: Applications of Multivariate Analytical Techniques. Chapter 11 In: *Marine Organisms as Indicators.* Springer-Verlag. New York: 247-326.

A black and white photograph of a person wearing a wetsuit and carrying gear, kneeling on a sandy beach near the water's edge. The person appears to be engaged in a monitoring or sampling activity. The background shows gentle waves washing onto the shore.

**APPENDIX C**

**BEACH  
MONITORING  
PROGRAM**

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Beach Monitoring Program  
1997 - 2013

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Appendix C-1a  
*Total coliform bacteria (MPN/100 mL) at shoreline stations  
 July 2012 - December 2012*

	Shoreline Station									
	<u>15</u>	<u>15E</u>	<u>16</u>	<u>17</u>	<u>18</u>	<u>19</u>	<u>20</u>	<u>21</u>	<u>21.1</u>	<u>22</u>
2-Jul-12	73	31	31	31	< 10	10			< 10	
9-Jul-12	637	41	20	20	< 10	20			1333	
10-Jul-12									< 10	
16-Jul-12	8164	20	20	73	< 10	31			20	
17-Jul-12	1112	107								
23-Jul-12	794	75	< 10	20	41	< 10			10	
30-Jul-12	2310	10	< 10	20	< 10	10			31	
6-Aug-12	< 10	457	41	< 10	10	< 10			10	
13-Aug-12	620	52	160	30	< 10	< 10			10	
20-Aug-12	3448	20	52	10	20	20			10	
27-Aug-12	1616	< 10	63	41	< 10	< 10			< 10	
4-Sep-12	435	3076	41	20	< 10	< 10			< 10	
5-Sep-12	1968	10								
7-Sep-12	<b>19863</b>	175								
8-Sep-12	6488	265								
10-Sep-12	9804	20	20	10	< 10	20			< 10	
17-Sep-12	3169	63	20	10	31	< 10			< 10	
24-Sep-12	571	183	285	20	20	10			10	
25-Sep-12	31	134	20							
1-Oct-12	1182	10	85	148	31	10			20	
9-Oct-12	703	< 10	41	10	< 10	< 10			< 10	
15-Oct-12	520	134	<b>12033</b>	328	345	41			771	
16-Oct-12	1259	41	1187		10				122	
17-Oct-12			3654						10	
22-Oct-12	3441	620	108	175	789	563			393	
23-Oct-12					63					
29-Oct-12	<b>10462</b>	< 10	31	< 10	31	< 10			10	
30-Oct-12	20	< 10								
5-Nov-12	9804	20	10	< 10	31	10			< 10	
13-Nov-12	10	2909	41	74	< 10	97			10	
19-Nov-12	2603	121	41	31	52	20			20	
21-Nov-12					<b>&gt; 24196</b>	4611	345	<b>&gt; 24196</b>	2359	6131
22-Nov-12					203	85		109		988
23-Nov-12					63					
24-Nov-12					52					
25-Nov-12					41					
26-Nov-12	20	41	< 10	20	10	< 10			20	
28-Nov-12	5172	262	216	41						
29-Nov-12	4352	148	3255		121					
30-Nov-12	1722	1376	3255		6131	<b>&gt; 24196</b>	3255	<b>&gt; 24196</b>	2755	591
1-Dec-12	<b>&gt; 24196</b>	275	282		520	323	318	529	404	
2-Dec-12	4884	4352	5475	4884	<b>&gt; 24196</b>		1483	<b>&gt; 24196</b>	<b>24196</b>	1529
3-Dec-12	404	201	216	175	121	2247		384	97	
5-Dec-12					<b>&gt; 24196</b>	<b>&gt; 24196</b>	2247	7701	613	<b>&gt; 24196</b>
6-Dec-12					74	3873		288		504
7-Dec-12						171				
10-Dec-12	2809	10	85	31	41	1565			10	
17-Dec-12	4352	121	512	75	41	85			31	
22-Dec-12	<b>&gt; 24196</b>	31	10							
23-Dec-12	5172	298								
24-Dec-12					86	1376	97	269	75	98
25-Dec-12						160				
26-Dec-12	<b>19863</b>	373	480	908	637	798	839	991	638	4884
27-Dec-12	235	246				179				201

**BOLD = > 10,000 MPN/100mL**

Appendix C-1b  
*Total coliform bacteria (MPN/100 mL) at shoreline stations*  
 January 2013 - June 2013

	Shoreline Station									
	<u>15</u>	<u>15E</u>	<u>16</u>	<u>17</u>	<u>18</u>	<u>19</u>	<u>20</u>	<u>21</u>	<u>21.1</u>	<u>22</u>
2-Jan-13	538	< 10	52	31	< 10	10		< 10		
7-Jan-13	2400	74	31	31	120	199		373		
14-Jan-13	657	< 10	< 10	< 10	10	20		< 10		
22-Jan-13	8164	10	31	10	< 10	< 10		< 10		
28-Jan-13	5475	20	10	10	< 10	< 10		10		
4-Feb-13	1246	63	< 10	< 10	< 10	10		< 10		
11-Feb-13	74	121	31	31	52	52		< 10		
19-Feb-13	246	144	< 10	20	20	20		< 10		
25-Feb-13	1354	10	20	20	10	< 10		< 10		
4-Mar-13	683	10	10	10	31	< 10		63		
11-Mar-13	5172	85	62	20	< 10	< 10		30		
18-Mar-13	173	< 10	20	10	10	10		< 10		
25-Mar-13	98	31	171	< 10	< 10	10		< 10		
1-Apr-13	6131	110	41	52	< 10	20		146		
2-Apr-13	583	41								
8-Apr-13	3654	31	10	10	10	< 10		< 10		
15-Apr-13	4611	< 10	52	10	< 10	41		< 10		
22-Apr-13	1935	< 10	10		10	< 10		< 10		
23-Apr-13				< 10						
29-Apr-13	1201	20	< 10	< 10	< 10	31		< 10		
6-May-13	5794	41	< 10	< 10	< 10	< 10		< 10		
13-May-13	315	97	52	< 10	10	< 10		< 10		
20-May-13	4884	20	< 10	10	< 10	10		< 10		
28-May-13	1664	41	10	< 10	10	10		10		
29-May-13	2187	30								
3-Jun-13	97	31	41	< 10	< 10	< 10		< 10		
10-Jun-13	3255	743	30	10	< 10	< 10		41		
17-Jun-13	3076	97	< 10	10	10	< 10		< 10		
24-Jun-13	1281	52	31	20	20	< 10		< 10		

**BOLD => 10,000 MPN/100mL**

Appendix C-1c  
*Escherichia Coli (MPN/100 mL) at shoreline stations*  
 July 2012 - December 2012

	Shoreline Station									
	<u>15</u>	<u>15E</u>	<u>16</u>	<u>17</u>	<u>18</u>	<u>19</u>	<u>20</u>	<u>21</u>	<u>21.1</u>	<u>22</u>
2-Jul-12	41	10	< 10	< 10	< 10	< 10			< 10	
9-Jul-12	< 10	< 10	20	< 10	< 10	< 10			<b>538</b>	
10-Jul-12									< 10	
16-Jul-12	<b>738</b>	10	10	31	< 10	31			< 10	
17-Jul-12	41	< 10								
23-Jul-12	20	< 10	< 10	10	31	< 10			10	
30-Jul-12	20	< 10	< 10	< 10	< 10	< 10			20	
6-Aug-12	< 10	31	< 10	< 10	< 10	< 10			10	
13-Aug-12	30	52	20	< 10	< 10	< 10			< 10	
20-Aug-12	10	< 10	10	< 10	10	20			10	
27-Aug-12	< 10	< 10	10	< 10	< 10	< 10			< 10	
4-Sep-12	10	< 10	< 10	10	< 10	< 10			< 10	
5-Sep-12	31	< 10								
7-Sep-12	<b>4352</b>	41								
8-Sep-12	63	10								
10-Sep-12	63	< 10	10	< 10	< 10	< 10			< 10	
17-Sep-12	10	10	< 10	< 10	< 10	< 10			< 10	
24-Sep-12	52	75	108	< 10	10	10			< 10	
25-Sep-12	10	< 10	10							
1-Oct-12	30	10	< 10	110	< 10	< 10			10	
9-Oct-12	< 10	< 10	< 10	< 10	< 10	< 10			< 10	
15-Oct-12	250	86	148	195	213	10			<b>520</b>	
16-Oct-12	31	20	<b>1187</b>		< 10				63	
17-Oct-12			< 10						< 10	
22-Oct-12	75	30	31	10	<b>538</b>	31			20	
23-Oct-12					< 10					
29-Oct-12	10	< 10	< 10	< 10	< 10	< 10			10	
30-Oct-12	< 10	< 10								
5-Nov-12	52	< 10	< 10	< 10	< 10	< 10			< 10	
13-Nov-12	< 10	134	20	31	10	30			< 10	
19-Nov-12	10	31	< 10	< 10	10	10			< 10	
21-Nov-12					<b>&gt; 24196</b>	<b>586</b>	75	<b>2098</b>	203	<b>464</b>
22-Nov-12					20	10		30		20
23-Nov-12					20					
24-Nov-12					20					
25-Nov-12					10					
26-Nov-12	< 10	41	< 10	< 10	< 10	< 10			10	
28-Nov-12	175	41	98	< 10						
29-Nov-12	298	86	<b>727</b>		10					
30-Nov-12	<b>521</b>	345	<b>1153</b>		<b>2282</b>	<b>24196</b>	<b>1210</b>	<b>&gt; 24196</b>	<b>583</b>	160
1-Dec-12	97	109	20		134	20	20	85	31	
2-Dec-12	<b>990</b>	<b>988</b>	<b>886</b>	<b>1396</b>	<b>4106</b>		383	<b>6867</b>	<b>4884</b>	216
3-Dec-12	173	10	10	31	30	132		< 10	< 10	
5-Dec-12					<b>4106</b>	<b>24196</b>	379	<b>703</b>	41	<b>3448</b>
6-Dec-12					< 10	<b>565</b>		63		52
7-Dec-12						52				
10-Dec-12	< 10	< 10	31	10	10	161			10	
17-Dec-12	41	< 10	20	< 10	10	< 10			< 10	
22-Dec-12	72	< 10	< 10							
23-Dec-12	< 10	< 10								
24-Dec-12					52	282	10	63	20	20
25-Dec-12						10				
26-Dec-12	201	63	41	191	134	185	75	175	85	<b>563</b>
27-Dec-12	20	< 10				30				< 10

**BOLD = > 400 MPN/100mL**

Appendix C-1d  
*Escherichia coli* (MPN/100 mL) at shoreline stations  
 January 2013 to June 2013

	Shoreline Station									
	<u>15</u>	<u>15E</u>	<u>16</u>	<u>17</u>	<u>18</u>	<u>19</u>	<u>20</u>	<u>21</u>	<u>21.1</u>	<u>22</u>
2-Jan-13	< 10	< 10	< 10	10	< 10	10				< 10
7-Jan-13	31	10	10	10	30	85				120
14-Jan-13	20	< 10	< 10	< 10	< 10	< 10				< 10
22-Jan-13	86	< 10	< 10	< 10	< 10	< 10				< 10
28-Jan-13	< 10	< 10	< 10	< 10	< 10	< 10				< 10
4-Feb-13	10	< 10	< 10	< 10	< 10	< 10				< 10
11-Feb-13	10	10	< 10	< 10	31	< 10				< 10
19-Feb-13	< 10	10	< 10	< 10	10	20				< 10
25-Feb-13	20	< 10	< 10	10	< 10	< 10				< 10
4-Mar-13	< 10	< 10	< 10	10	20	< 10				< 10
11-Mar-13	52	10	10	10	< 10	< 10				30
18-Mar-13	< 10	< 10	10	< 10	< 10	< 10				< 10
25-Mar-13	10	31	132	< 10	< 10	< 10				< 10
1-Apr-13	146	< 10	< 10	< 10	< 10	10				< 10
2-Apr-13	< 10	10								
8-Apr-13	41	< 10	< 10	10	< 10	< 10				< 10
15-Apr-13	41	< 10	< 10	< 10	< 10	< 10				< 10
22-Apr-13	10	< 10	< 10		10	< 10				10
23-Apr-13				< 10						
29-Apr-13	< 10	< 10	< 10	< 10	< 10	20				< 10
6-May-13	20	< 10	< 10	< 10	< 10	< 10				< 10
13-May-13	< 10	< 10	52	< 10	< 10	< 10				< 10
20-May-13	20	< 10	< 10	< 10	< 10	< 10				< 10
28-May-13	122	10	< 10	< 10	10	< 10				< 10
29-May-13	20	< 10								
3-Jun-13	< 10	< 10	< 10	< 10	< 10	< 10				< 10
10-Jun-13	31	31	10	10	< 10	< 10				10
17-Jun-13	41	10	< 10	< 10	< 10	< 10				< 10
24-Jun-13	< 10	10	10	10	10	< 10				< 10

#

**BOLD => 400 MPN/100 ml**

Appendix C-1e  
*Enterococcus (MPN/100 mL) at shoreline stations*  
 July 2012 - December 2012

	Shoreline Station									
	<u>15</u>	<u>15E</u>	<u>16</u>	<u>17</u>	<u>18</u>	<u>19</u>	<u>20</u>	<u>21</u>	<u>21.1</u>	<u>22</u>
2-Jul-12	86	< 10	< 10	10	< 10	< 10			< 10	
9-Jul-12	< 10	< 10	< 10	< 10	< 10	< 10			74	
10-Jul-12									< 10	
16-Jul-12	<b>146</b>	< 10	< 10	20	< 10	20			10	
17-Jul-12	20	< 10								
23-Jul-12	< 10	< 10	10	< 10	< 10	< 10			< 10	
30-Jul-12	10	< 10	< 10	< 10	< 10	10			< 10	
6-Aug-12	< 10	< 10	< 10	< 10	< 10	< 10			20	
13-Aug-12	10	< 10	< 10	< 10	10	< 10			< 10	
20-Aug-12	97	10	< 10	< 10	< 10	< 10			< 10	
27-Aug-12	20	< 10	< 10	< 10	< 10	< 10			< 10	
4-Sep-12	98	<b>223</b>	< 10	< 10	< 10	< 10			< 10	
5-Sep-12	20	<b>199</b>								
7-Sep-12	<b>2851</b>	10								
8-Sep-12	41	10								
10-Sep-12	96	< 10	< 10	< 10	< 10	< 10			< 10	
17-Sep-12	20	< 10	< 10	20	< 10	< 10			< 10	
24-Sep-12	<b>146</b>	75	<b>441</b>	30	< 10	< 10			62	
25-Sep-12	10	20	< 10							
1-Oct-12	10	< 10	< 10	10	< 10	< 10			< 10	
9-Oct-12	10	< 10	< 10	< 10	< 10	< 10			< 10	
15-Oct-12	<b>134</b>	20	98	74	<b>262</b>	31			86	
16-Oct-12	< 10	30	<b>121</b>		< 10				<b>134</b>	
17-Oct-12			< 10						31	
22-Oct-12	41	< 10	10	20	20	41			20	
23-Oct-12					< 10					
29-Oct-12	41	10	20	20	< 10	20			< 10	
30-Oct-12	< 10	10								
5-Nov-12	41	< 10	< 10	10	< 10	< 10			< 10	
13-Nov-12	< 10	10	< 10	41	20	75			< 10	
19-Nov-12	10	10	10	< 10	10	10			< 10	
21-Nov-12					<b>345</b>	<b>259</b>	20	<b>512</b>	63	<b>323</b>
22-Nov-12					< 10	31		31		20
23-Nov-12					< 10					
24-Nov-12					10					
25-Nov-12					10					
26-Nov-12	< 10	10	< 10	10	< 10	< 10			< 10	
28-Nov-12	<b>173</b>	20	10	20						
29-Nov-12	<b>441</b>	<b>109</b>	<b>3873</b>		74					
30-Nov-12	<b>1396</b>	<b>1019</b>	<b>4106</b>		<b>723</b>	<b>&gt; 24196</b>	<b>884</b>	<b>12997</b>	<b>369</b>	<b>120</b>
1-Dec-12	<b>275</b>	<b>331</b>	<b>173</b>		<b>464</b>	63	63	41	<b>350</b>	
2-Dec-12	<b>213</b>	<b>228</b>	<b>146</b>	<b>393</b>	<b>908</b>		85	<b>1050</b>	<b>1553</b>	86
3-Dec-12	10	10	10	20	< 10	31		< 10	< 10	
5-Dec-12					<b>556</b>	<b>6867</b>	96	<b>155</b>	63	<b>2143</b>
6-Dec-12					10	<b>110</b>		< 10		10
7-Dec-12						31				
10-Dec-12	41	< 10	20	< 10	< 10	10			< 10	
17-Dec-12	52	< 10	10	< 10	10	< 10			< 10	
22-Dec-12	<b>160</b>	31	74							
23-Dec-12	30	10								
24-Dec-12					< 10	<b>135</b>	10	< 10	< 10	< 10
25-Dec-12						< 10				
26-Dec-12	85	41	31	20	10	<b>106</b>	63	63	98	<b>110</b>
27-Dec-12	< 10	< 10				20				< 10

**BOLD = > 104 MPN/100mL**

Appendix C-1f  
*Enterococcus (MPN/100 mL) at shoreline stations*  
 January 2013 to June 2013

	Shoreline Station									
	<u>15</u>	<u>15E</u>	<u>16</u>	<u>17</u>	<u>18</u>	<u>19</u>	<u>20</u>	<u>21</u>	<u>21.1</u>	<u>22</u>
2-Jan-13	< 10	< 10	< 10	41	10	< 10			< 10	
7-Jan-13	20	< 10	10	< 10	10	< 10			31	
14-Jan-13	20	< 10	< 10	< 10	< 10	< 10			< 10	
22-Jan-13	10	10	< 10	< 10	10	< 10			< 10	
28-Jan-13	< 10	< 10	< 10	< 10	< 10	< 10			< 10	
4-Feb-13	20	20	< 10	< 10	< 10	< 10			< 10	
11-Feb-13	74	41	20	< 10	< 10	< 10			< 10	
19-Feb-13	< 10	< 10	< 10	< 10	< 10	< 10			< 10	
25-Feb-13	95	< 10	< 10	< 10	< 10	< 10			< 10	
4-Mar-13	10	< 10	< 10	< 10	< 10	< 10			< 10	
11-Mar-13	75	20	< 10	< 10	< 10	< 10			41	
18-Mar-13	< 10	< 10	< 10	< 10	10	< 10			< 10	
25-Mar-13	63	< 10	52	< 10	10	< 10			< 10	
1-Apr-13	<b>185</b>	10	< 10	10	< 10	< 10			10	
2-Apr-13	10	< 10								
8-Apr-13	74	< 10	< 10	< 10	< 10	10			< 10	
15-Apr-13	63	< 10	< 10	< 10	10	< 10			< 10	
22-Apr-13	< 10	10	< 10		< 10	< 10			< 10	
23-Apr-13				20						
29-Apr-13	10	< 10	< 10	10	< 10	< 10			< 10	
6-May-13	41	< 10	< 10	< 10	< 10	< 10			< 10	
13-May-13	< 10	< 10	< 10	< 10	< 10	< 10			< 10	
20-May-13	31	< 10	< 10	< 10	< 10	< 10			< 10	
28-May-13	<b>173</b>	< 10	< 10	20	10	< 10			< 10	
29-May-13	10	< 10								
3-Jun-13	< 10	10	< 10	< 10	< 10	< 10			< 10	
10-Jun-13	41	10	< 10	< 10	< 10	10			< 10	
17-Jun-13	74	20	< 10	< 10	< 10	< 10			< 10	
24-Jun-13	30	< 10	< 10	< 10	< 10	< 10			< 10	

**BOLD => 104 MPN/100 ml**

APPENDIX C-2a  
*Bacteria counts greater than single sample maximum limit*  
 July 2008 - June 2009

Sample Date	Station	Total Coliform Count > 10,000 MPN/100mL	<i>E. coli</i> Count > 400 MPN/100mL	Enterococcus Count > 104 MPN/100mL	Follow-up sample date	Total Coliform Count	<i>E. coli</i> Count	Enterococcus Count
29-Jul-08	15	5794	86	<b>185</b>	30-Jul-08	4106	86	86
6-Aug-08	15	> <b>24196</b>	<b>441</b>	97	7-Aug-08	2098	10	<10
12-Aug-08	15	9804	52	<b>199</b>	13-Aug-08	9208	20	73
20-Aug-08	15	<b>24196</b>	<b>504</b>	<b>638</b>	21-Aug-08	<b>12033</b>	75	74
					22-Aug-08	63	41	20
26-Aug-08	15	> <b>24196</b>	<b>487</b>	<b>462</b>	27-Aug-08	201	10	85
27-Aug-08	15E	<b>14136</b>	<b>583</b>	<b>959</b>	28-Aug-08	10	<10	10
1-Oct-08	15	<b>17329</b>	<b>473</b>	<b>884</b>	2-Oct-08	<b>17329</b>	<b>295</b>	<b>173</b>
					3-Oct-08	<b>12033</b>	110	51
					4-Oct-08	256	<10	<10
1-Oct-08	15E	256	199	<b>203</b>	2-Oct-08	189	86	63
1-Oct-08	16	269	85	<b>142</b>	2-Oct-08	97	31	<10
15-Oct-08	16	7701	107	<b>298</b>	16-Oct-08	10	<10	<10
29-Oct-08	15	> <b>24196</b>	<b>910</b>	<b>670</b>	30-Oct-08	4106	228	52
2-Nov-08	19	1872	262	<b>187</b>	3-Nov-08	120	<10	<10
12-Nov-08	19	3076	<b>613</b>	52	13-Nov-08	10	10	10
18-Nov-08	15	<b>11199</b>	279	<b>131</b>	19-Nov-08	1989	73	63
18-Nov-08	18	<10	<10	<b>199</b>	19-Nov-08	20	10	<10
25-Nov-08	17	546	323	<b>109</b>	26-Nov-08	272	171	40
13-Jan-09	15	2481	226	<b>108</b>	15-Jan-09	20	<10	<10
4-Feb-09	18	932	63	<b>243</b>	5-Feb-09	10	<10	10
15-Feb-09	16	98	10	<b>213</b>	16-Feb-09	146	10	10
16-Feb-09	18	3654	<b>884</b>	<b>243</b>	17-Feb-09	231	41	<10
16-Feb-09	19	<b>12033</b>	<b>2489</b>	<b>637</b>	17-Feb-09	3448	377	98
16-Feb-09	20	2359	<b>733</b>	<b>148</b>	17-Feb-09	275	31	10
16-Feb-09	21	3076	<b>845</b>	<b>230</b>	17-Feb-09	637	41	31
16-Feb-09	21.1	3255	<b>882</b>	<b>185</b>	17-Feb-09	187	<10	10
16-Feb-09	22	<b>17329</b>	<b>2481</b>	<b>2603</b>	17-Feb-09	<b>12033</b>	<b>1664</b>	<b>1860</b>
					18-Feb-09	20	<10	<10
5-Mar-09	21.1	3873	<b>613</b>	<b>121</b>	6-Mar-09	74	20	31
15-Apr-09	15E	1872	84	<b>132</b>	16-Apr-09	10	<10	<10
27-May-09	15	<b>14136</b>	<b>727</b>	<b>985</b>	28-May-09	4884	41	62



APPENDIX C-2b  
*Bacteria counts greater than single sample maximum limit*  
 July 2009 - June 2010

Sample Date	Station	Total Coliform Count > 10,000 MPN/100mL	<i>E. coli</i> Count > 400 MPN/100mL	Enterococcus Count > 104 MPN/100mL	Follow-up sample date	Total Coliform Count	<i>E. coli</i> Count	Enterococcus Count
8-Jul-09	15	880	201	<b>121</b>	10-Jul-09	3076	10	<b>132</b>
					11-Jul-09	988	<10	<b>393</b>
					12-Jul-09	327	<10	<10
14-Jul-09	15	> <b>24196</b>	<b>441</b>	<b>631</b>	15-Jul-09	8164	74	<b>183</b>
					16-Jul-09	<b>12997</b>	97	<b>1336</b>
					17-Jul-09	2851	52	72
11-Aug-09	15	> <b>24196</b>	279	<b>544</b>	12-Aug-09	<b>15531</b>	75	<b>122</b>
					13-Aug-09	<b>17329</b>	259	<b>909</b>
					14-Aug-09	<b>17329</b>	233	<b>294</b>
					15-Aug-09	1017	10	52
19-Aug-09	15	<b>24196</b>	373	<b>135</b>	20-Aug-09	98	<10	<10
8-Sep-09	15	<b>12033</b>	369	<b>246</b>	9-Sep-09	> <b>24196</b>	185	86
					10-Sep-09	959	<10	20
22-Sep-09	15	> <b>24196</b>	<b>780</b>	<b>886</b>	23-Sep-09	862	10	<10
30-Sep-09	15	<b>10462</b>	158	158	1-Oct-09	<b>24196</b>	158	<b>161</b>
					2-Oct-09	> <b>24196</b>	426	<b>292</b>
					3-Oct-09	<b>3873</b>	<b>3448</b>	<b>175</b>
					4-Oct-09	377	10	52
3-Oct-09	15E	74	41	<b>108</b>	4-Oct-09	31	20	<10
13-Oct-09	15	<b>12997</b>	355	<b>359</b>	14-Oct-09	<b>4611</b>	<b>631</b>	<b>336</b>
					15-Oct-09	246	97	52
13-Oct-09	16	2382	75	<b>305</b>	14-Oct-09	906	122	52
13-Oct-09	18	> <b>24196</b>	> <b>24196</b>	<b>6867</b>	14-Oct-09	364	10	41
13-Oct-09	19	> <b>24196</b>	> <b>24196</b>	> <b>24196</b>	14-Oct-09	4611	<b>496</b>	86
					15-Oct-09	213	74	20
13-Oct-09	20	> <b>24196</b>	<b>14136</b>	<b>4611</b>	14-Oct-09	156	20	52
13-Oct-09	21	> <b>24196</b>	> <b>24196</b>	> <b>24196</b>	14-Oct-09	487	98	52
13-Oct-09	21.1	> <b>24196</b>	> <b>24196</b>	<b>2142</b>	14-Oct-09	281	52	52
14-Oct-09	15E	1860	<b>576</b>	<b>624</b>	15-Oct-09	259	52	10
20-Oct-09	15	6867	<b>471</b>	<b>301</b>	21-Oct-09	<b>24196</b>	<b>512</b>	<b>408</b>
					22-Oct-09	> <b>24196</b>	<b>1607</b>	<b>1145</b>
					23-Oct-09	> <b>24196</b>	108	<b>171</b>
					24-Oct-09	<b>15531</b>	31	85
					25-Oct-09	<b>4884</b>	<b>546</b>	75
					26-Oct-09	8164	41	52
22-Oct-09	15E	> <b>24196</b>	<b>909</b>	> <b>24196</b>	23-Oct-09	538	52	63
3-Nov-09	15	7270	98	<b>121</b>	4-Nov-09	63	31	<10
4-Nov-09	15E	5172	98	<b>108</b>	6-Nov-09	10	<10	<10
10-Nov-09	15	<b>24196</b>	110	<b>187</b>	11-Nov-09	<b>14136</b>	<10	75
					12-Nov-09	372	20	<10
17-Nov-09	15E	457	301	<b>216</b>	18-Nov-09	241	145	<b>373</b>
					19-Nov-09	20	10	<10
6-Jan-10	15	<b>12997</b>	20	52	7-Jan-10	> <b>24196</b>	373	<b>1043</b>
					8-Jan-10	<b>15531</b>	20	52
					9-Jan-10	1396	<10	<10
12-Jan-10	16	3873	63	<b>309</b>	13-Jan-10	173	10	97
12-Jan-10	19	776	213	<b>496</b>	13-Jan-10	31	10	86
12-Jan-10	21.1	216	<10	<b>420</b>	13-Jan-10	52	10	<10
18-Jan-10	15	1145	146	<b>364</b>	19-Jan-10	960	246	98

APPENDIX C-2b (cont.)  
*Bacteria counts greater than single sample maximum limit*  
 July 2009 - June 2010

Sample Date	Station	Total Coliform Count > 10,000 MPN/100mL	<i>E. coli</i> Count > 400 MPN/100mL	Enterococcus Count > 104 MPN/100mL	Follow-up sample date	Total Coliform Count	<i>E. coli</i> Count	Enterococcus Count
18-Jan-10	15E	563	120	<b>216</b>	19-Jan-10	650	134	52
18-Jan-10	16	1414	183	<b>538</b>	19-Jan-10	613	98	52
18-Jan-10	17	<b>11199</b>	<b>1515</b>	<b>450</b>	19-Jan-10	1050	226	<b>144</b>
					20-Jan-10	359	73	30
18-Jan-10	18	7270	<b>583</b>	<b>327</b>	19-Jan-10	1333	<b>767</b>	<b>402</b>
					20-Jan-10	379	63	<b>122</b>
					21-Jan-10	122	41	41
18-Jan-10		8164	<b>933</b>	<b>450</b>	19-Jan-10	6867	<b>2909</b>	<b>759</b>
					20-Jan-10	670	41	<b>110</b>
					21-Jan-10	213	52	41
18-Jan-10	21	789	134	<b>122</b>	19-Jan-10	712	75	<b>145</b>
					20-Jan-10	379	41	<b>121</b>
					21-Jan-10	132	<10	10
19-Jan-10	21.1	908	201	<b>256</b>	20-Jan-10	960	75	<b>262</b>
					21-Jan-10	97	20	20
26-Feb-10	15	2481	135	<b>109</b>	27-Feb-10	4106	119	<b>228</b>
					28-Feb-10	98	10	20
26-Feb-10	17	563	305	<b>216</b>	27-Feb-10	2224	10	10
9-Mar-10	19	<10	<10	<b>253</b>	10-Mar-10	10	10	<10
20-Apr-10	19	880	<10	<b>292</b>	21-Apr-10	63	20	20
20-Apr-10	21.1	529	<10	<b>1439</b>	21-Apr-10	30	<10	<10
18-May-10	15	5172	86	<b>132</b>	19-May-10	776	52	20
26-May-10	15	7270	63	<b>301</b>	27-May-10	<b>&gt;24196</b>	96	<b>488</b>
					28-May-10	2909	<10	31
29-Jun-10	15	7701	<b>432</b>	<b>1223</b>	30-Jun-10	2755	31	31

APPENDIX C-2c  
*Bacteria counts greater than single sample maximum limit  
 July 2010 - June 2011*

Sample Date	Station	Total Coliform Count > 10,000 MPN/100mL	<i>E. coli</i> Count > 400 MPN/100mL	Enterococcus Count > 104 MPN/100mL	Follow-up sample date	Total Coliform Count	<i>E. coli</i> Count	Enterococcus Count
3-Aug-10	15	<b>14136</b>	10	41	4-Aug-10	959	63	30
17-Aug-10	16	7270	63	<b>135</b>	18-Aug-10	<10	<10	<10
14-Sep-10	15E	8664	121	<b>336</b>	15-Sep-10	683	63	52
15-Sep-10	15	<b>12997</b>	<b>2359</b>	<b>2755</b>	16-Sep-10	<b>&gt;24196</b>	185	<b>265</b>
					17-Sep-10	9804	256	<b>292</b>
					18-Sep-10	5475	<b>1106</b>	<b>836</b>
					19-Sep-10	<b>17329</b>	301	<b>435</b>
					20-Sep-10	417	<10	31
28-Sep-10	15	8664	108	<b>275</b>	29-Sep-10	5475	98	<b>173</b>
					30-Sep-10	<b>17329</b>	279	<b>645</b>
					1-Oct-10	<b>17329</b>	121	<b>450</b>
					2-Oct-10	<b>12033</b>	52	96
					3-Oct-10	1989	148	<b>226</b>
					4-Oct-10	<b>&gt;24196</b>	<b>605</b>	<b>228</b>
					5-Oct-10	98	20	31
3-Oct-10	15E	120	63	<b>420</b>	4-Oct-10	86	<10	<10
20-Oct-10	15	2187	75	<b>122</b>	21-Oct-10	10	<10	63
20-Oct-10	15E	52	31	<b>132</b>	21-Oct-10	41	10	20
8-Nov-10	21.1	158	132	<b>317</b>	9-Nov-10	156	52	<10
17-Nov-10	15	<b>12997</b>	75	20	18-Nov-10	9208	10	<10
22-Nov-10	17	203	52	<b>389</b>	23-Nov-10	20	<10	10
19-Dec-10	18	<b>11199</b>	<b>2382</b>	<b>717</b>	20-Dec-10	8164	<b>4352</b>	<b>4106</b>
					21-Dec-10	75	<10	31
19-Dec-10	19	4106	<b>703</b>	<b>441</b>	20-Dec-10	771	135	<b>148</b>
					21-Dec-10	855	63	10
19-Dec-10	20	521	135	<b>285</b>	20-Dec-10	891	74	<b>161</b>
					21-Dec-10	74	<10	<10
19-Dec-10	21	479	189	<b>191</b>	20-Dec-10	884	197	<b>393</b>
					21-Dec-10	52	20	10
19-Dec-10	21.1	650	161	<b>231</b>	20-Dec-10	448	75	<b>158</b>
					21-Dec-10	75	20	10
19-Dec-10	22	833	203	<b>175</b>	20-Dec-10	495	31	74
20-Dec-10	15	1450	<b>833</b>	<b>839</b>	21-Dec-10	389	52	<b>144</b>
					22-Dec-10	404	20	<10
20-Dec-10	15E	3654	<b>3255</b>	<b>209</b>	21-Dec-10	199	20	52
20-Dec-10	16	1989	<b>1334</b>	<b>1250</b>	21-Dec-10	41	10	20
20-Dec-10	17	548	122	<b>109</b>	21-Dec-10	134	31	10
29-Dec-10	19	3873	<b>565</b>	<b>146</b>	30-Dec-10	<b>12033</b>	<b>743</b>	<b>144</b>
					31-Dec-10	259	20	20
29-Dec-10	21	<b>17329</b>	<b>1565</b>	<b>512</b>	30-Dec-10	146	10	<10
18-Jan-11	15	2046	359	<b>441</b>	19-Jan-11	134	10	10
18-Jan-11	15E	529	355	<b>256</b>	19-Jan-11	10	<10	<10
18-Jan-11	16	2359	<b>663</b>	<b>1187</b>	19-Jan-11	41	10	10
18-Mar-11	15	6488	309	<b>173</b>	19-Mar-11	960	10	20
19-Mar-11	22	1670	275	<b>231</b>	20-Mar-11	9208	<b>860</b>	<b>605</b>
					21-Mar-11	109	<10	10
22-Jun-11	15	<b>&gt;24196</b>	63	52	23-Jun-11	960	<10	<10

Appendix C-2d  
*Bacteria counts greater than single sample maximum limit and follow-up samples*  
 July 2011 - June 2012

Sample Date	Station	Total Coliform Count > 10,000 MPN/100mL	<i>E. coli</i> Count > 400 MPN/100mL	Enterococcus Count > 104 MPN/100mL	Follow-up sample date	Total Coliform Count	<i>E. coli</i> Count	Enterococcus Count
12-Sep-11	15	5794	327	<b>1076</b>	13-Sep-11	4884	31	41
5-Oct-11	15	<b>&gt;24196</b>	62	<b>146</b>	6-Oct-11	644	31	10
11-Oct-11	17	789	<b>432</b>	<b>110</b>	12-Oct-11	3255	<b>1374</b>	<b>602</b>
					13-Oct-11	41	<10	<10
24-Oct-11	15	683	97	<b>171</b>	25-Oct-11	2909	10	74
24-Oct-11	21.1	134	74	<b>119</b>	25-Oct-11	10	<10	10
21-Jan-12	22	586	331	<b>231</b>	22-Jan-12	1334	63	85
22-Jan-12	18	185	41	<b>187</b>	23-Jan-12	122	10	41
22-Jan-12	19	670	131	<b>145</b>	23-Jan-12	109	<10	41
23-Apr-12	18	836	<b>573</b>	41	24-Apr-12	<10	<10	<10
25-Jun-12	15	<b>11199</b>	20	52	26-Jun-12	960	<10	97
25-Jun-12	15E	281	<10	<b>496</b>	26-Jun-12	62	<10	<10

Appendix C-2e  
*Bacteria counts greater than single sample maximum limit and follow-up samples*  
 July 2012 - June 2013

Sample Date	Station	Total Coliform Count > 10,000 MPN/100mL	<i>E. coli</i> Count > 400 MPN/100mL	Enterococcus Count > 104 MPN/100mL	Follow-up sample date(s)	Total Coliform Count	<i>E. coli</i> Count	Enterococcus Count
9-Jul-12	21.1	1333	<b>538</b>	74	10-Jul-12	<10	<10	<10
16-Jul-12	15	8164	<b>738</b>	<b>146</b>	17-Jul-12	1112	41	20
4-Sep-12	15E	3076	<10	<b>223</b>	5-Sep-12	10	<10	<b>199</b>
					7-Sep-12	175	41	10
7-Sep-12	15	<b>19863</b>	<b>4352</b>	<b>2851</b>	8-Sep-12	6488	63	41
24-Sep-12	15	571	52	<b>146</b>	25-Sep-12	31	10	10
24-Sep-12	16	285	108	<b>441</b>	25-Sep-12	20	10	<10
15-Oct-12	15	520	250	<b>134</b>	16-Oct-12	1259	31	<10
15-Oct-12	16	<b>12033</b>	148	98	16-Oct-12	1187	<b>1187</b>	<b>121</b>
					17-Oct-12	3654	<10	<10
15-Oct-12	18	345	213	<b>262</b>	16-Oct-12	10	<10	<10
15-Oct-12	21.1	771	<b>520</b>	86	16-Oct-12	122	63	<b>134</b>
					17-Oct-12	10	<10	31
22-Oct-12	18	789	<b>538</b>	20	23-Oct-12	63	<10	<10
29-Oct-12	15	<b>10462</b>	10	41	30-Oct-12	20	<10	<10
21-Nov-12	18	> <b>24196</b>	> <b>24196</b>	<b>345</b>	22-Nov-12	203	20	<10
21-Nov-12	19	4611	<b>586</b>	<b>259</b>	22-Nov-12	85	10	31
21-Nov-12	21	> <b>24196</b>	<b>2098</b>	<b>512</b>	22-Nov-12	109	30	31
21-Nov-12	22	6131	<b>464</b>	<b>323</b>	22-Nov-12	988	20	20
28-Nov-12	15	5172	175	<b>173</b>	29-Nov-12	4352	298	<b>441</b>
					30-Nov-12	1722	<b>521</b>	<b>1396</b>
					1-Dec-12	> <b>24196</b>	97	<b>275</b>
					2-Dec-12	4884	<b>990</b>	<b>213</b>
					3-Dec-12	404	173	10
29-Nov-12	15E	148	86	<b>109</b>	30-Nov-12	1376	345	<b>1019</b>
					1-Dec-12	275	109	<b>331</b>
					2-Dec-12	4352	<b>988</b>	<b>228</b>
					3-Dec-12	201	10	10
29-Nov-12	16	3255	<b>727</b>	<b>3873</b>	30-Nov-12	3255	<b>1153</b>	<b>4106</b>
					1-Dec-12	282	20	<b>173</b>
					2-Dec-12	5475	<b>886</b>	<b>146</b>
					3-Dec-12	216	10	10
30-Nov-12	18	6131	<b>2282</b>	<b>723</b>	1-Dec-12	520	134	<b>464</b>
					2-Dec-12	> <b>24196</b>	<b>4106</b>	<b>908</b>
					3-Dec-12	121	30	<10
30-Nov-12	19	> <b>24196</b>	<b>24196</b>	> <b>24196</b>	1-Dec-12	323	20	63
30-Nov-12	20	3255	<b>1210</b>	<b>884</b>	1-Dec-12	318	20	63
30-Nov-12	21	> <b>24196</b>	> <b>24196</b>	<b>12997</b>	1-Dec-12	529	85	41
30-Nov-12	21.1	2755	<b>583</b>	<b>369</b>	1-Dec-12	404	31	<b>350</b>
					2-Dec-12	<b>24196</b>	<b>4884</b>	<b>1553</b>
					3-Dec-12	97	<10	<10
30-Nov-12	22	591	160	<b>120</b>	2-Dec-12	1529	216	86
2-Dec-12	17	4884	<b>1396</b>	<b>393</b>	3-Dec-12	175	31	20
2-Dec-12	21	> <b>24196</b>	<b>6867</b>	<b>1050</b>	3-Dec-12	384	<10	<10
5-Dec-12	18	> <b>24196</b>	<b>4106</b>	<b>556</b>	6-Dec-12	74	<10	10
5-Dec-12	19	> <b>24196</b>	<b>24196</b>	<b>6867</b>	6-Dec-12	3873	<b>565</b>	<b>110</b>
					7-Dec-12	171	52	31
5-Dec-12	21	7701	<b>703</b>	<b>155</b>	6-Dec-12	288	63	<10

Appendix C-2e (cont.)  
*Bacteria counts greater than single sample maximum limit and follow-up samples*  
 July 2012 - June 2013

Sample Date	Station	Total Coliform Count > 10,000 MPN/100mL	<i>E. coli</i> Count > 400 MPN/100mL	Enterococcus Count > 104 MPN/100mL	Follow-up sample date(s)	Total Coliform Count	<i>E. coli</i> Count	Enterococcus Count
5-Dec-12	22	>24196	3448	2143	6-Dec-12	504	52	10
22-Dec-12	15	>24196	72	160	23-Dec-12	5172	<10	30
24-Dec-12	19	1376	282	135	25-Dec-12	160	10	<10
26-Dec-12	15	19863	201	85	27-Dec-12	235	20	<10
26-Dec-12	19	798	185	106	27-Dec-12	179	30	20
26-Dec-12	22	4884	563	110	27-Dec-12	201	<10	<10
1-Apr-13	15	6131	146	185	2-Apr-13	583	<10	10
28-May-13	15	1664	122	173	29-May-13	2187	20	10

Appendix C-3  
 Summary of treated Combined Sewer Discharges (CSD)  
 and rainfall July 2012 - June 2013

Month	Day	CSD Location and Duration Times					Number of Treated Discharge Events	Inches of Rain	Number of Rain Days	Monthly Rain Total (inches)
		Sea Cliff I	Sea Cliff II	Lincoln	Vicente	Lake Merced				
October	21							0.04		
	22							0.76		
	23							0.17		
	24							0.41		
	25							0.05		
	31							0.16	6	1.59
November	1							0.3		
	8							0.11		
	9							0.01		
	16							0.69		
	17							0.84		
	20							0.54		
	21			4 hr 40 min	4 hr 31 min	3 hr 44 min	1	0.54		
	28	6 min	18 min				1	0.82		
	29			9 hr 5 min	9 hr	8 hr 53 min	1	0.22		
	30							1.91	10	5.98
December	1							1.24		
	2	7 min	31 min	*	2 hr 30 min	2 hr 19 min	1	1.16		
	5			*	1 hr 20 min	48 min	1	1.15		
	11							0.03		
	12							0.06		
	14							0.03		
	15							0.12		
	16							0.08		
	17							0.24		
	21							0.24		
	22	3 min	15 min				1	0.89		
	23			*	1 hr 6 min	1 hr 41 min	1	1.48		
	25			*	2hr 46 min	1hr 11 min	1	1.18		
	26							0.24		
28							0.42			
29							0.19	16	8.75	
January	5							0.32		
	6							0.01		
	23							0.23	3	0.56
February	7							0.31		
	8							0.26		
	19							0.29	3	0.86
March	5							0.15		
	6							0.06		
	7							0.08		
	19							0.06		
	20							0.06		
	28							0.01		
	30							0.26		
31							0.16	8	0.84	
April	1							0.4		
	4							0.43		
	7							0.02	3	0.85
May	27							0.03		
	28							0.01	2	0.04
June	23							0.02		
	24							0.03		
	25							0.16		
	26							0.02	4	0.23
<b>Total number of treated CSDs July 2012 - June 2013</b>		<b>3</b>	<b>3</b>	<b>6</b>	<b>6</b>	<b>6</b>	<b>8</b>	<b>Total in. of rain 19.7</b>	<b>Total rain days 55</b>	

\* No data from the Lincoln structure due to lack of telemetry for this event.





Appendix C-5

Number of Users Engaged in Water Contact Recreation After a Treated Combined Sewer Discharge July 2012 - June 2013.  
Multiple observations on the same date indicates different people posting, deposting, or sampling.

Discharge			Recreational Use Observations					
Date and Structure	Duration	End Time	Date and Activity	Station	Time	Water Contact Number of Users		
						Full	Partial	Non
11/21/2012			11/21/2012	18	0540	0	0	0
	Lincoln 4 hr 40 min	0601	Posting	19	0600	0	0	0
	Vicente 4 hr 31 min	0532		20	0620	0	0	0
	Lake Merced 3 hr 44 min	0038		20	0625	0	0	0
				21	0630	0	0	0
				21.1	0645	0	0	0
				22	0655	0	0	0
			11/22/2012	20	1159	0	6	50
			De-Posting	21.1	1210	0	12	60
11/28/2012			11/28/2012	15	1351	0	0	1
	Sea Cliff I 6 min	1040	Sampling	15E	1357	0	0	2
	Sea Cliff II 18 min	1055		16	1347	0	0	1
				17	1415	4	0	0
			11/28/2012	15	1440	0	0	5
			Posting	15E	1445	0	0	3
				16	1450	0	0	2
				17	1430	8	0	5
11/30/2012			11/30/2012	18	0820	0	0	2
	Lincoln 9 hr 5 min	1305	Sampling	19	0725	0	0	1
	Vicente 9 hr	1251		20	1005	0	0	1
	Lake Merced 8 hr 53 min	1259		21	0940	0	0	0
			11/30/2012	18	1040	0	0	2
			Posting	19	1025	0	0	0
				20	1015	0	0	1
				20	1020	0	0	0
				21	1010	0	0	0
				21.1	1004	0	0	2
				22	0940	0	0	1
12/2/2012			12/2/2012	15	1245	0	0	3
	Sea Cliff I 7 min	0803	Posting	15E	1250	0	0	3
	Sea Cliff II 31 min	0829		16	1300	0	0	4
	Lincoln *	*		17	1240	0	0	6
	Vicente 2 hr 30 min	1047						
	Lake Merced 2 hr 19 min	1031	12/2/1012	18	1130	NR	NR	NR
			Sampling	20	1540	NR	NR	NR
				21	1550	NR	NR	NR
				21.1	1120	NR	NR	NR
				22	1610	NR	NR	NR
12/5/2012			12/5/2012	18	0815	0	0	5
	Lincoln *	*	Posting	19	0810	0	0	5
	Vicente 1 hr 20 min	0831		20	0845	1	0	5
	Lake Merced 48 min	0749		20	0855	0	0	2
				21	0827	0	0	0
				21.1	0820	0	0	0
				22	0755	0	0	0
			12/5/2012	18	1235	0	0	8
			Sampling	19	1259	0	0	3
				20	1315	0	0	4
				21	1328	0	0	1
				21.1	1215	1	0	0
				22	1130	0	0	0
			12/6/2012	20	1530	0	0	7
			De-Posting	21.1	0910	0	0	3
			12/7/2012	18	0845	0	0	5
			De-Posting	21	0850	0	0	2
				22	0840	0	0	0

\* No data from the Lincoln structure due to lack of telemetry for this event.

Appendix C-5 (cont.)

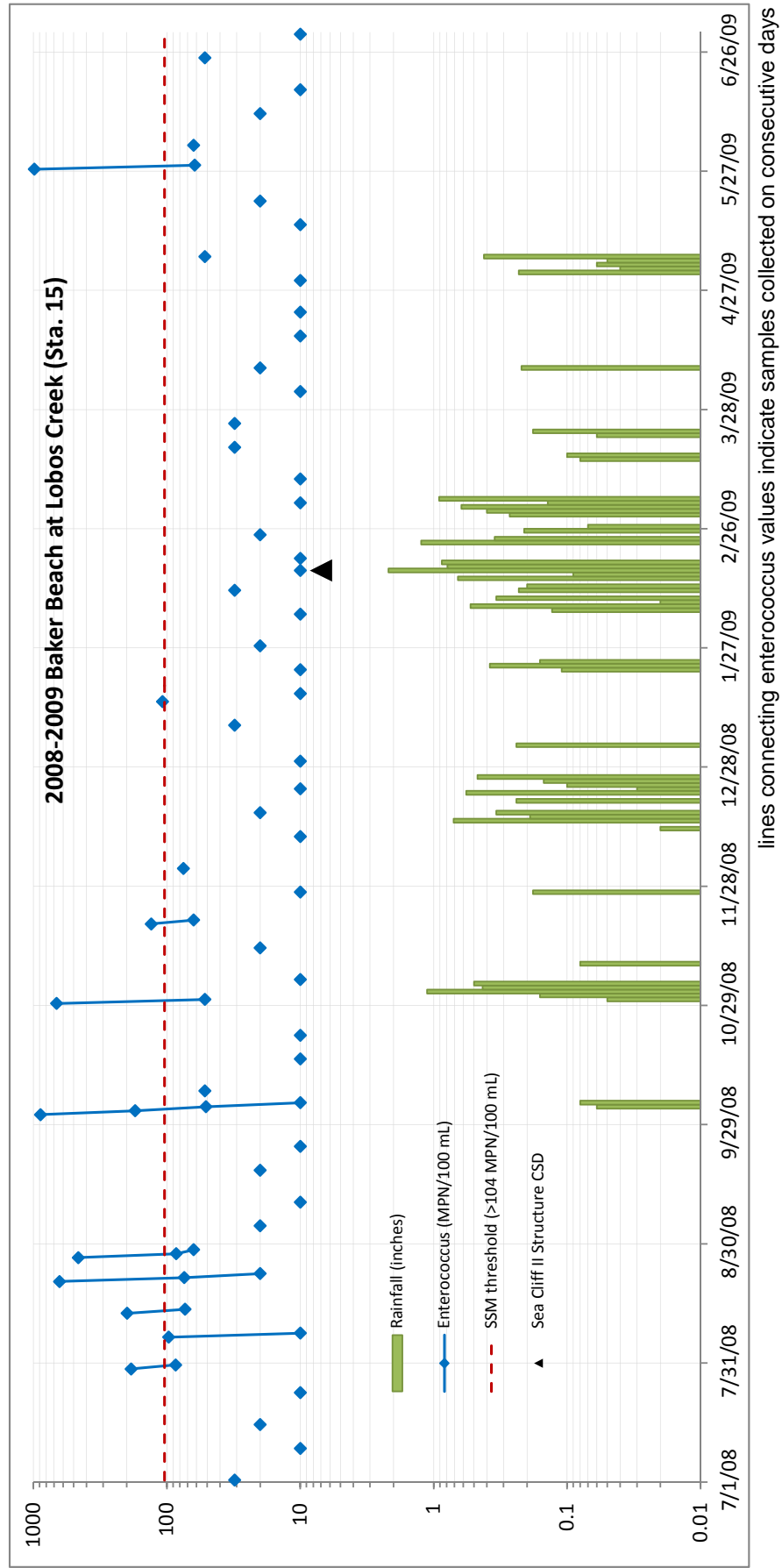
Number of Users Engaged in Water Contact Recreation After a Treated Combined Sewer Discharge July 2012 - June 2013.

Multiple observations on the same date indicates different people posting, deposting, or sampling.

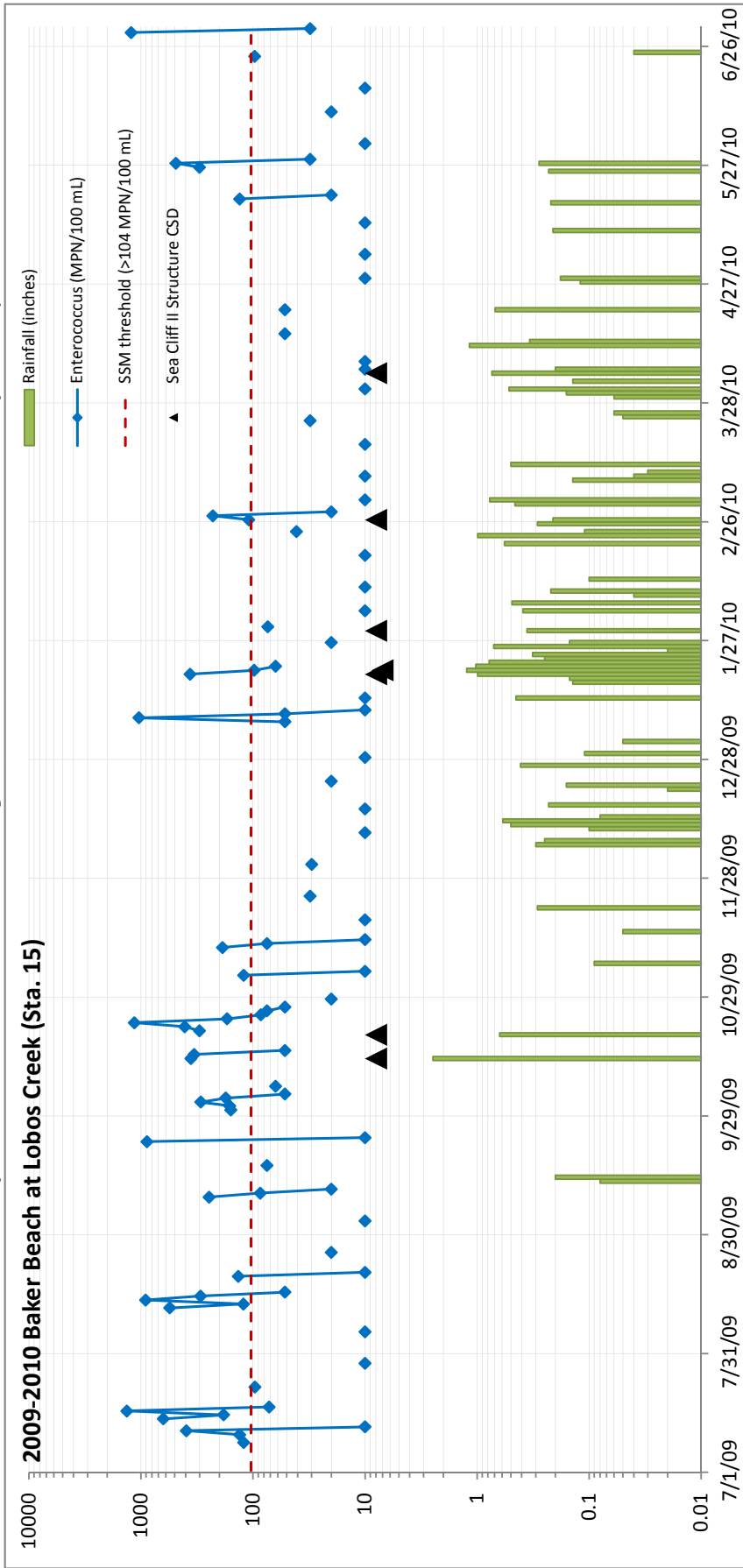
Discharge			Recreational Use Observations									
Date and Structure	Duration	End Time	Date and Activity	Station	Time	Water Contact Number of Users						
						Full	Partial	Non				
12/22/2012 Sea Cliff II	15 min	0355	12/22/2012	15	0820	0	0	2				
			Sampling/Posting	15E	0820	0	0	3				
				16	0820	0	0	2				
				12/24/2012	15	1147	0	0	55			
			De-Posting	15E	1145	2	0	18				
				16	1147	0	0	10				
12/23/2012	18	0855		0	0	10						
12/23/2012 Lincoln * Vicente 1 hr 6 min Lake Merced 1 hr 41 min	*	*	12/24/2012 Posting	19	0930	0	0	11				
				20	1005	0	0	4				
				21	1000	0	0	3				
				21.1	0950	0	0	3				
				22	0830	0	0	0				
				12/25/2012 De-Posting	18	0930	3	0	7			
			20	20	0945	0	0	7				
				21	0950	0	0	5				
				21.1	1000	6	0	5				
				22	0900	0	0	0				
				12/25/2012 Lincoln * Vicente 2 hr 46 min Lake Merced 1 hr 11 min	*	*	12/26/2012 Posting	18	0900	0	0	5
								19	0915	0	0	2
20	0920	0	0					0				
21	0925	0	0					2				
21.1	0930	0	0					0				
22	0835	0	0					0				
12/26/2012 Sampling	18	0910	0				0	2				
	19	0918	0				0	5				
	20	0928	0				0	3				
	21	0940	0				0	4				
	21.1	0947	0				0	1				
	12/27/2012 De-Posting	18	1025				0	1	15			
20		1012	0	0	0							
21		1010	0	2	2							
21.1		1200	0	1	5							
12/28/2012 De-Posting	19	0915	0	5	0							
	22	0830	0	0	0							

\* No data from the Lincoln structure due to lack of telemetry for this event.

Appendix C-6a  
*Enterococcus* values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 15 from July 1, 2008-June 30, 2009



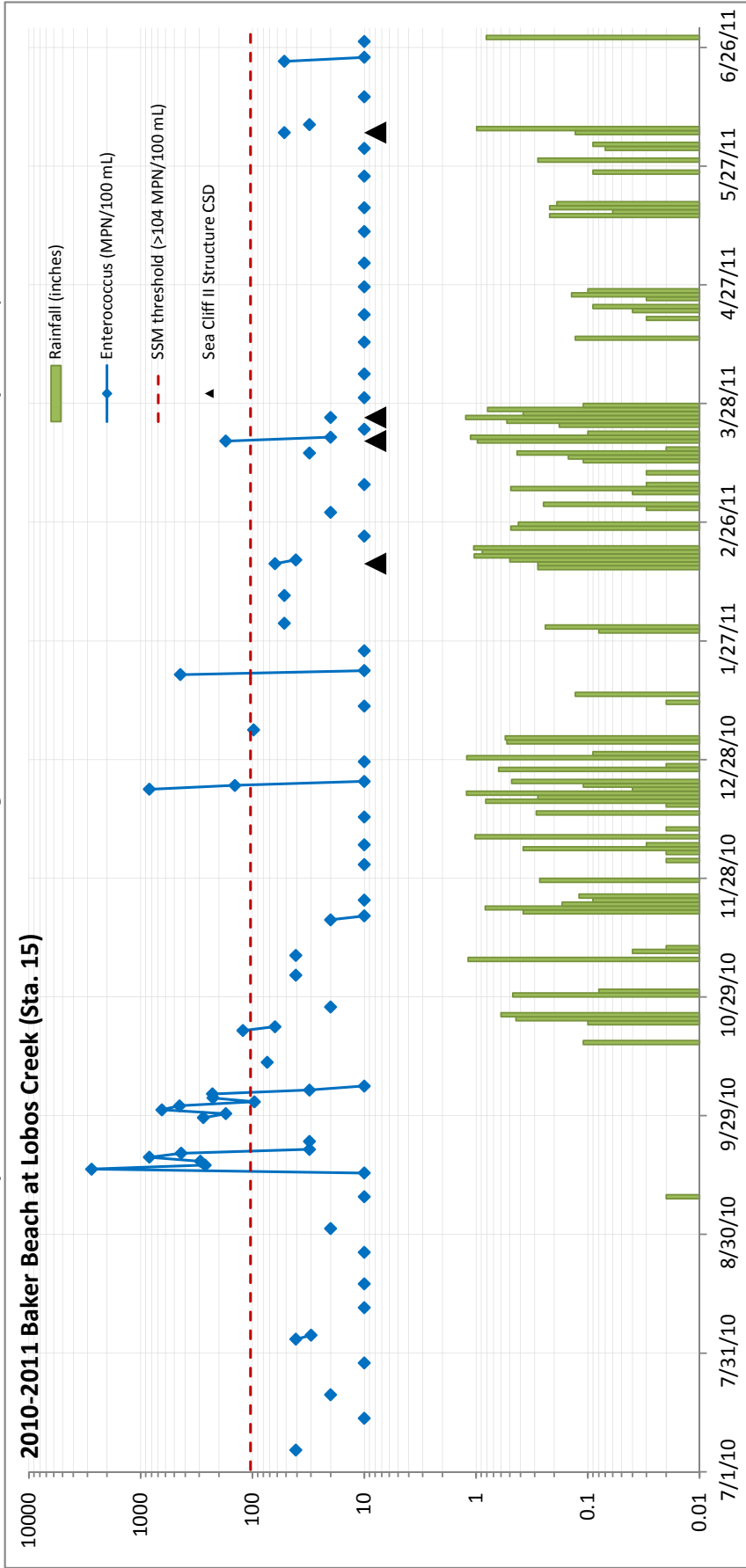
Appendix C-6b  
*Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 15 from July 1, 2009-June 30, 2010*



lines connecting enterococcus values indicate samples collected on consecutive days

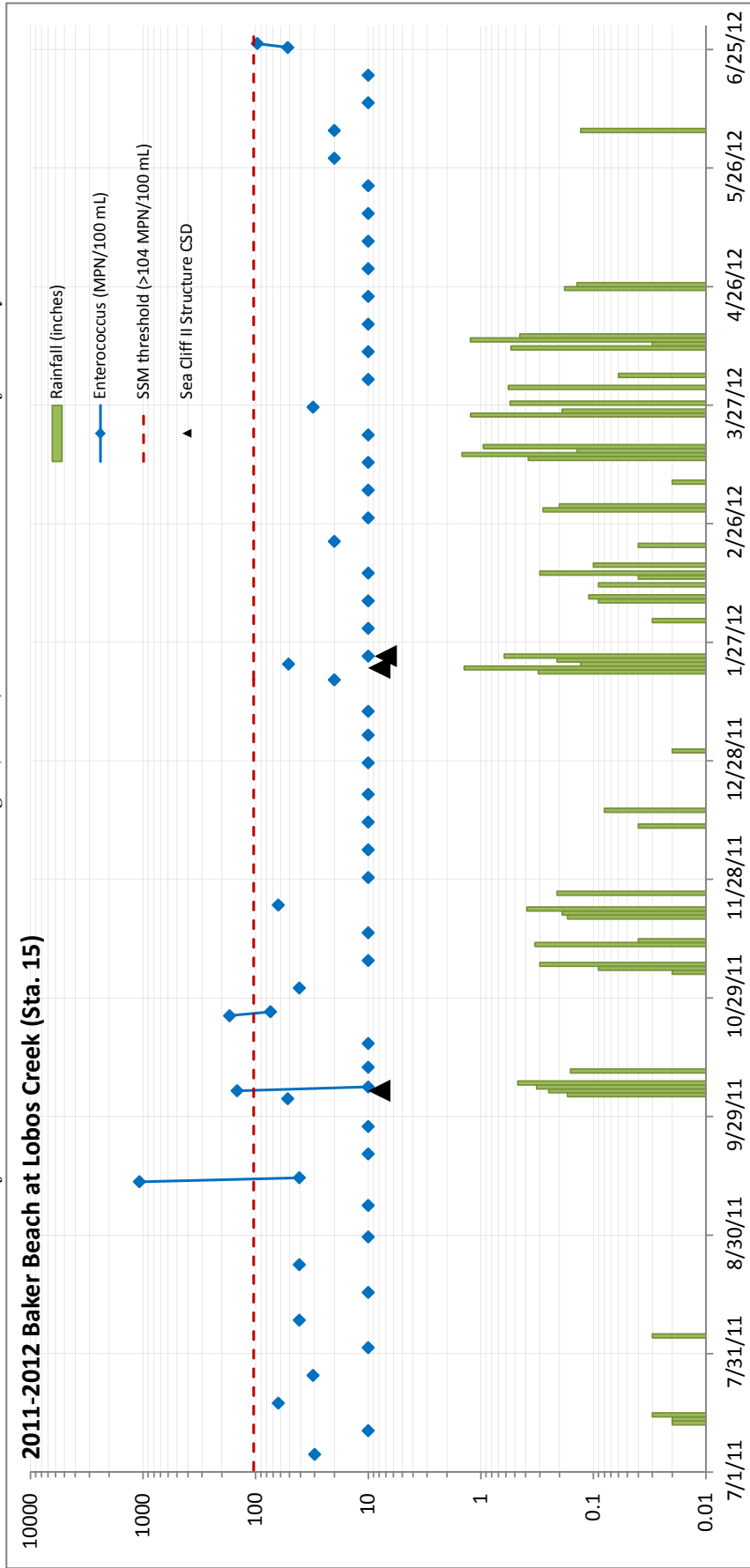
Appendix C-6c

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 15 from July 1, 2010-June 30, 2011



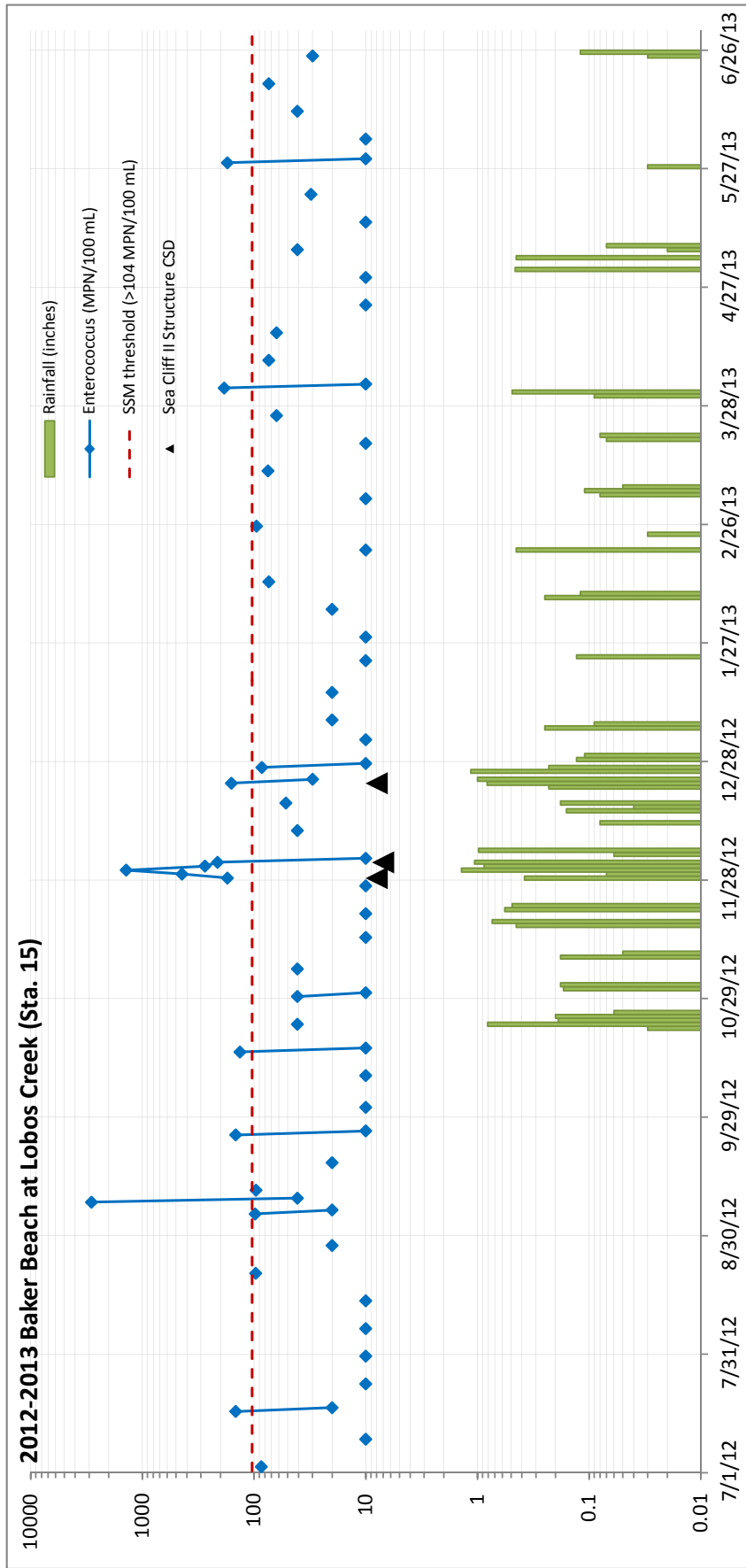
Appendix C-6d

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 15 from July 1, 2011-June 30, 2012

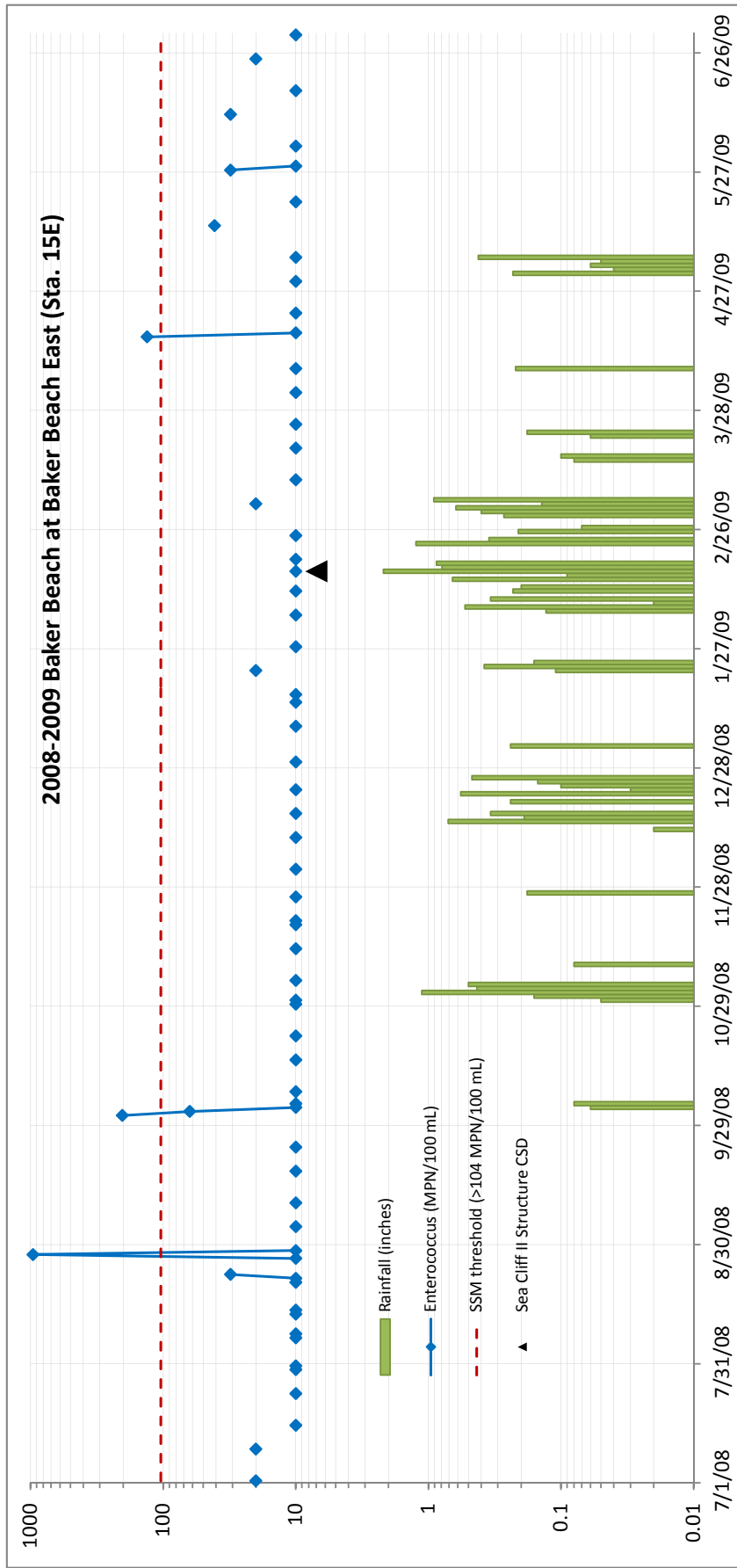


lines connecting enterococcus values indicate samples collected on consecutive days

Appendix C-6e  
*Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 15 from July 1, 2012-June 30, 2013*



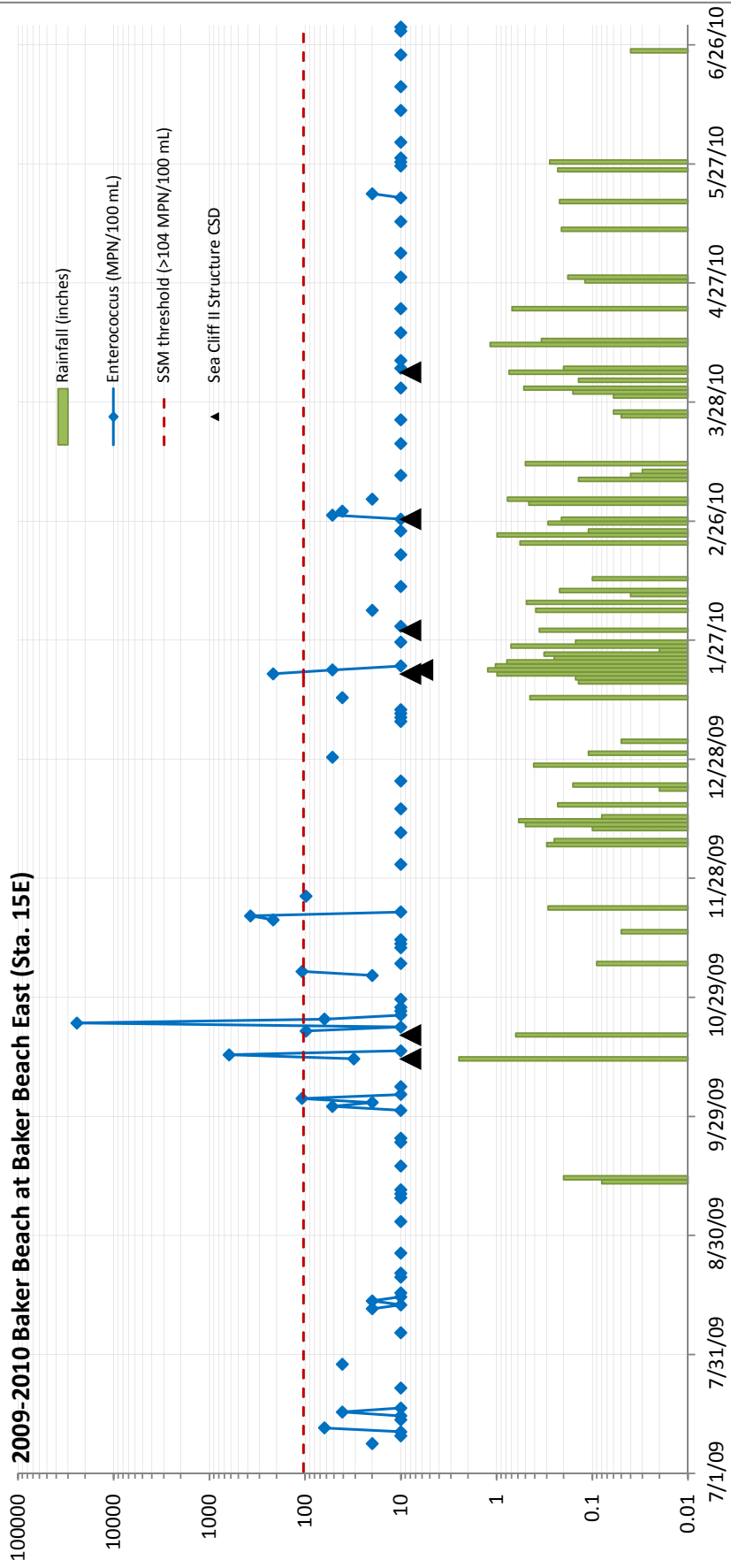
Appendix C-7a  
*Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 15E from July 1, 2008-June 30, 2009*



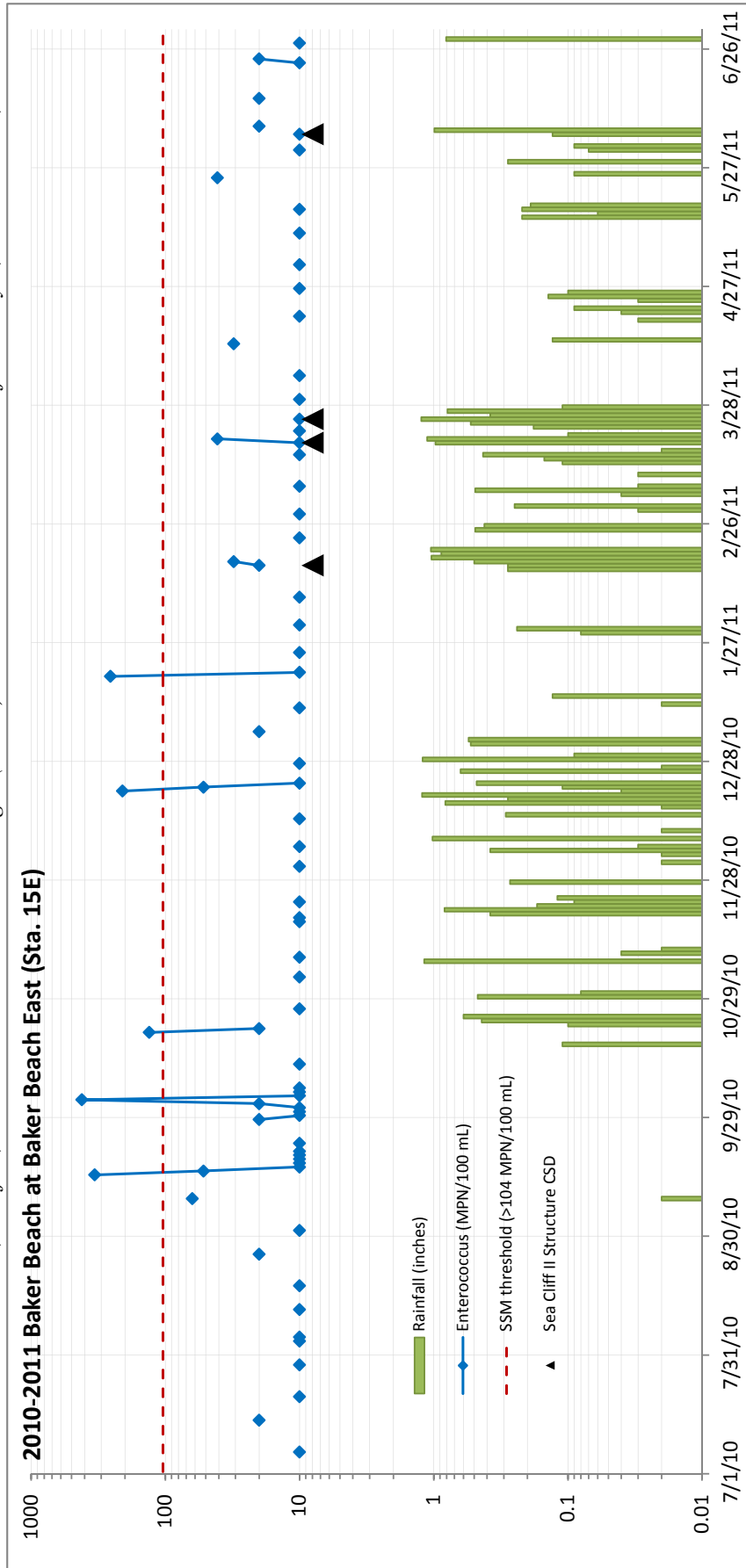


Appendix C-7b

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 15E from July 1, 2009-June 30, 2010



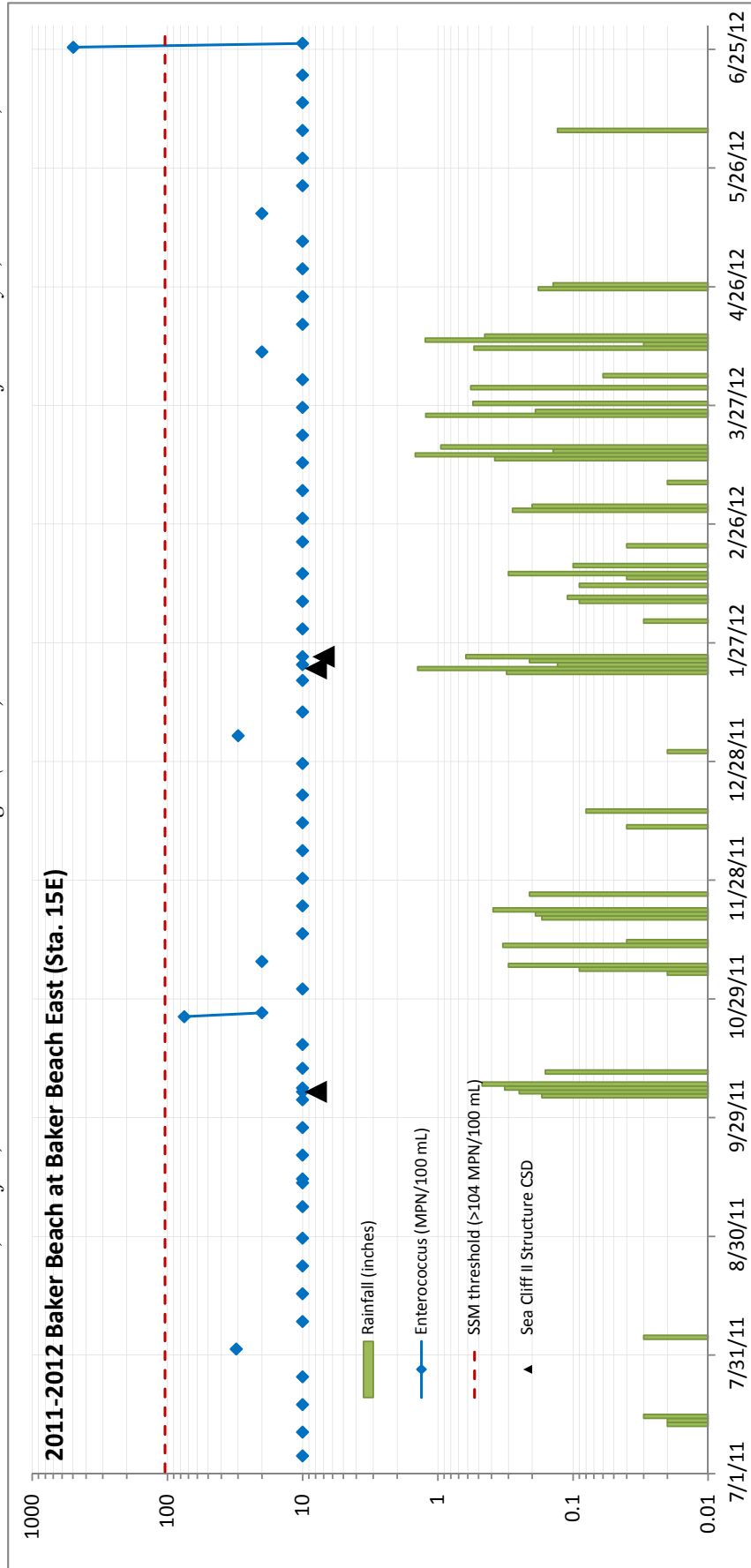
Appendix C-7c  
*Enterococcus* values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 15E from July 1, 2010-June 30, 2011



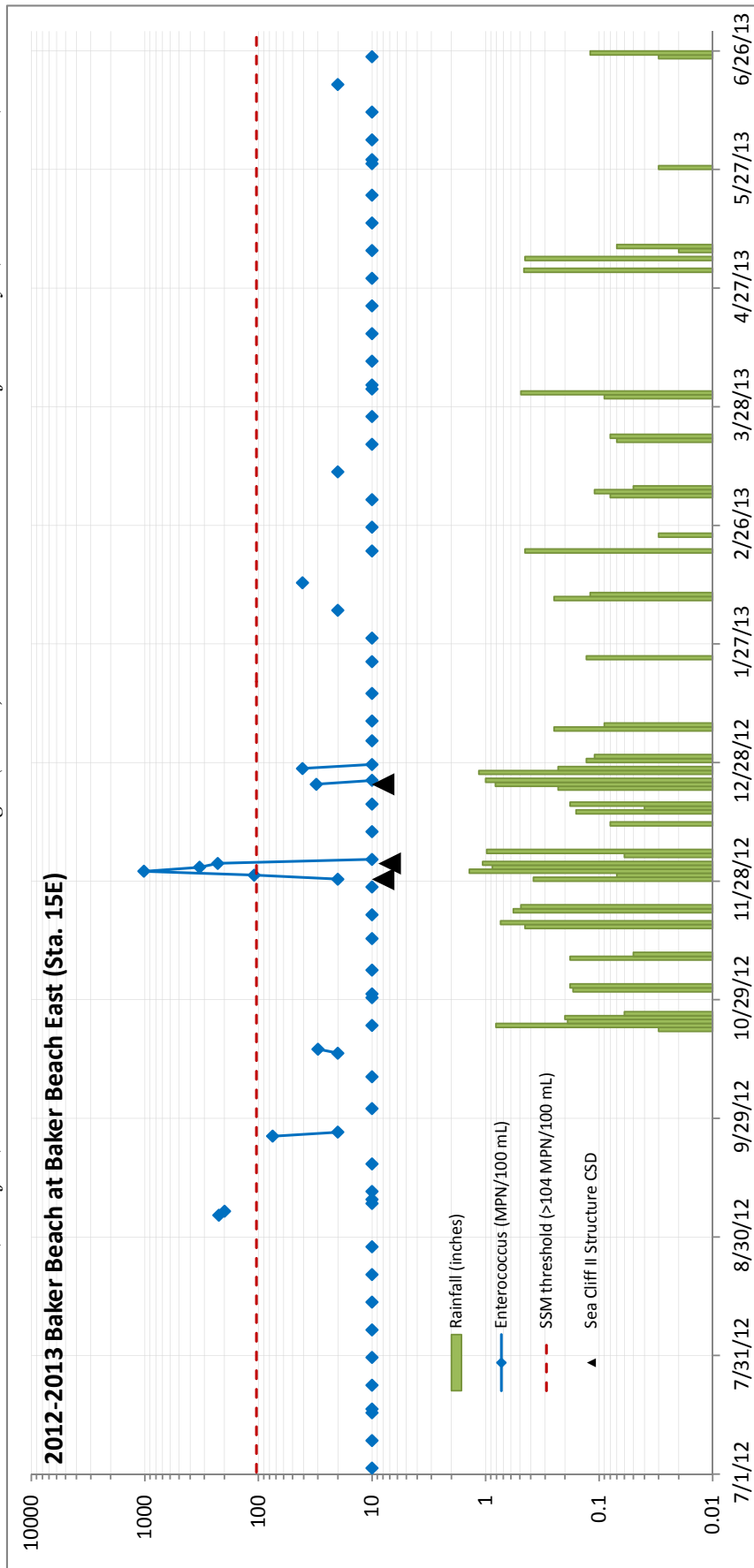
lines connecting enterococcus values indicate samples collected on consecutive days

Appendix C-7d

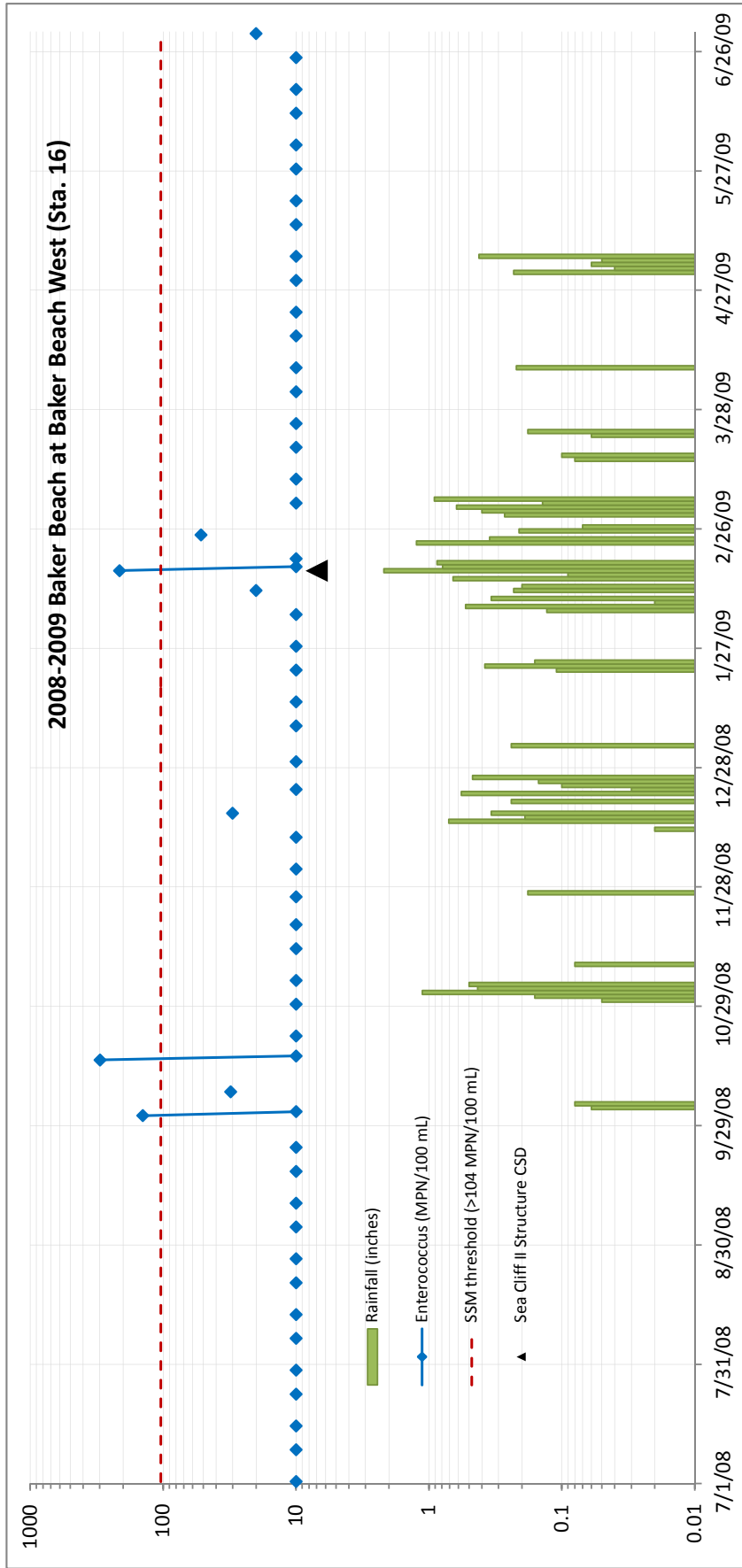
Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 15E from July 1, 2011–June 30, 2012



Appendix C-7e  
*Enterococcus* values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 15E from July 1, 2012-June 30, 2013

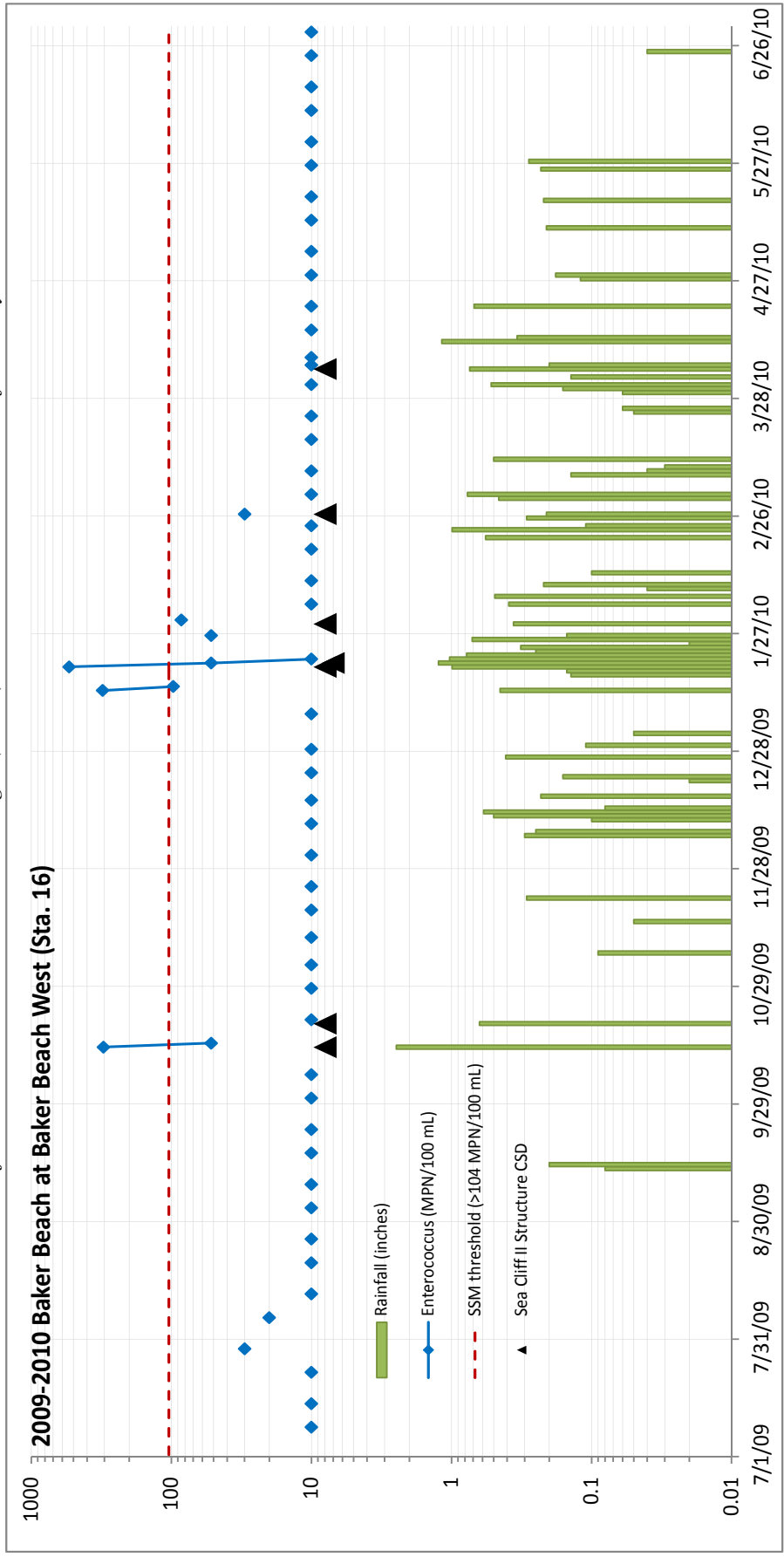


Appendix C-8a  
*Enterococcus* values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 16 from July 1, 2008-June 30, 2009



Appendix C-8b

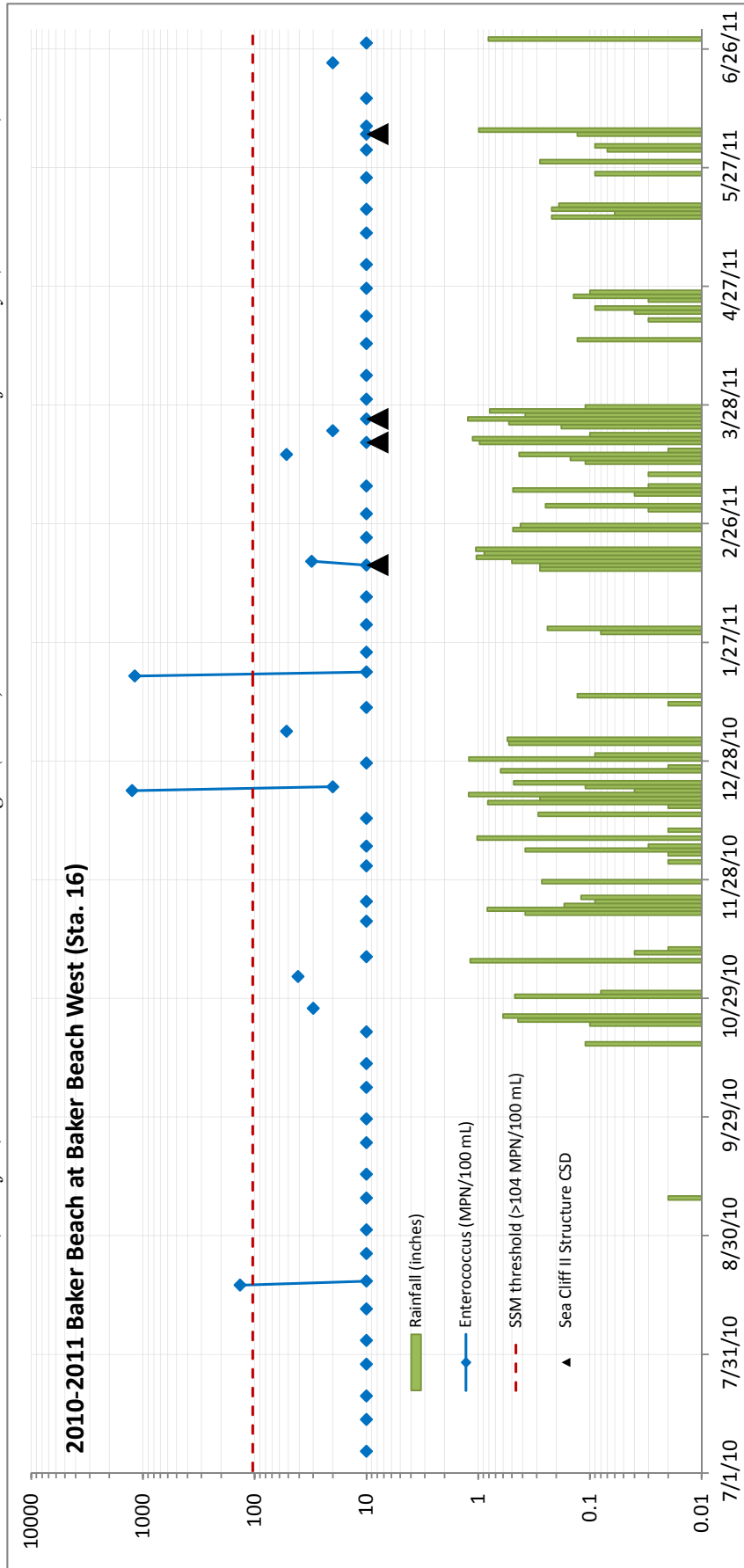
Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 16 from July 1, 2009-June 30, 2010



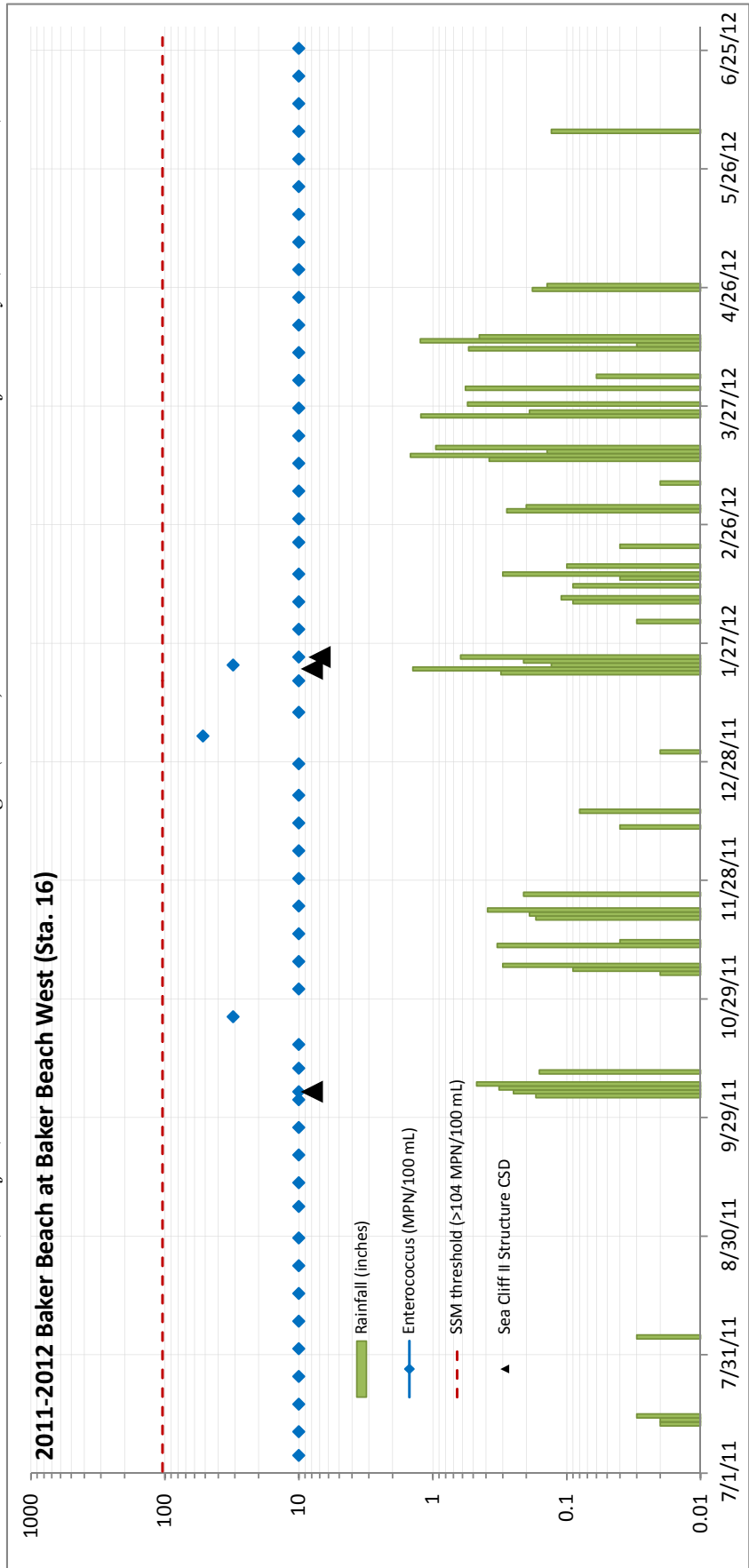
lines connecting enterococcus values indicate samples collected on consecutive days

Appendix C-8c

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 16 from July 1, 2010-June 30, 2011

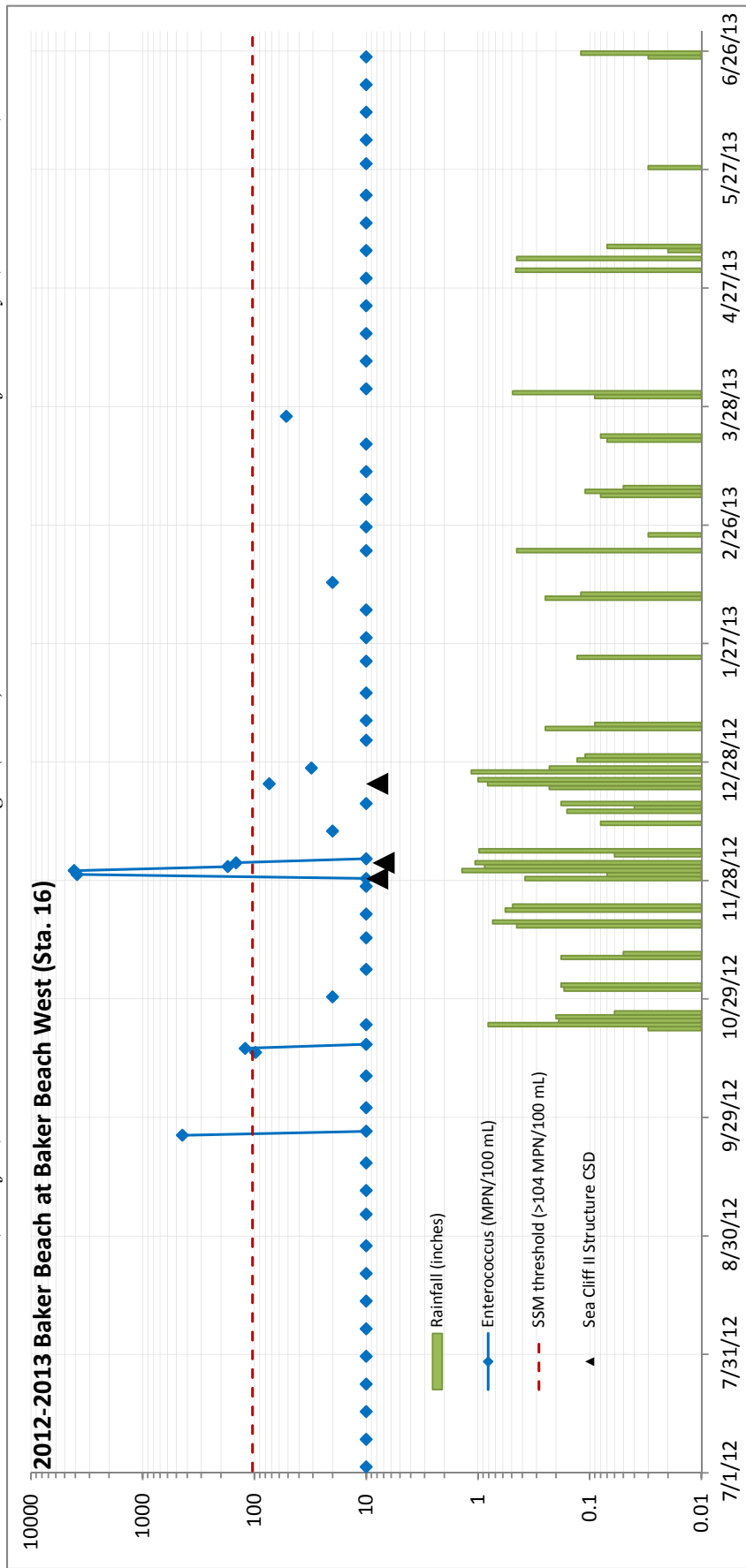


Appendix C-8d  
 Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 16 from July 1, 2011-June 30, 2012

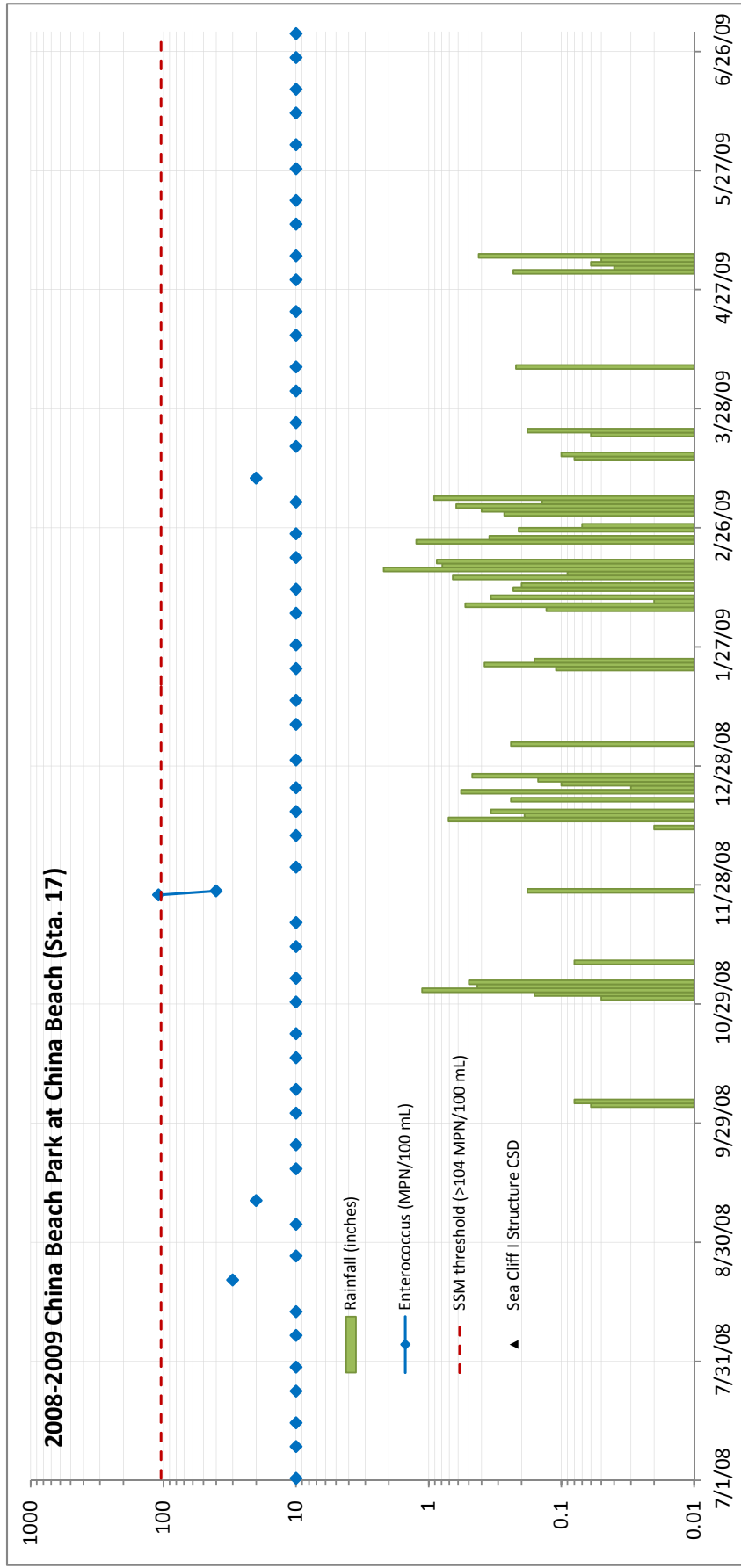




Appendix C-8e  
*Enterococcus* values, rainfall, and Combined Sewer Discharges (CSDs) at Baker Beach Station 16 from July 1, 2012-June 30, 2013

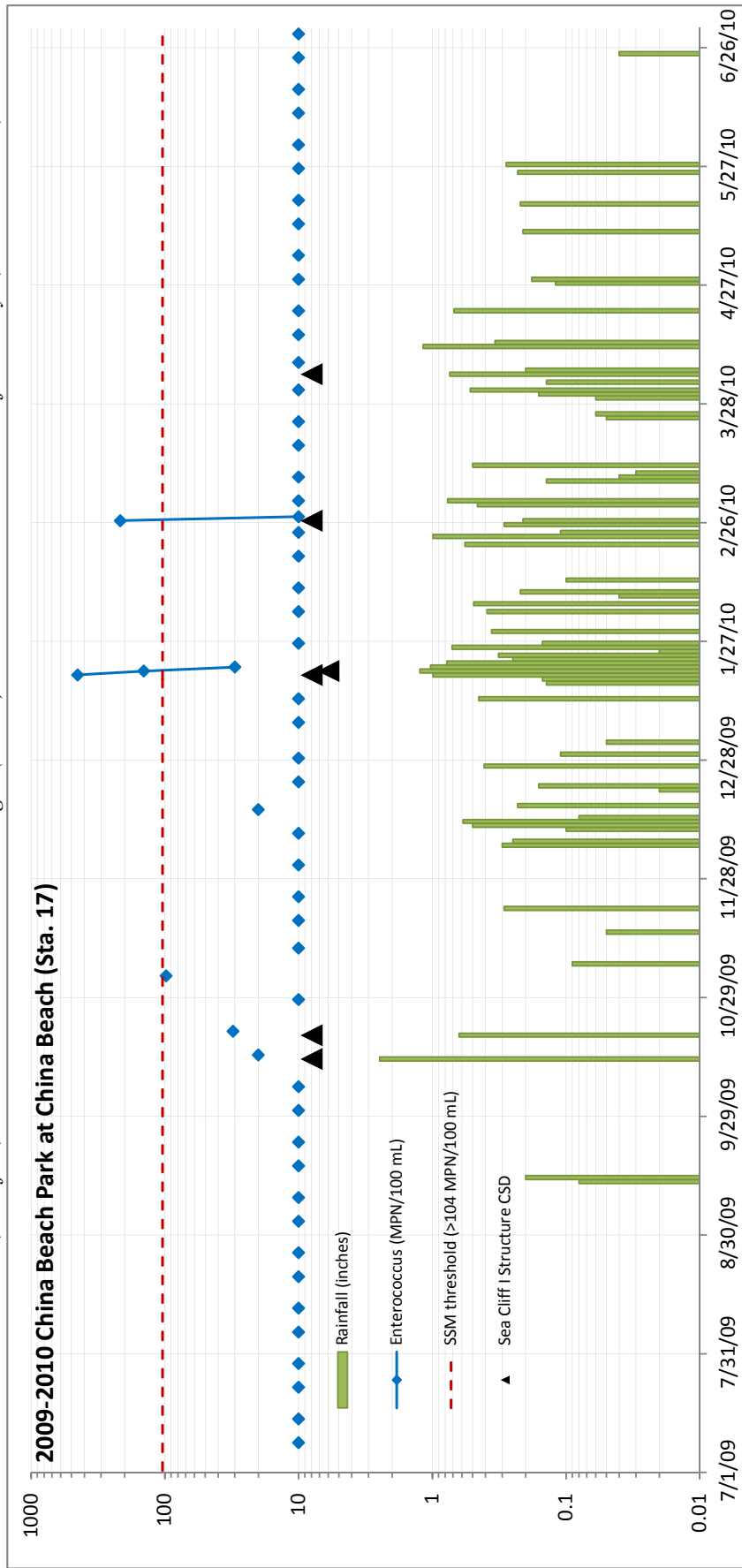


Appendix C-9a  
*Enterococcus* values, rainfall, and Combined Sewer Discharges (CSDs) at China Beach Station 17 from July 1, 2008-June 30, 2009



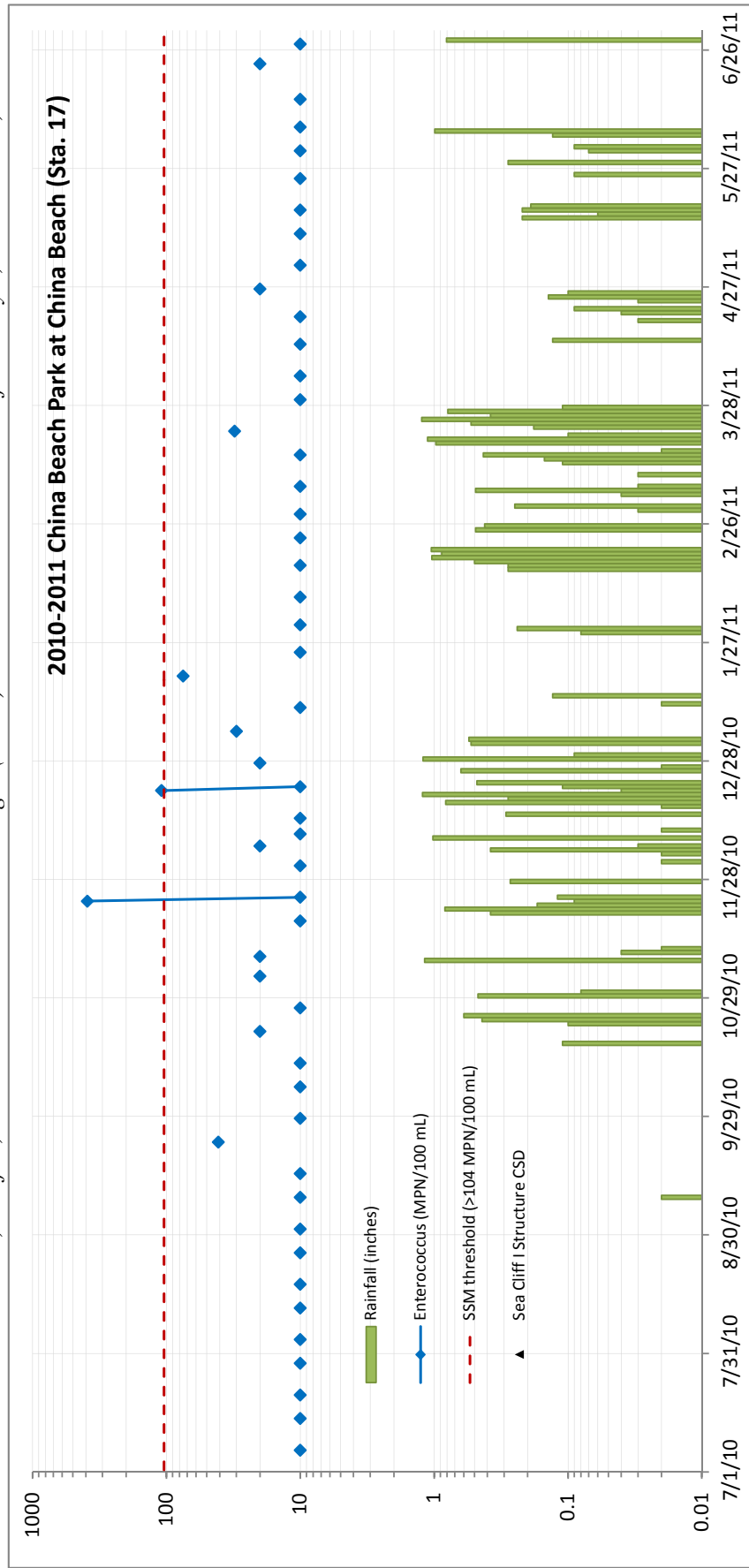
lines connecting enterococcus values indicate samples collected on consecutive days

Appendix C-9b  
*Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at China Beach Station 17 from July 1, 2009-June 30, 2010*

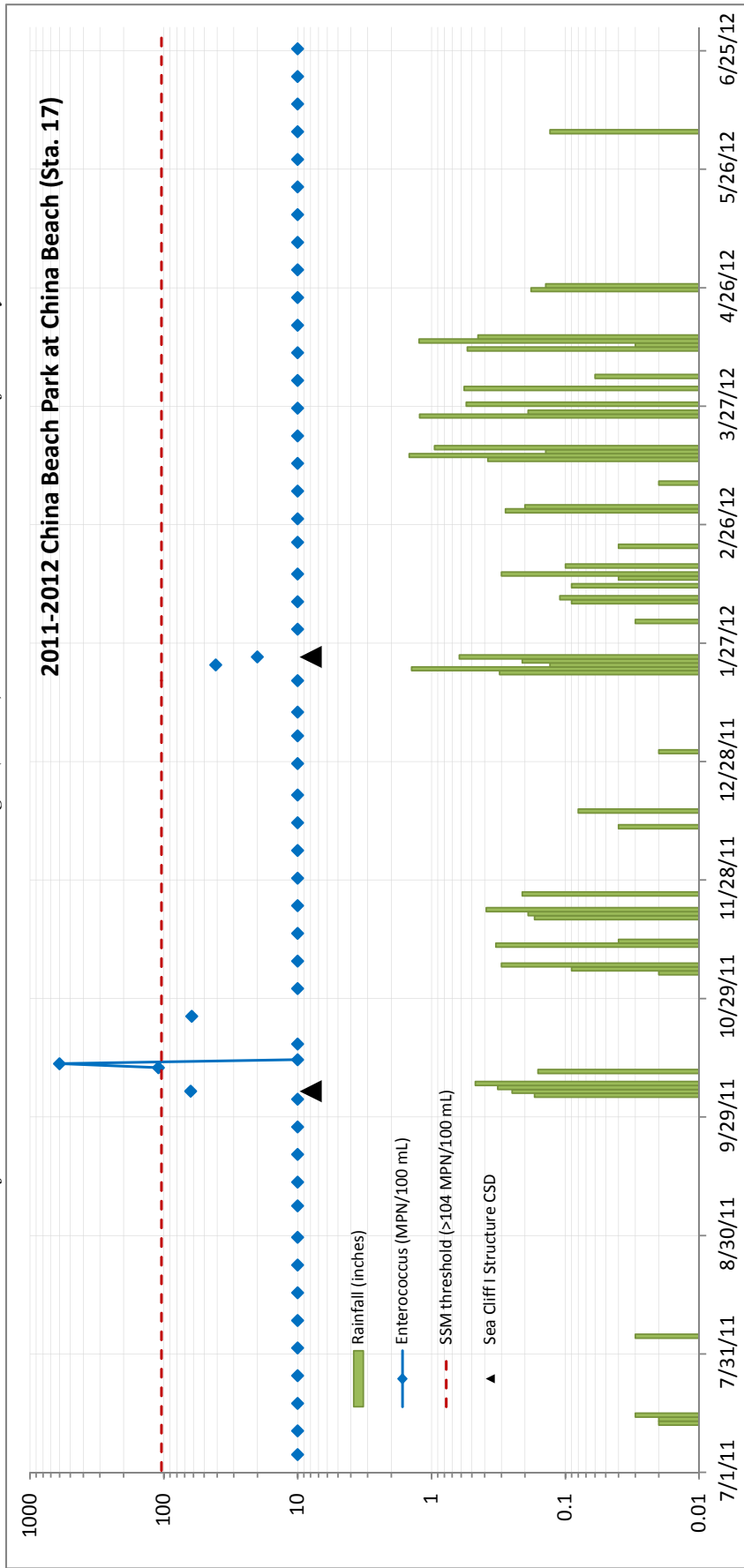


lines connecting enterococcus values indicate samples collected on consecutive days

Appendix C-9c  
 Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at China Beach Station 17 from July 1, 2010-June 30, 2011

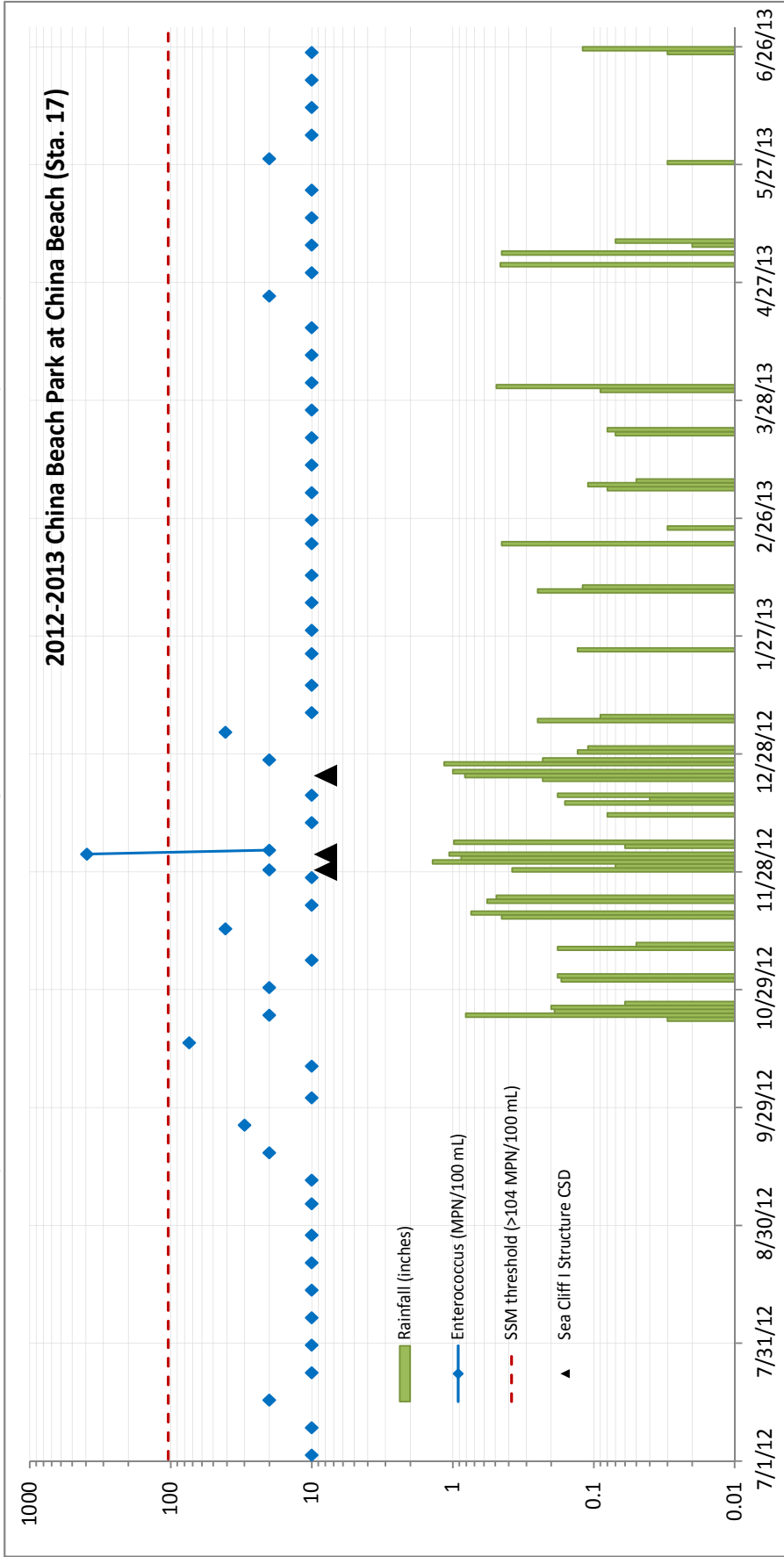


Appendix C-9d  
 Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at China Beach Station 17 from July 1, 2011-June 30, 2012



Appendix C-9e

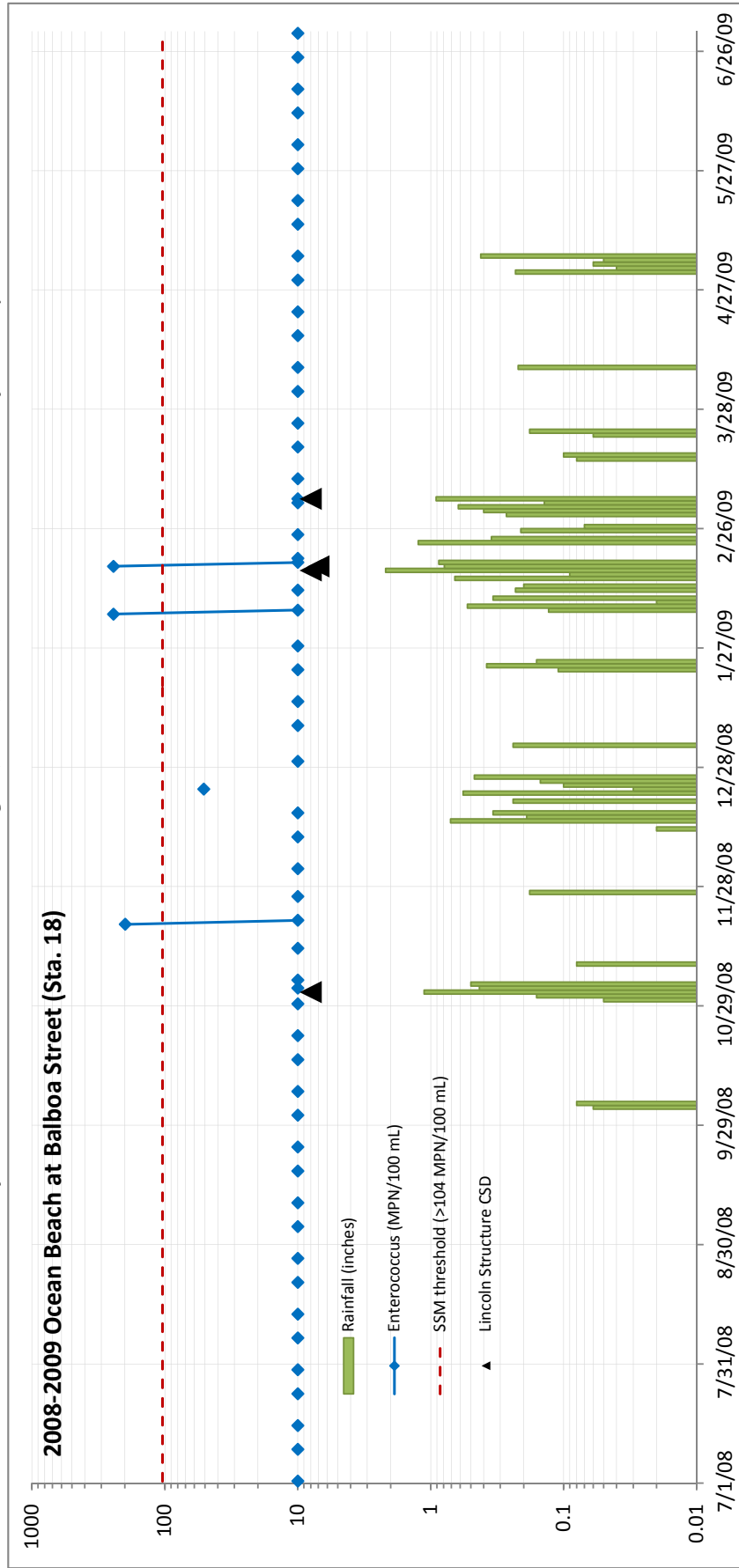
Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at China Beach Station 17 from July 1, 2012-June 30, 2013



lines connecting enterococcus values indicate samples collected on consecutive days

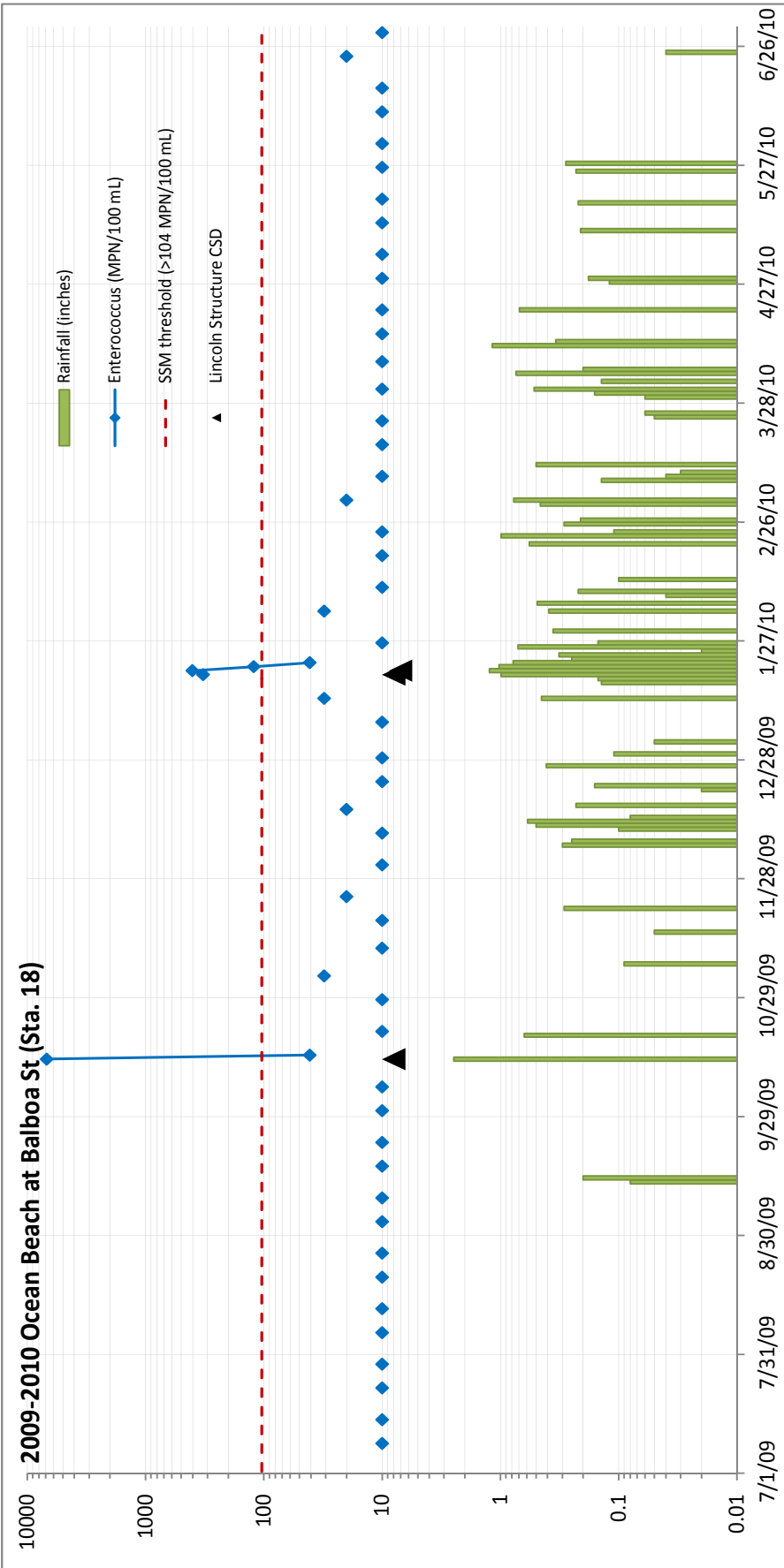
Appendix C-10a

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 18 from July 1, 2008-June 30, 2009



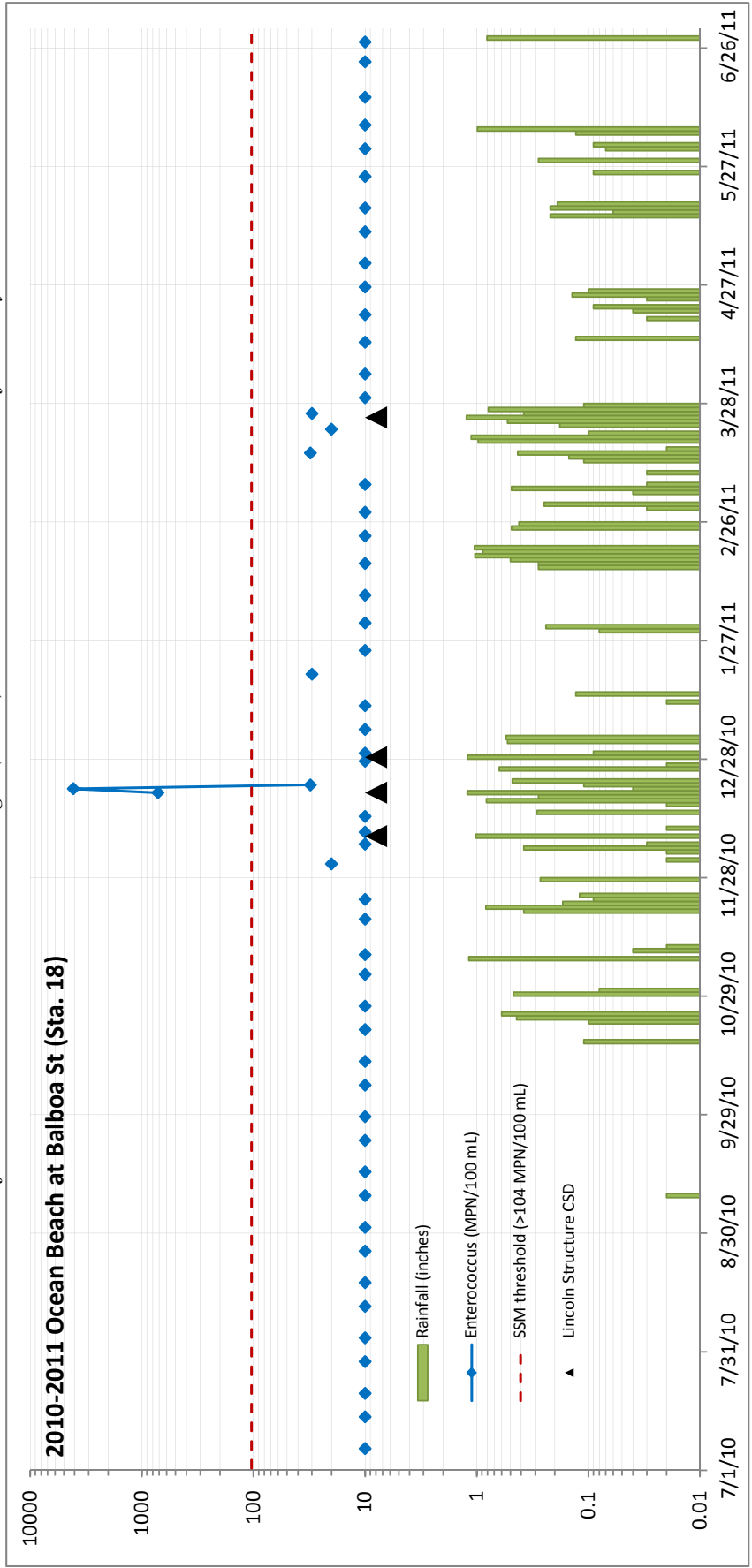
Appendix C-10b

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 18 from July 1, 2009-June 30, 2010



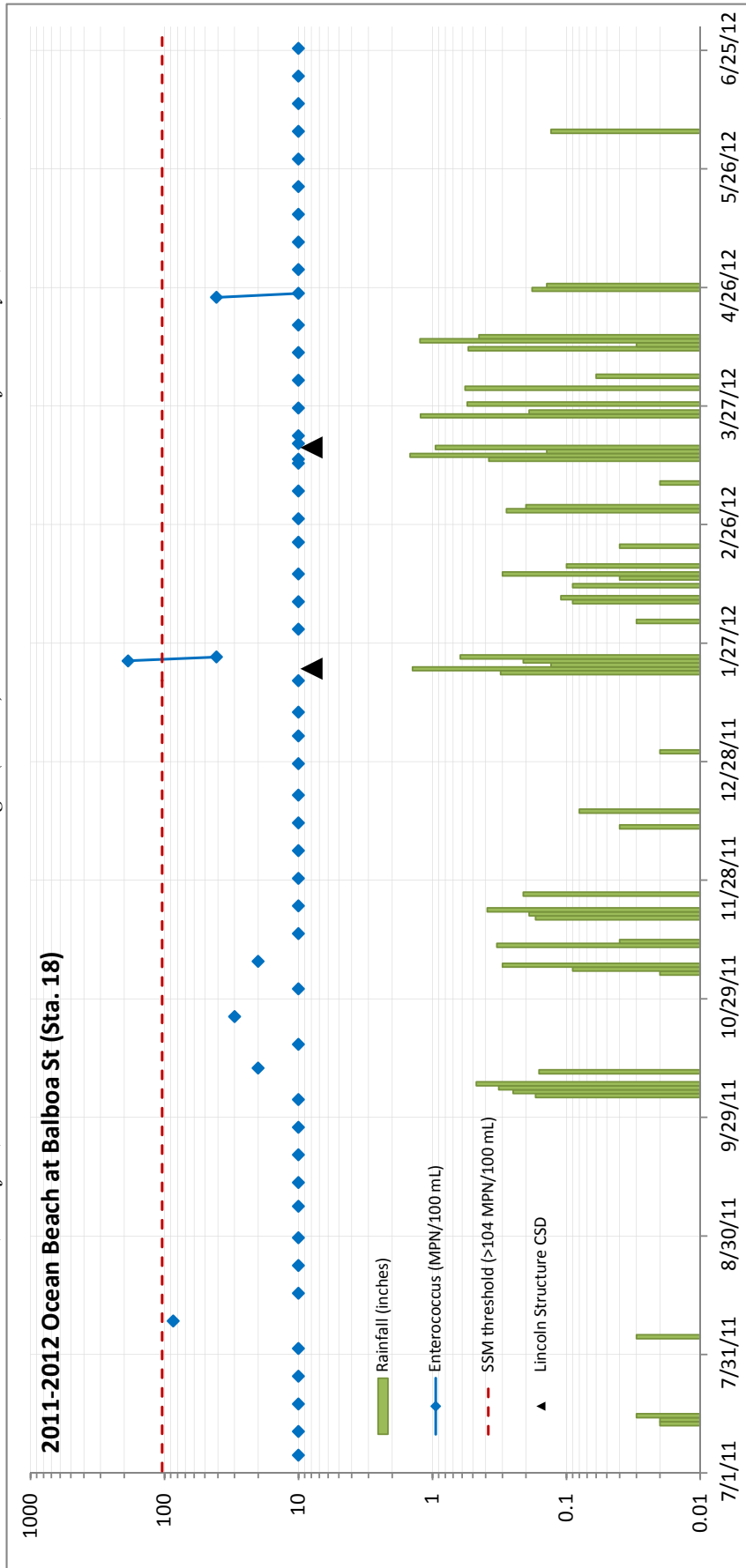


Appendix C-10c  
*Enterococcus* values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 18 from July 1, 2010-June 30, 2011

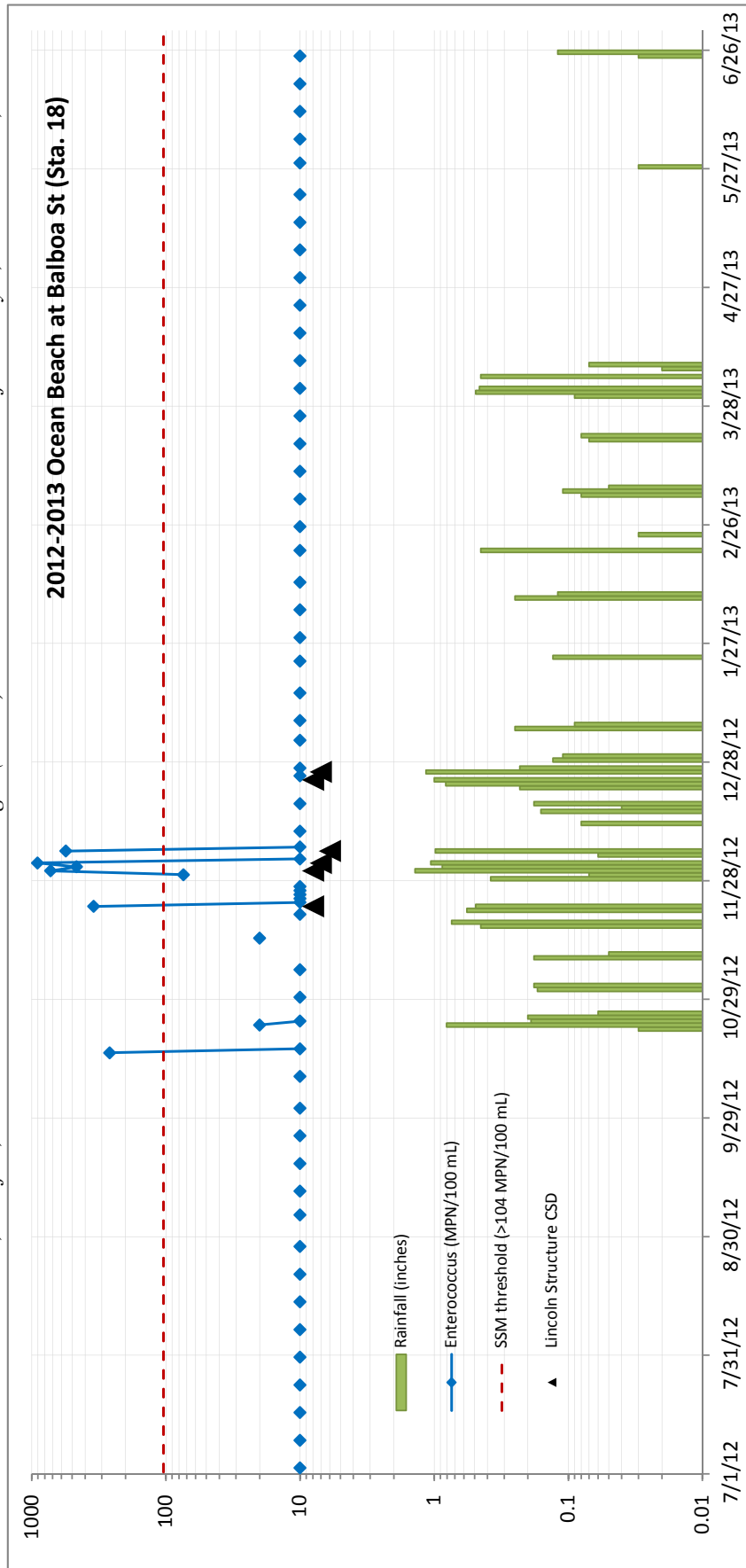


Appendix C-10d

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 18 from July 1, 2011-June 30, 2012

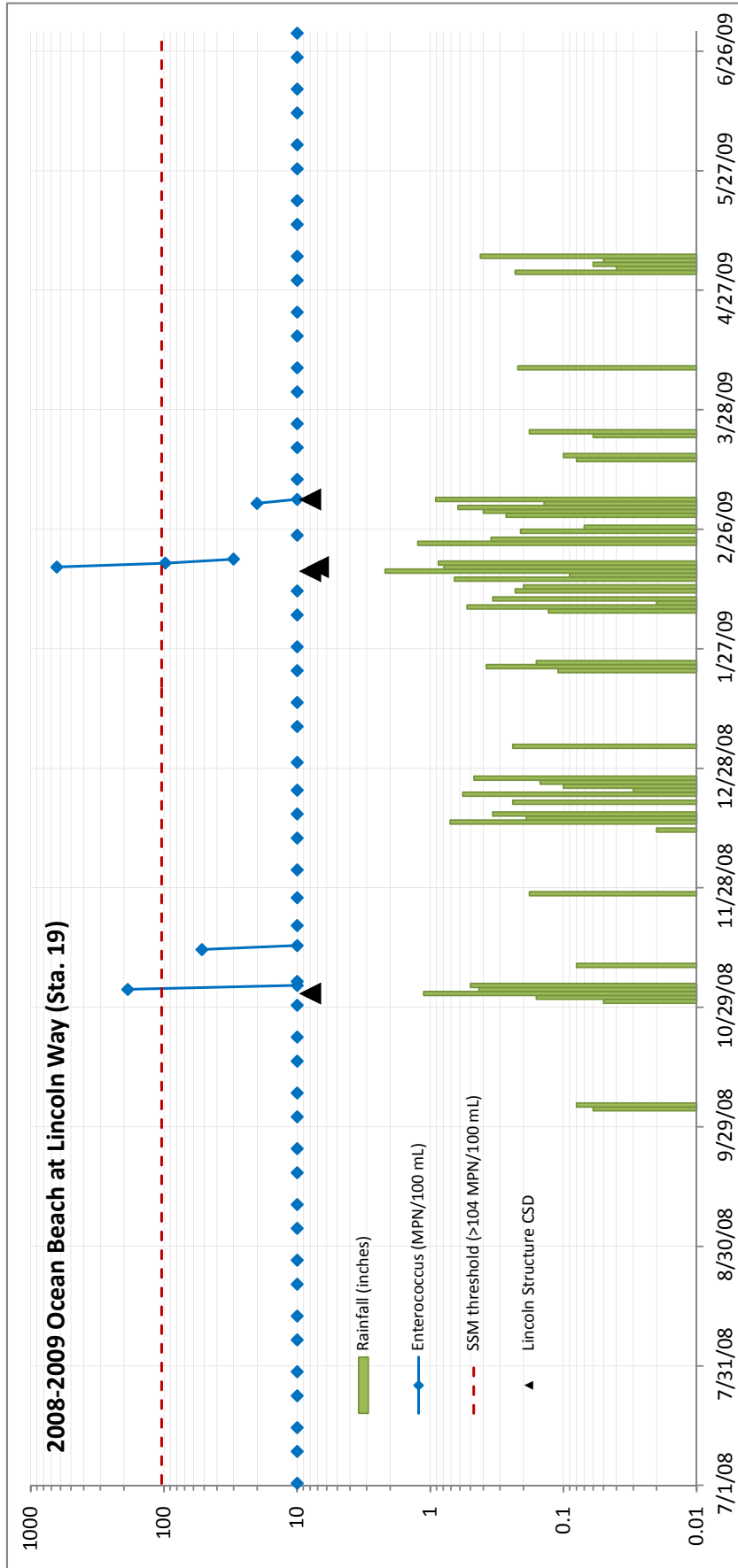


Appendix C-10e  
*Enterococcus* values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 18 from July 1, 2012-June 30, 2013

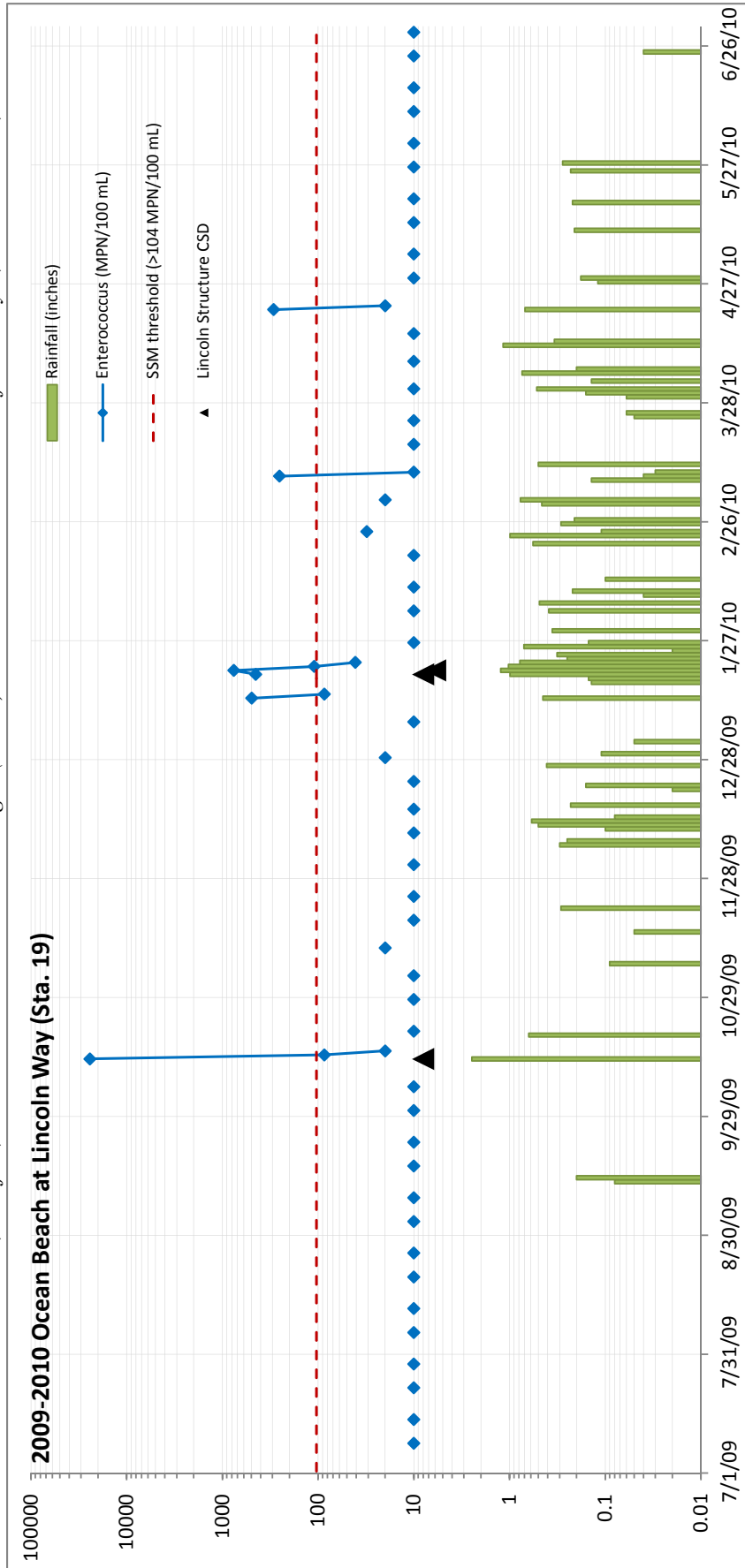


Appendix C-11a

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 19 from July 1, 2008-June 30, 2009

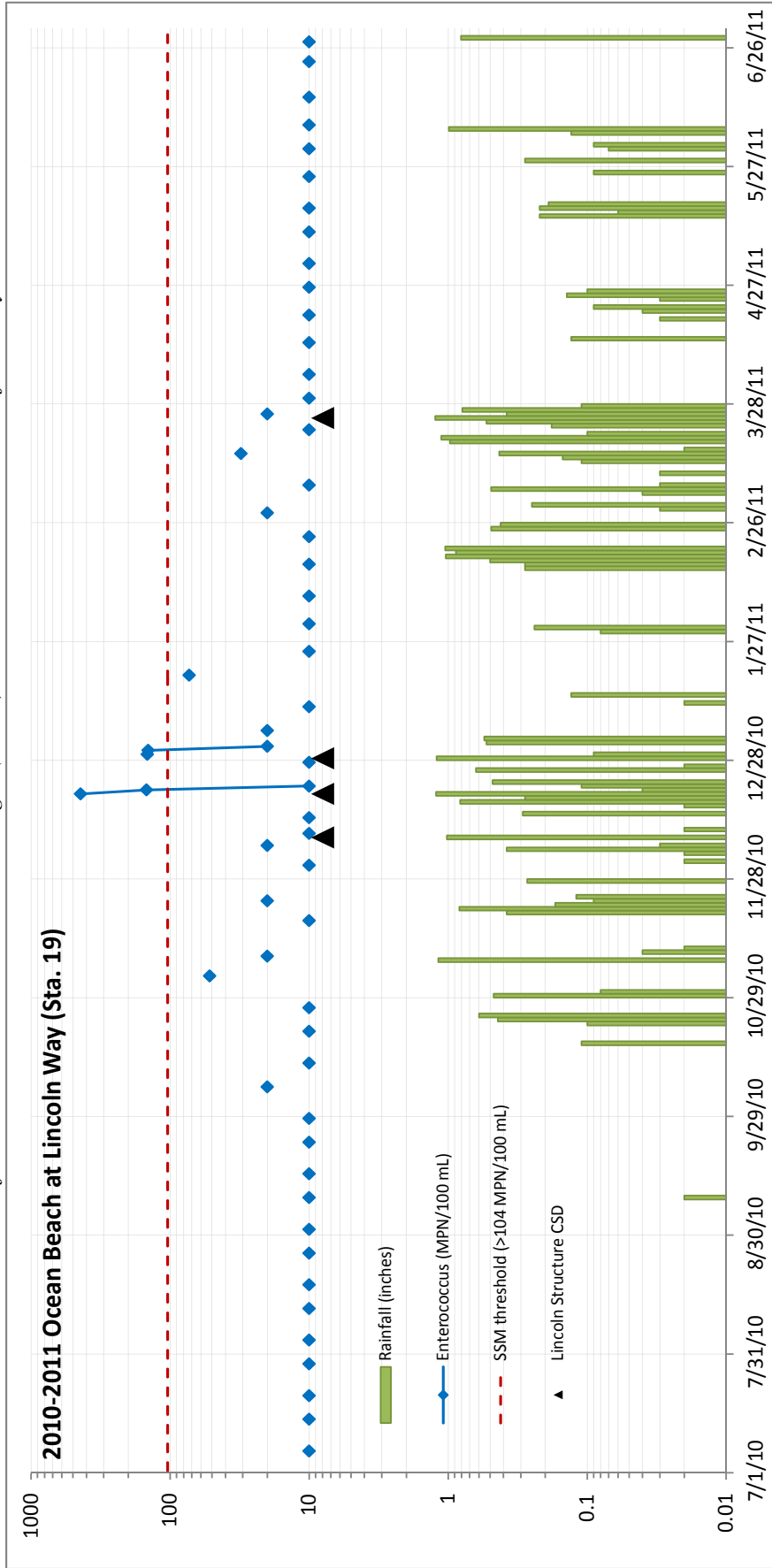


Appendix C-11b  
*Enterococcus* values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 19 from July 1, 2009-June 30, 2010



lines connecting enterococcus values indicate samples collected on consecutive days

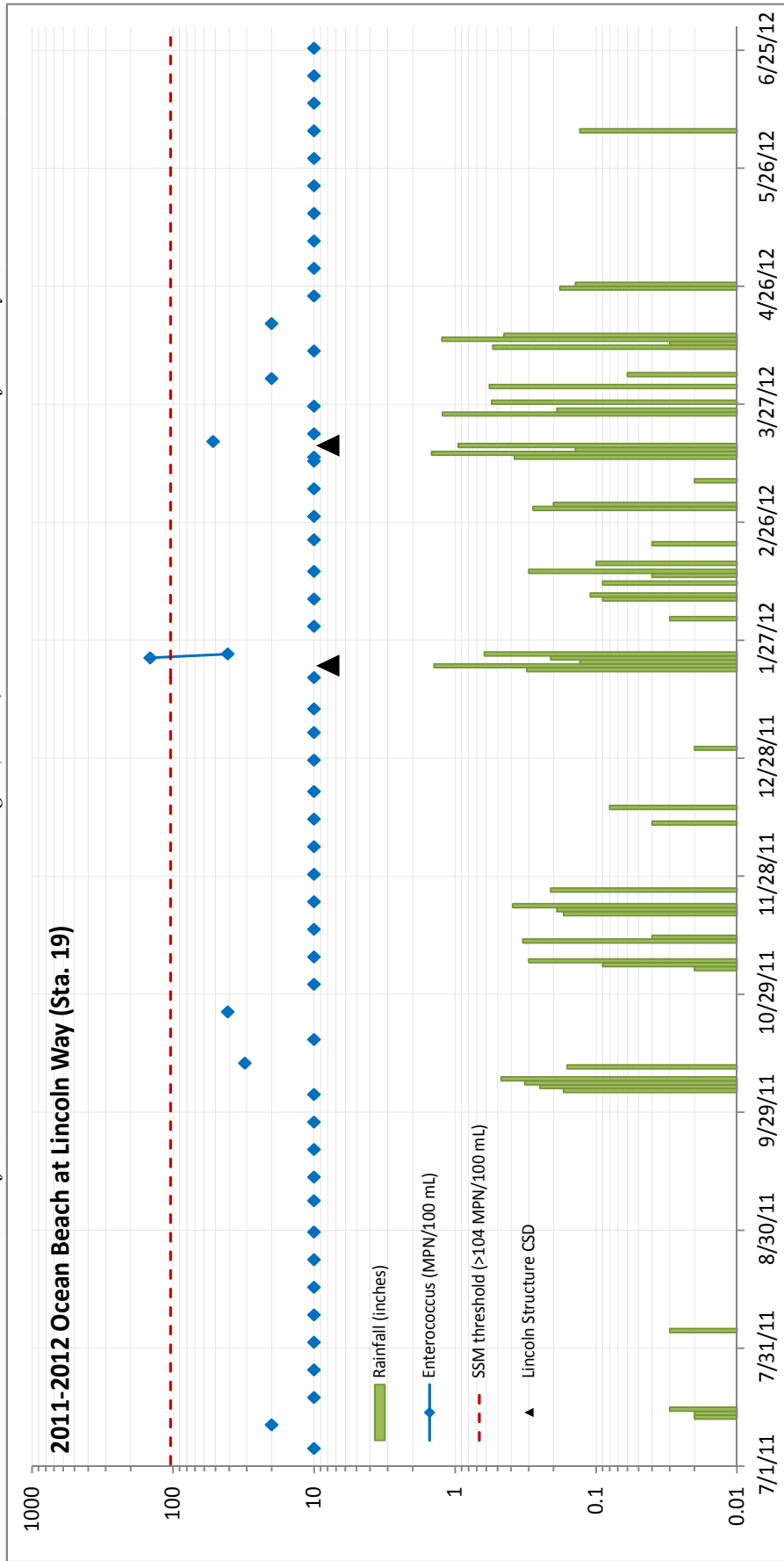
Appendix C-11c  
*Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 19 from July 1, 2010-June 30, 2011*



lines connecting enterococcus values indicate samples collected on consecutive days

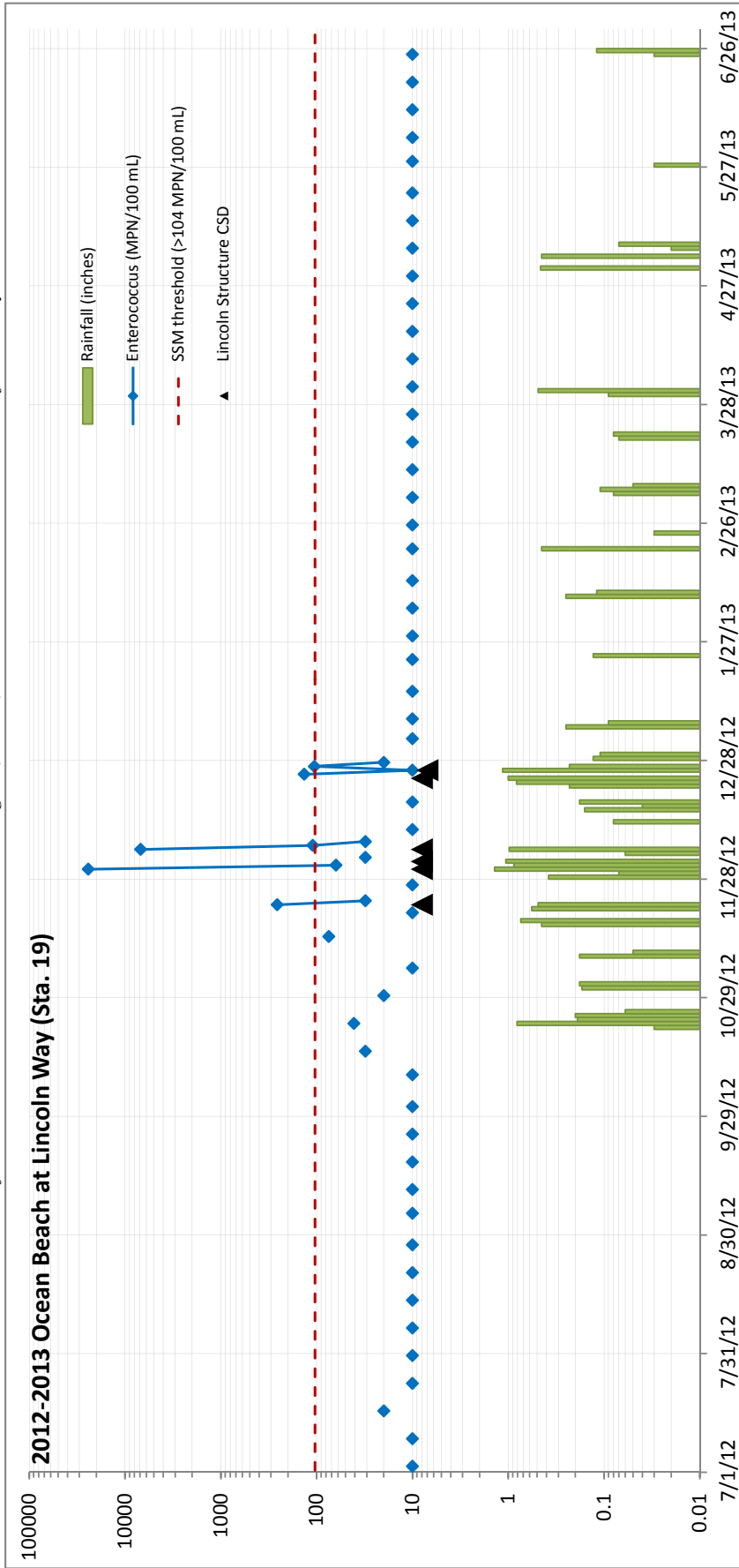
Appendix C-11d

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 19 from July 1, 2011-June 30, 2012



lines connecting enterococcus values indicate samples collected on consecutive days

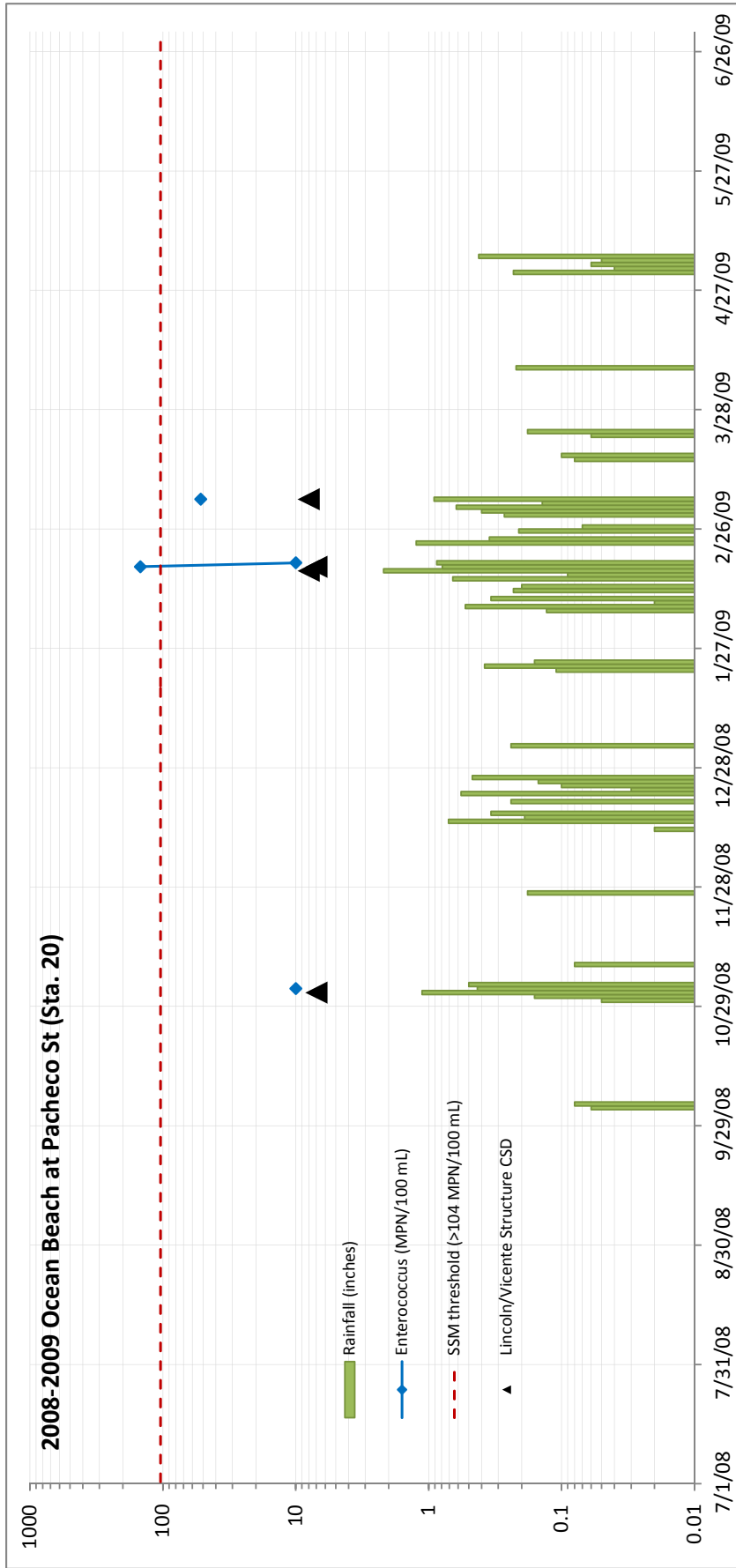
Appendix C-11e  
 Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 19 from July 1, 2012-June 30, 2013





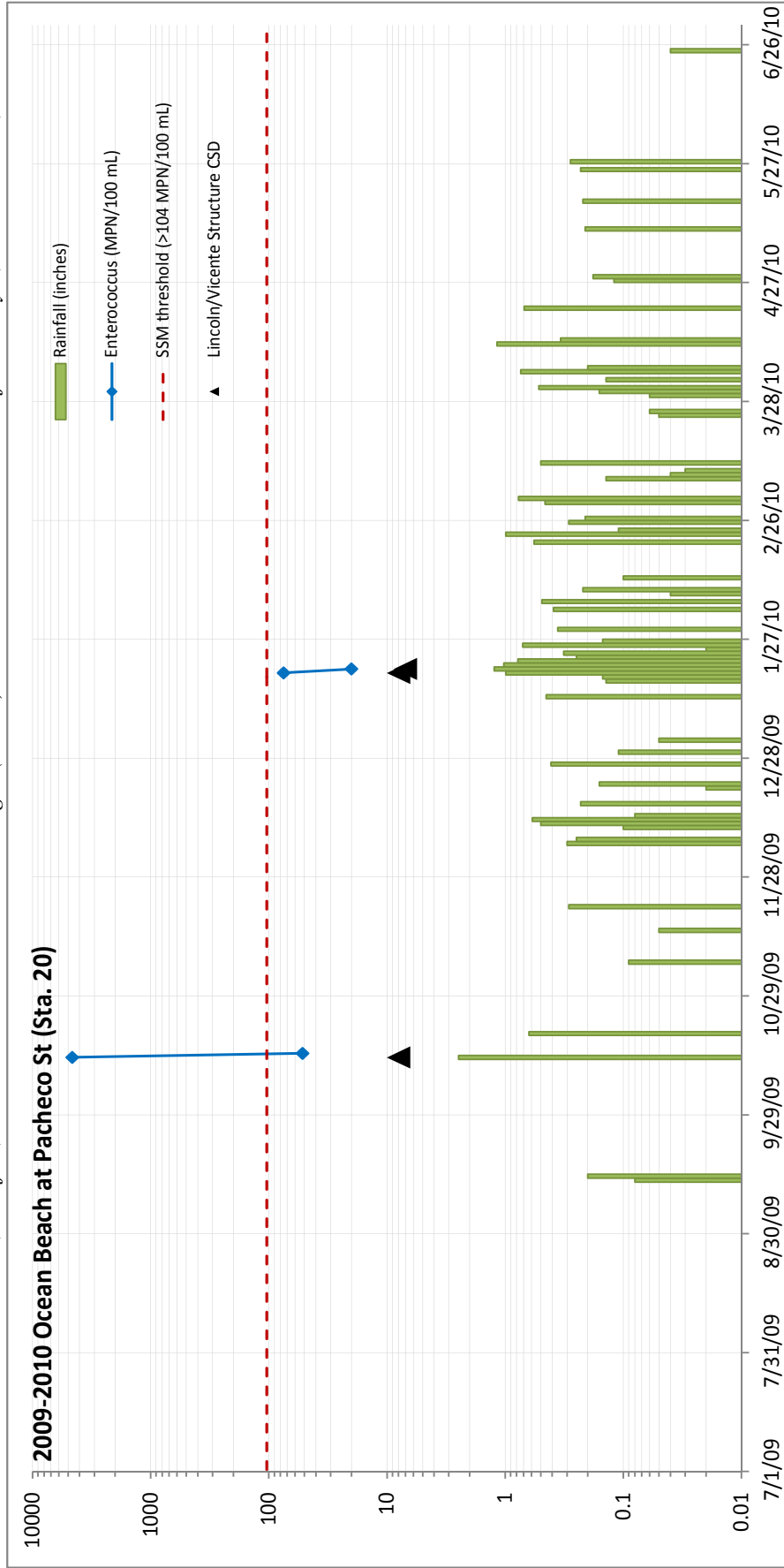
Appendix C-12a

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 20 from July 1, 2008-June 30, 2009



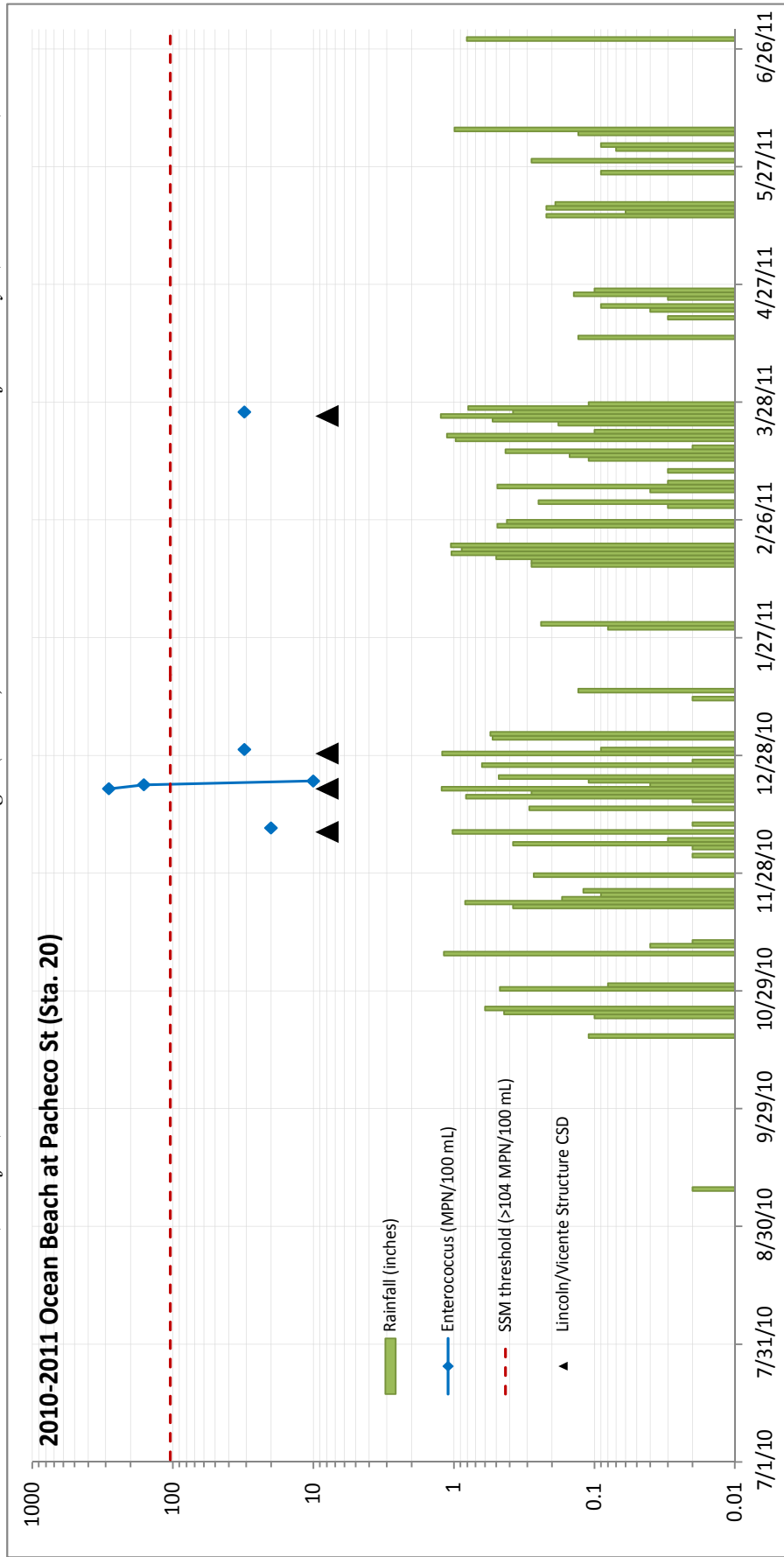
Appendix C-12b

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 20 from July 1, 2009-June 30, 2010

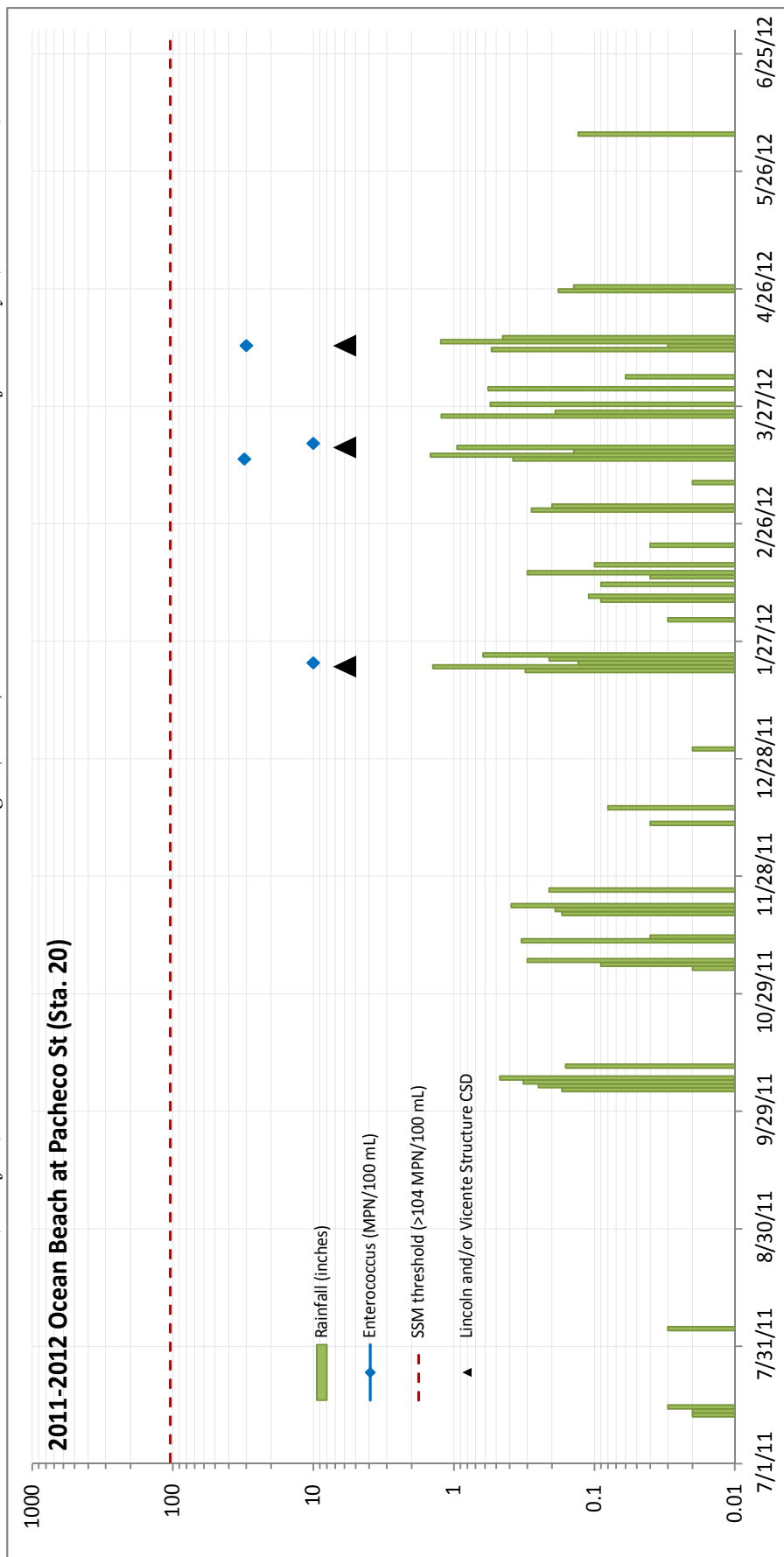


Appendix C-12c

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 20 from July 1, 2010-June 30, 2011

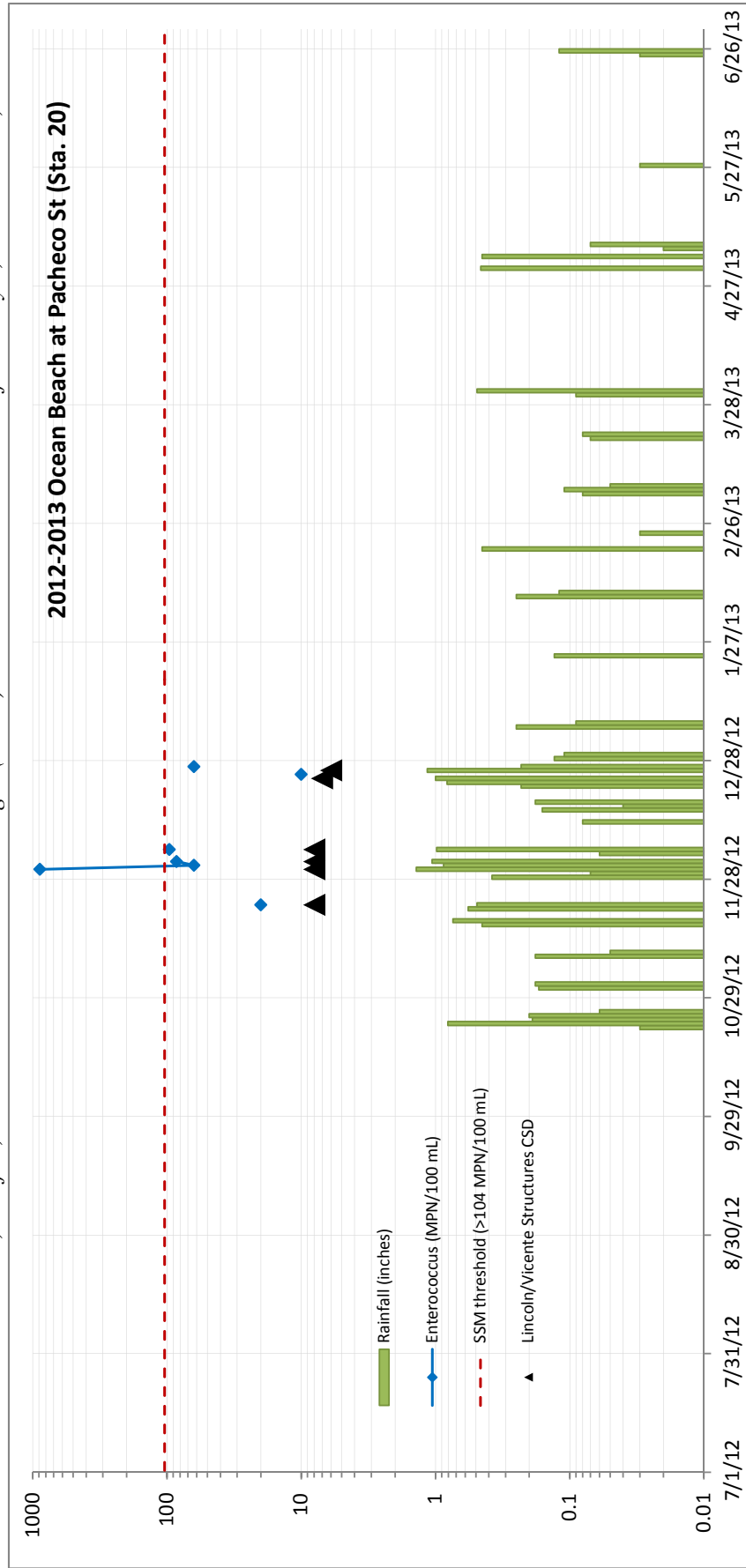


Appendix C-12d  
 Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 20 from July 1, 2011-June 30, 2012



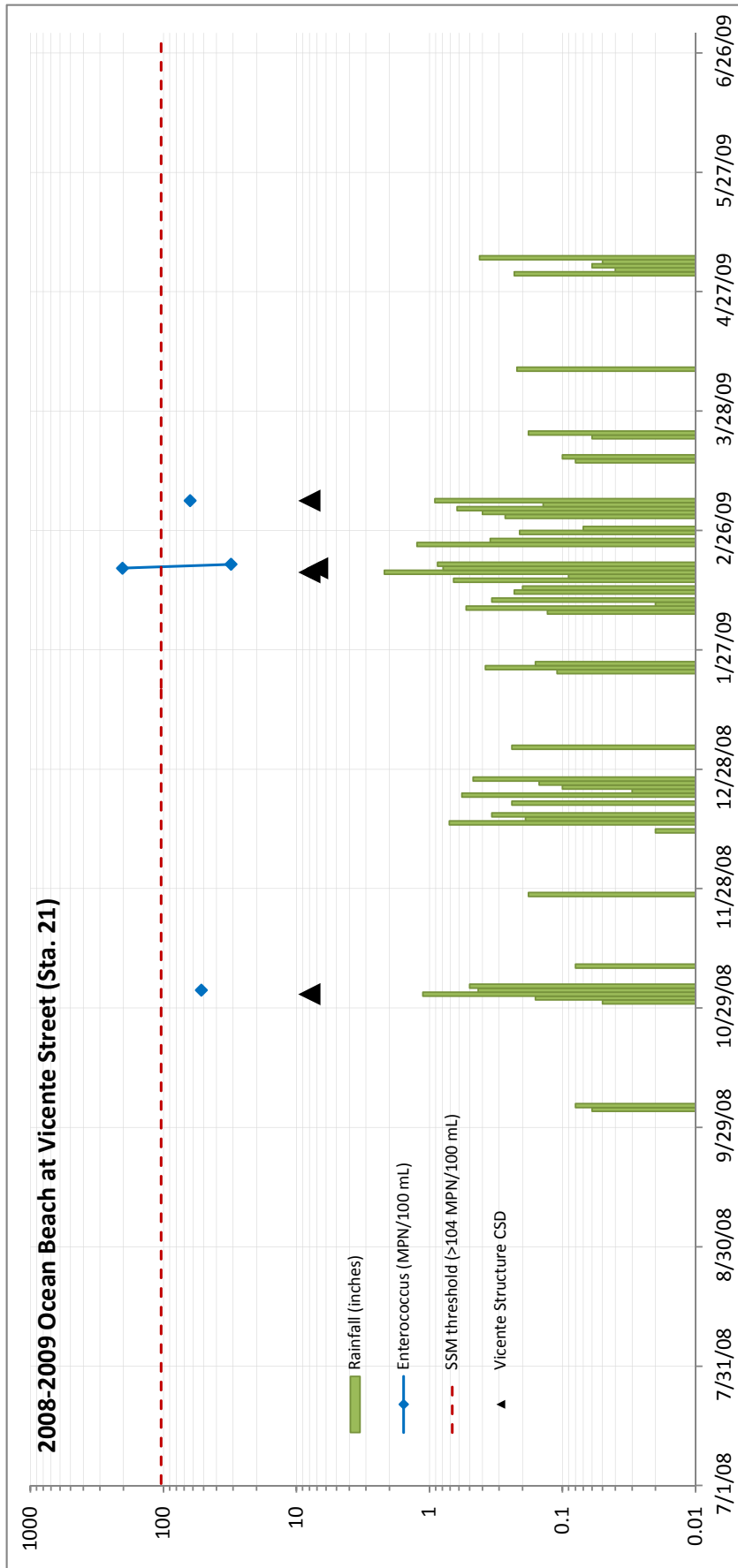
Appendix C-12e

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 20 from July 1, 2012-June 30, 2013

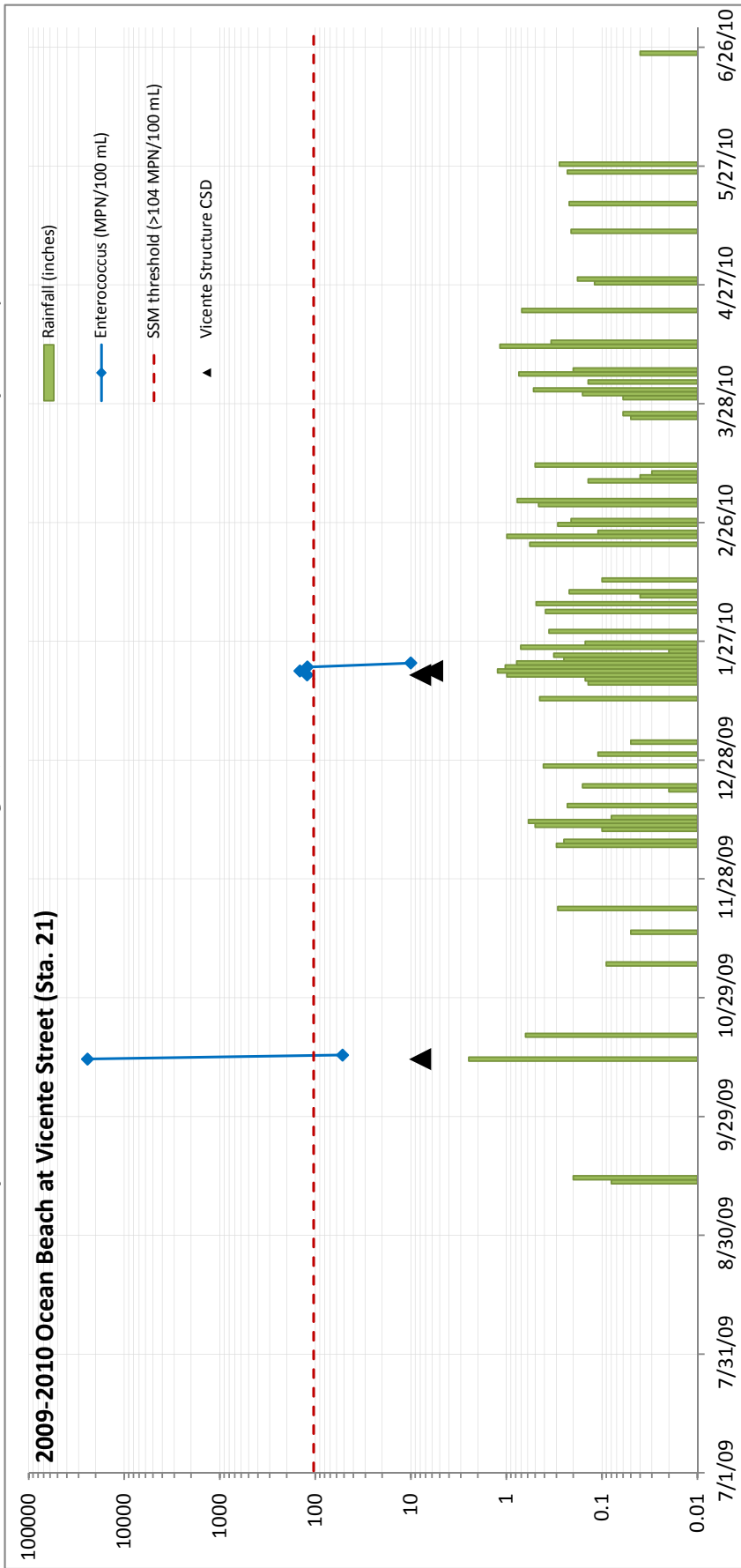


Appendix C-13a

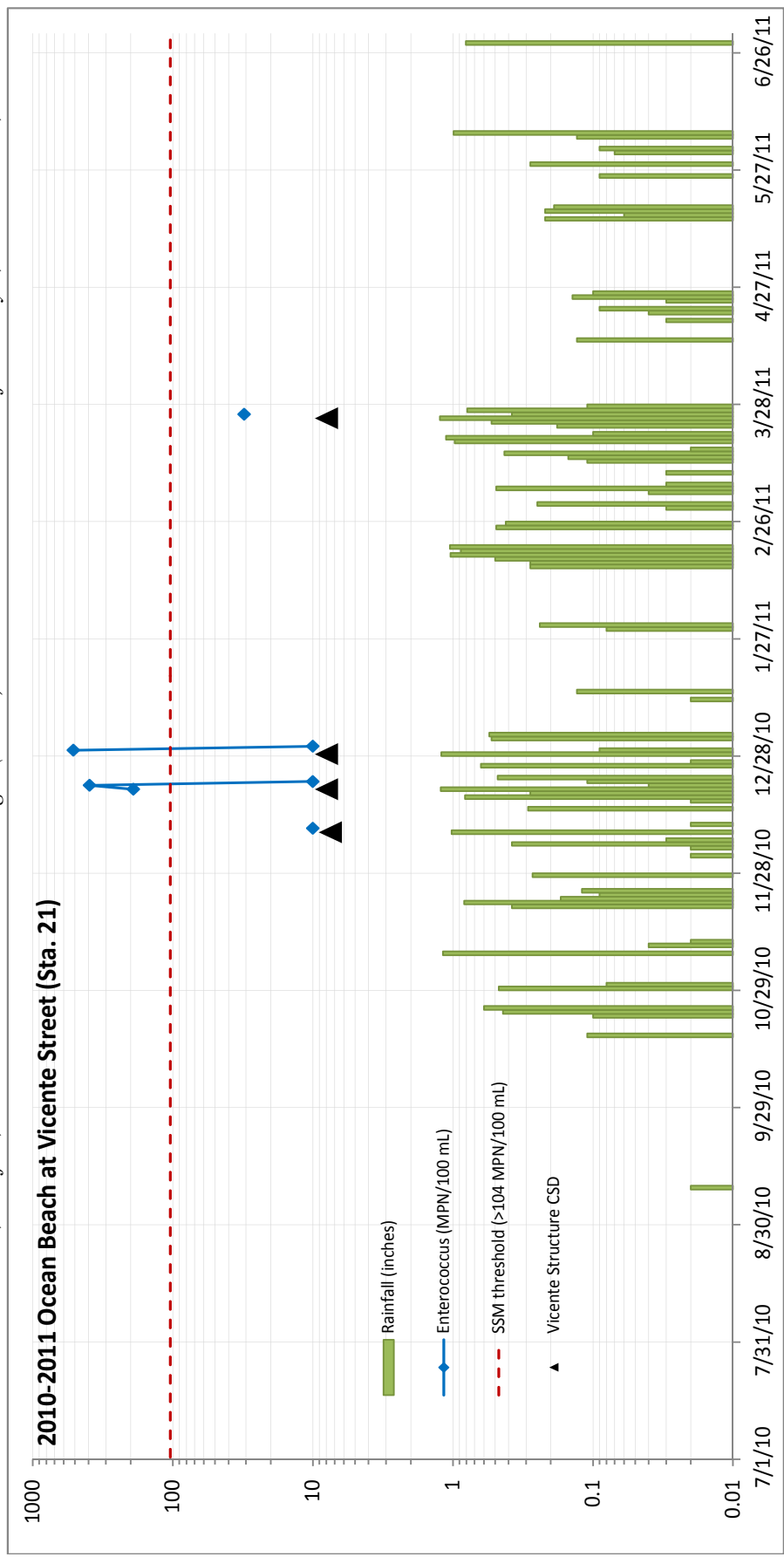
Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 21 from July 1, 2008-June 30, 2009



Appendix C-13b  
*Enterococcus* values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 21 from July 1, 2009-June 30, 2010

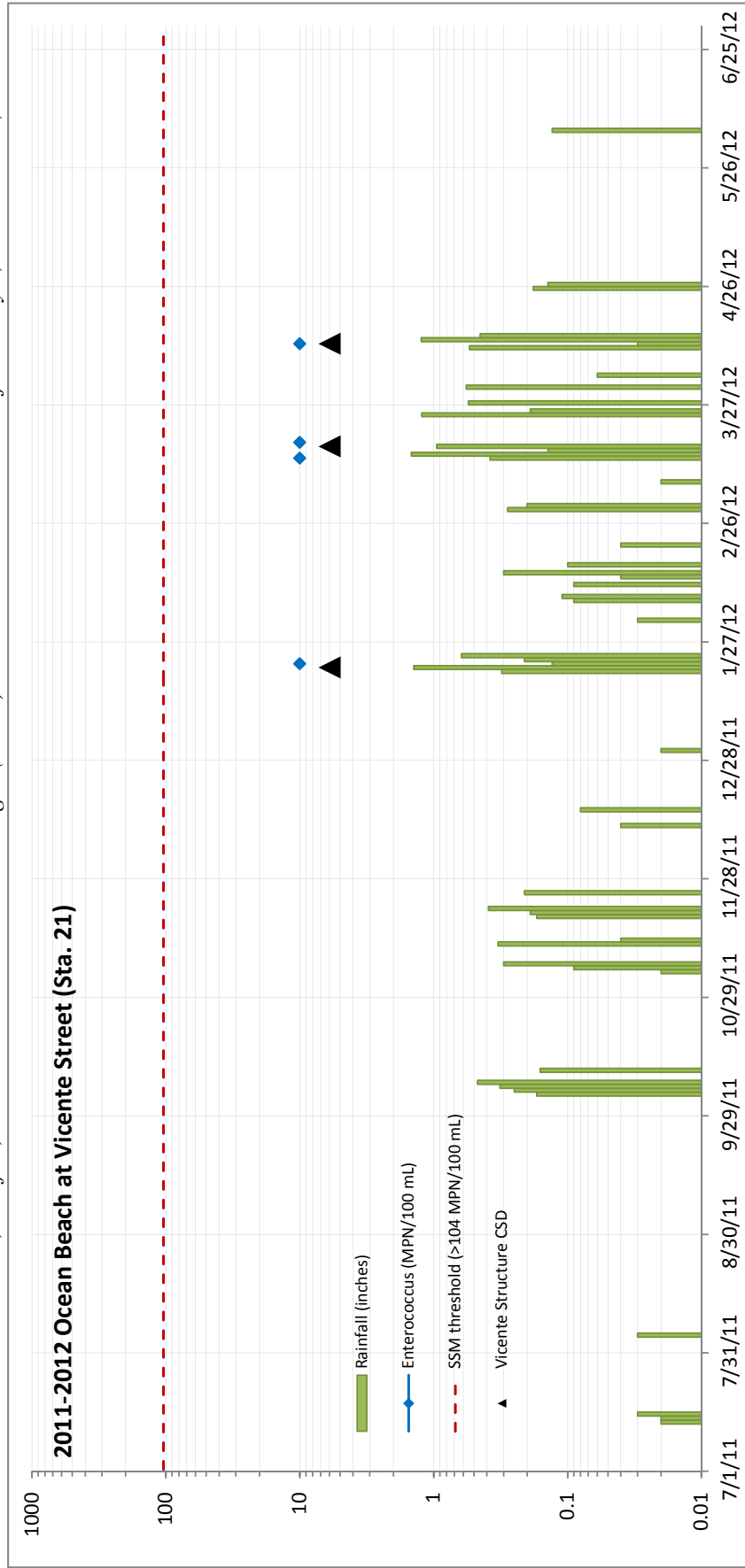


Appendix C-13c  
*Enterococcus* values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 21 from July 1, 2010-June 30, 2011

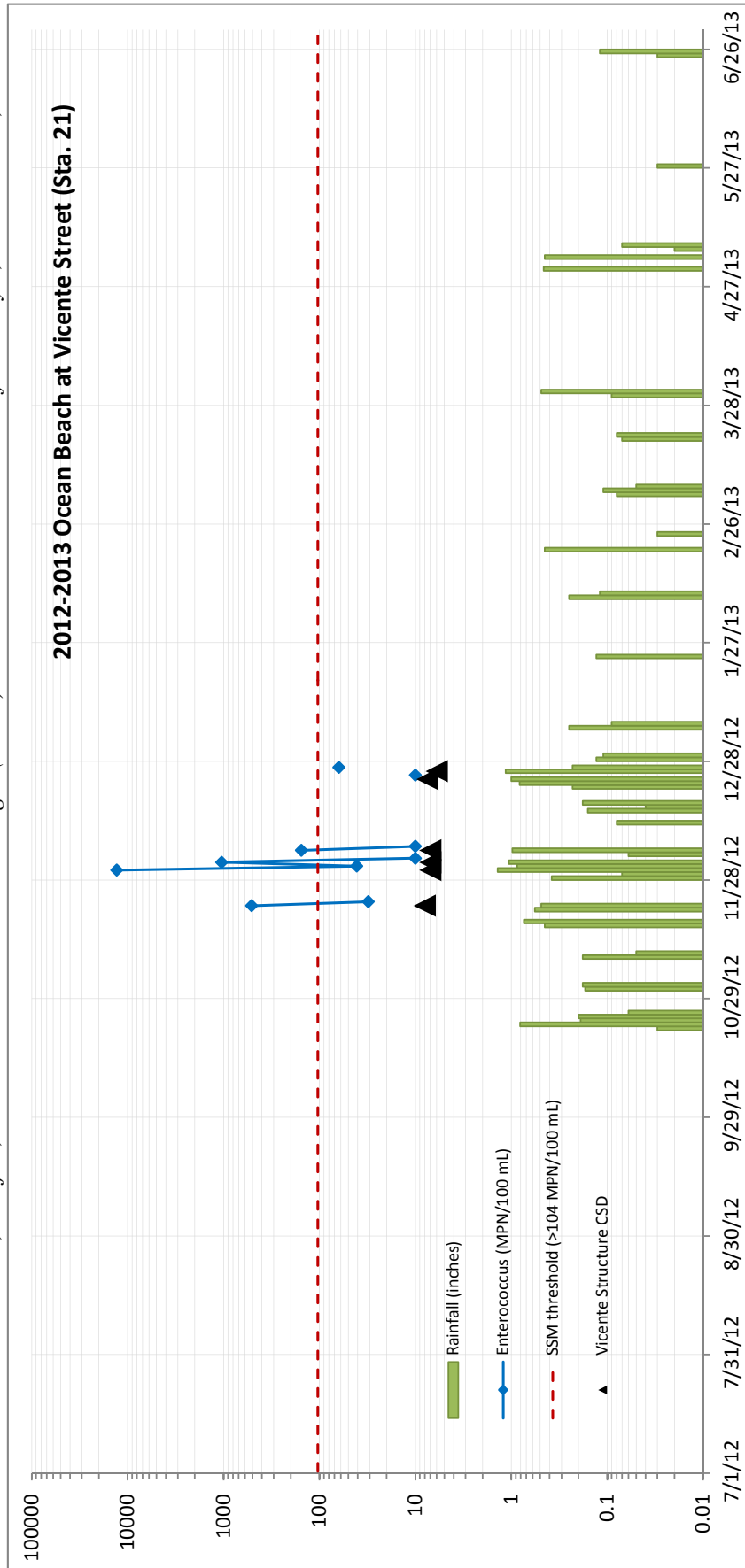




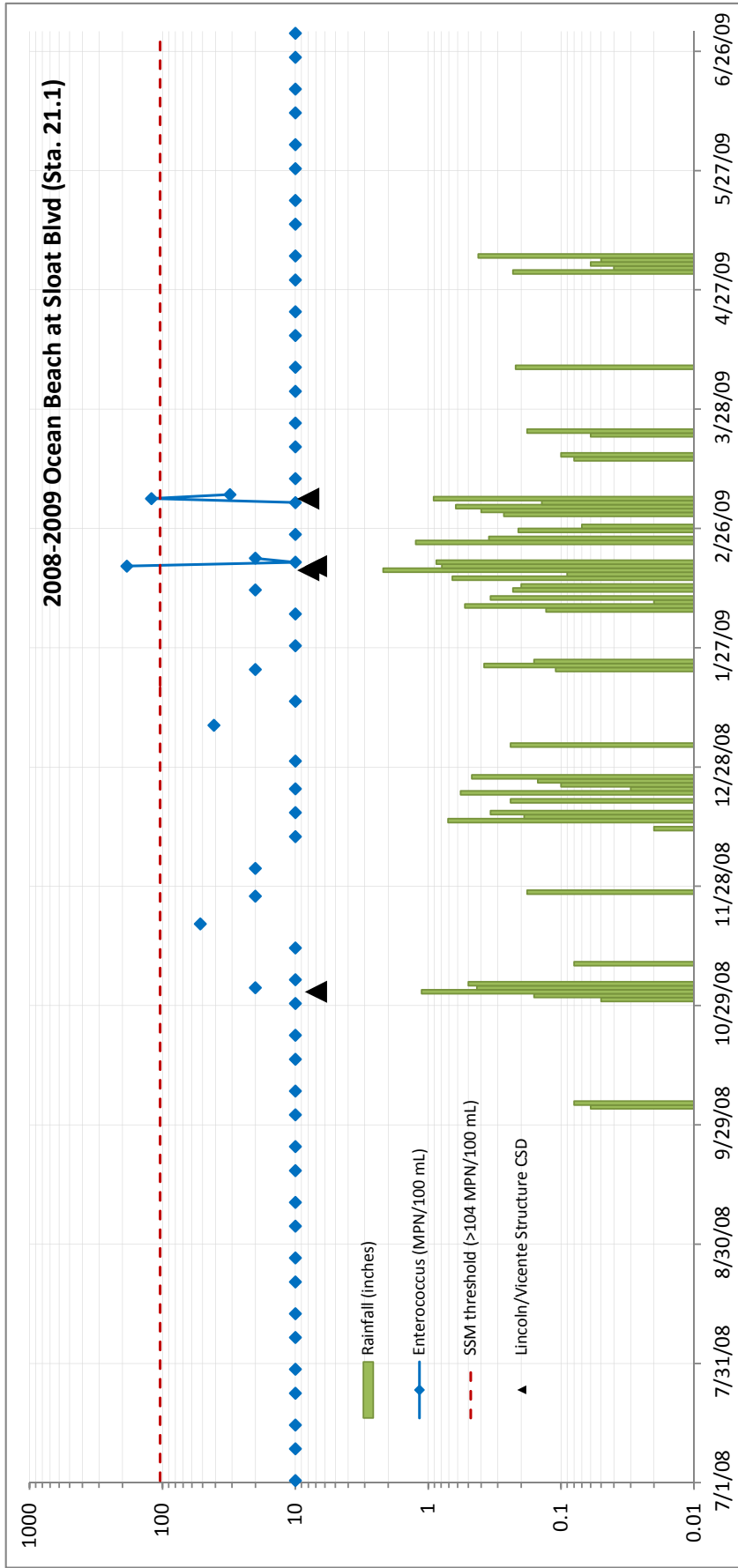
Appendix C-13d  
 Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 21 from July 1, 2011-June 30, 2012



Appendix C-13e  
 Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 21 from July 1, 2012-June 30, 2013

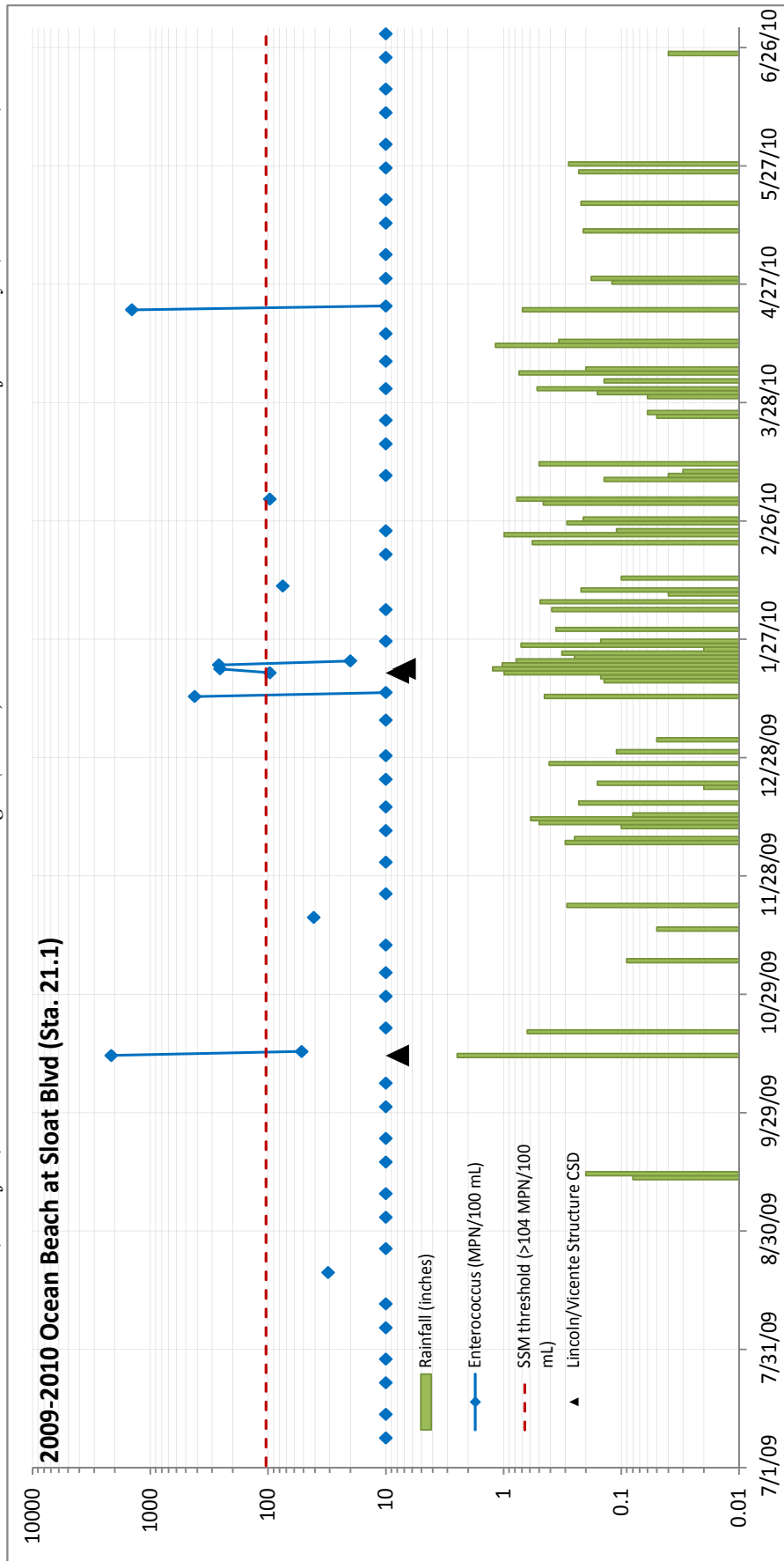


Appendix C-14a  
 Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 21.1 from July 1, 2008-June 30, 2009



Appendix C-14b

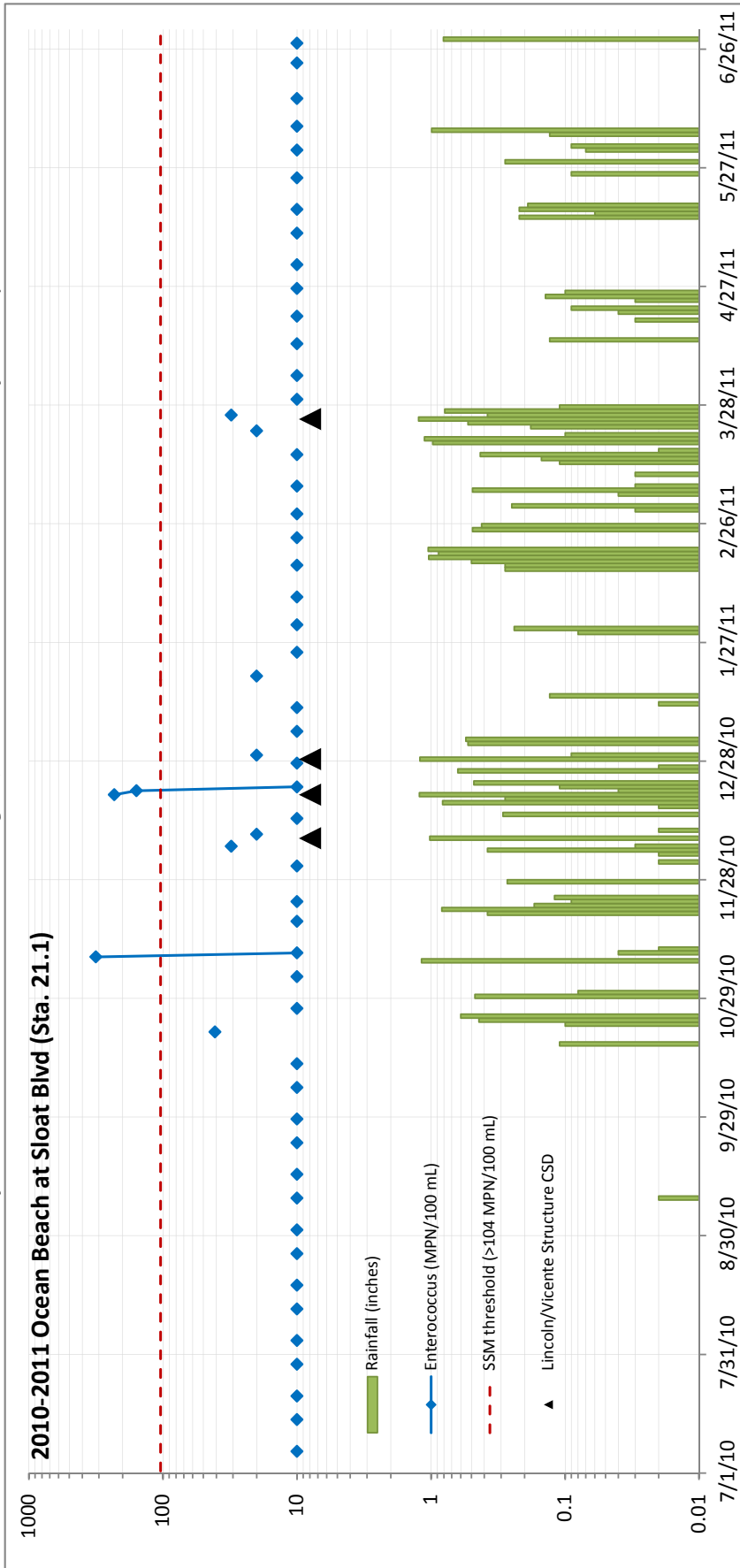
Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 21.1 from July 1, 2009-June 30, 2010



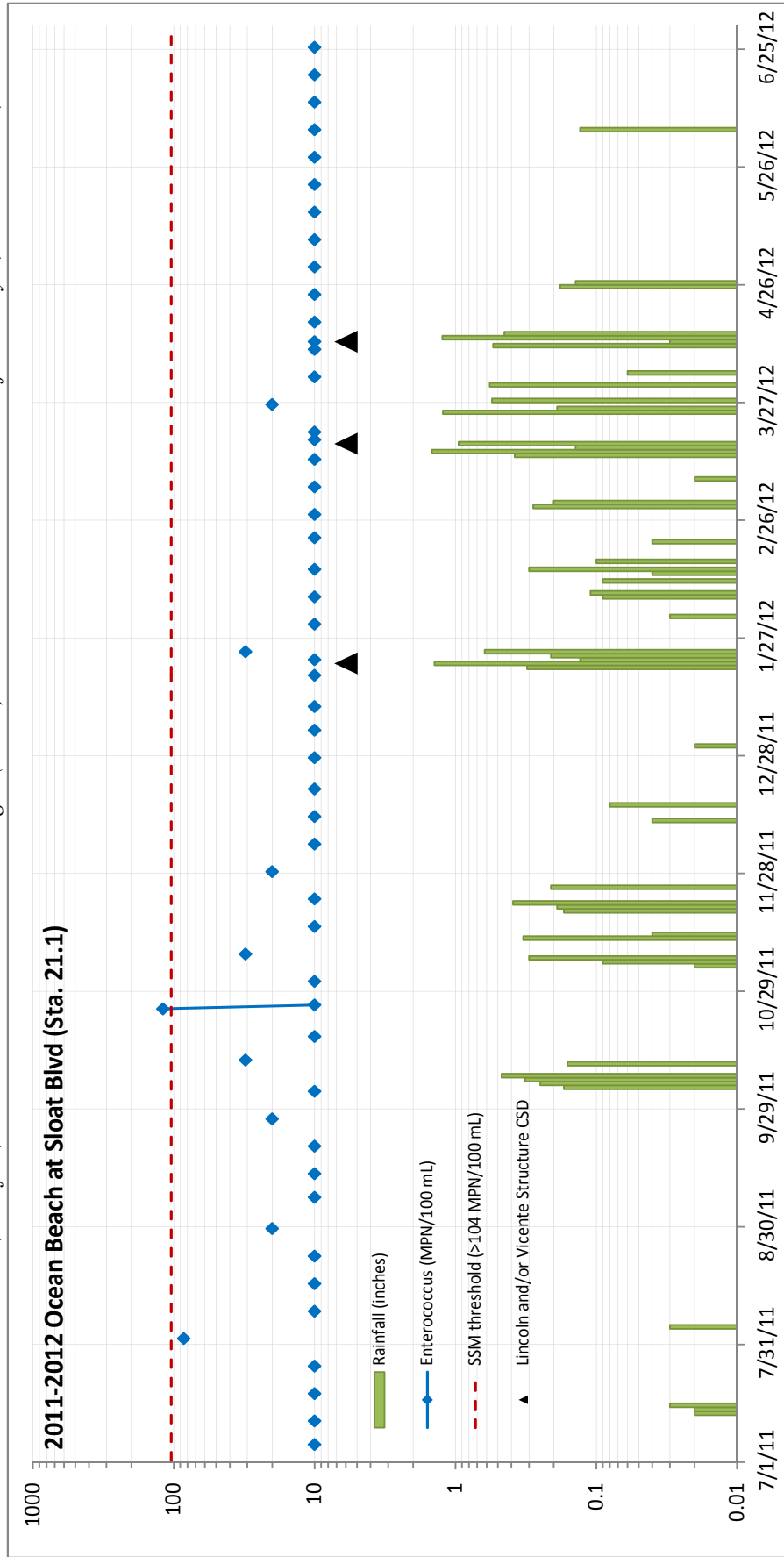
lines connecting enterococcus values indicate samples collected on consecutive days

Appendix C-14c

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 21.1 from July 1, 2010-June 30, 2011



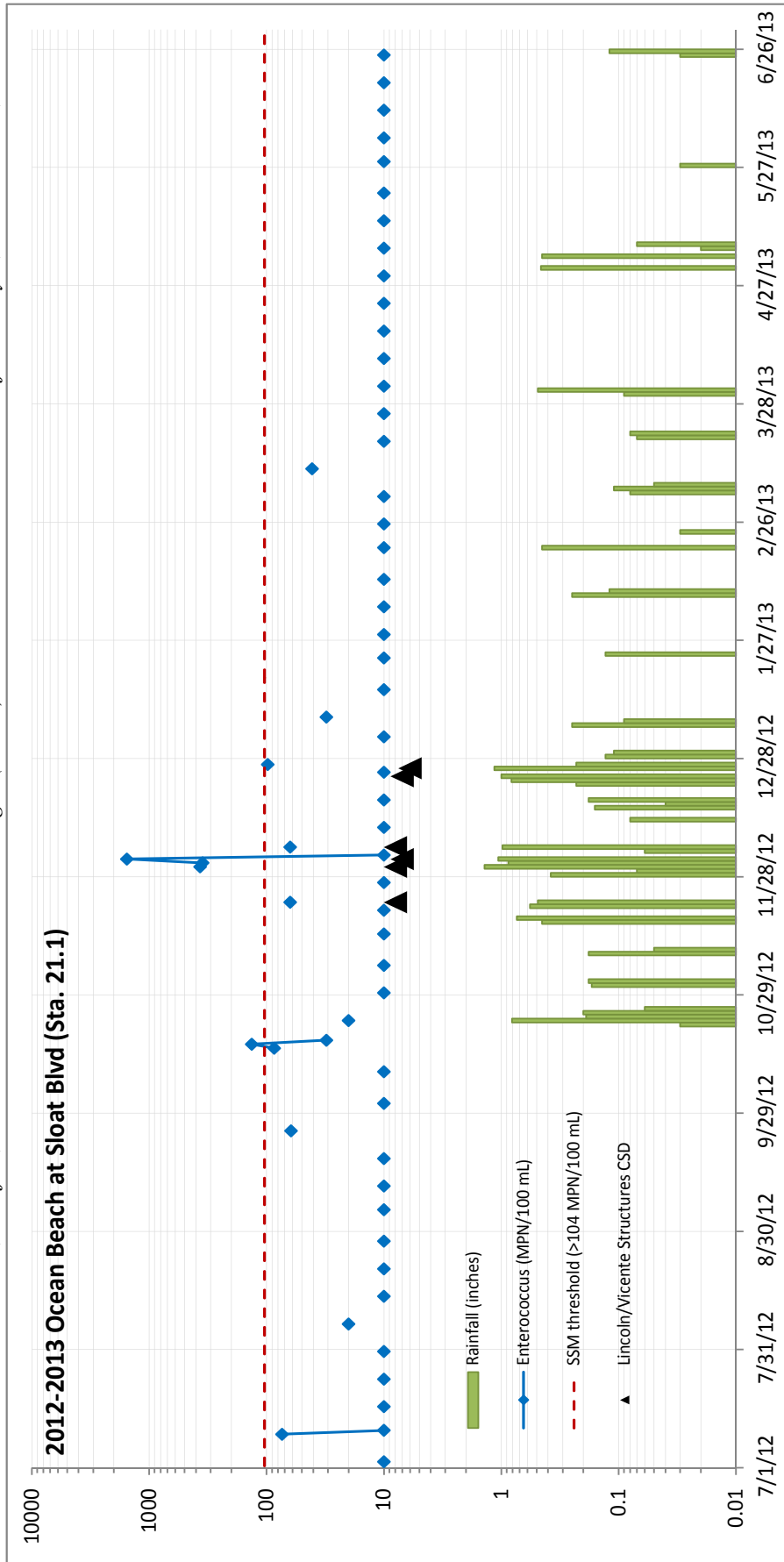
Appendix C-14d  
 Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 21.1 from July 1, 2011-June 30, 2012



lines connecting enterococcus values indicate samples collected on consecutive days

Appendix C-14e

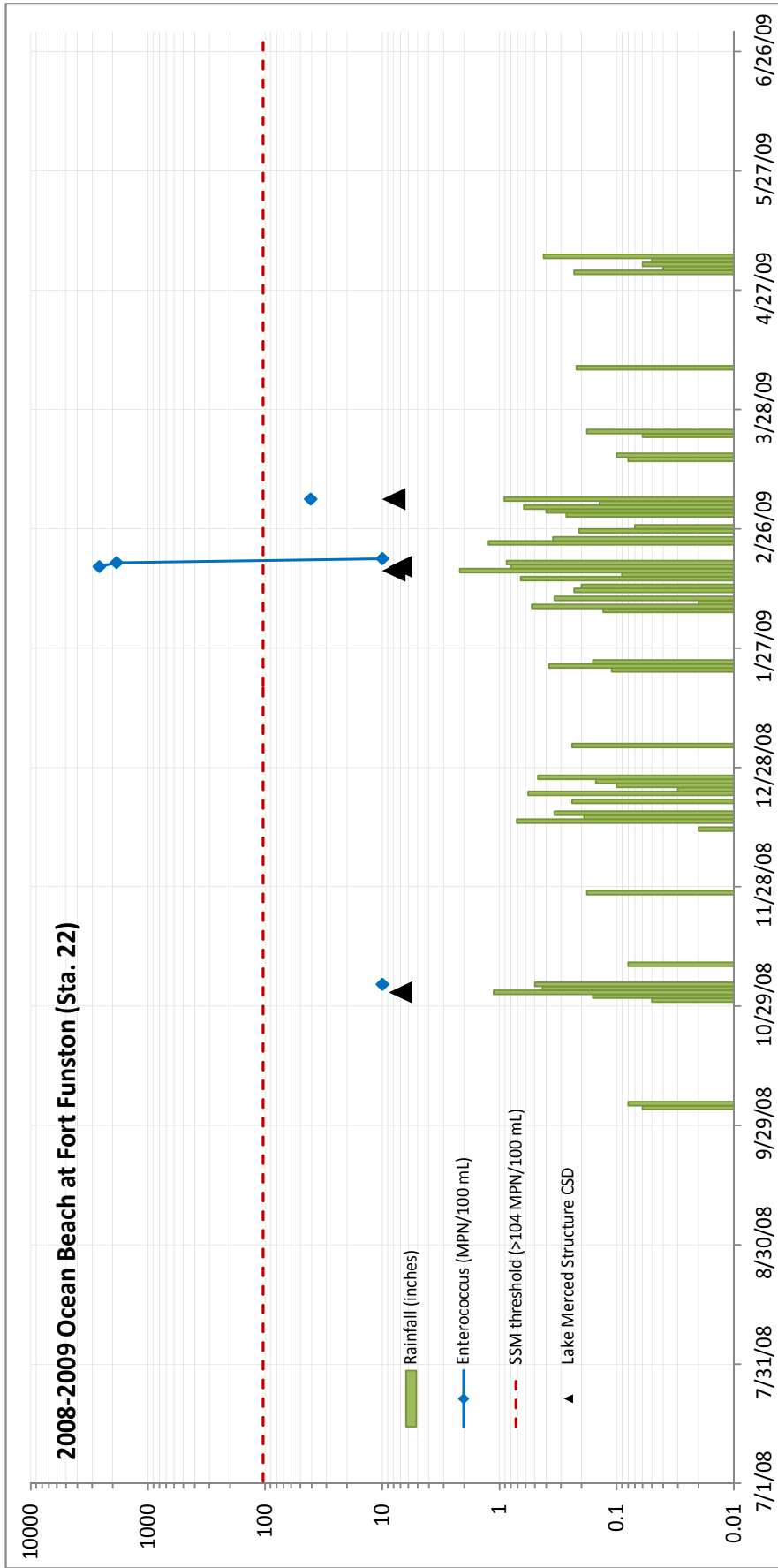
*Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 21.1 from July 1, 2012-June 30, 2013*



lines connecting enterococcus values indicate samples collected on consecutive days

Appendix C-15a

Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 22 from July 1, 2008-June 30, 2009

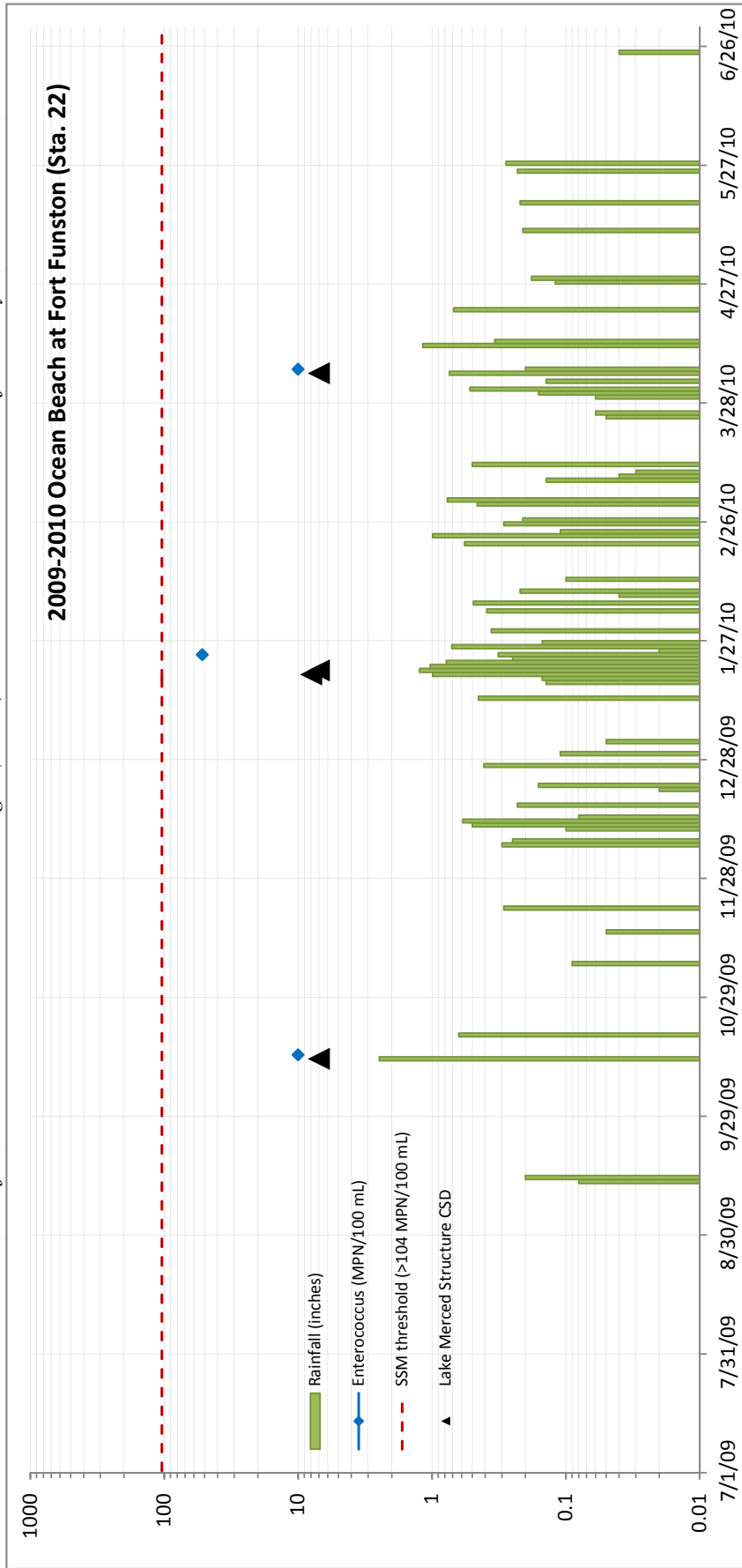


lines connecting enterococcus values indicate samples collected on consecutive days

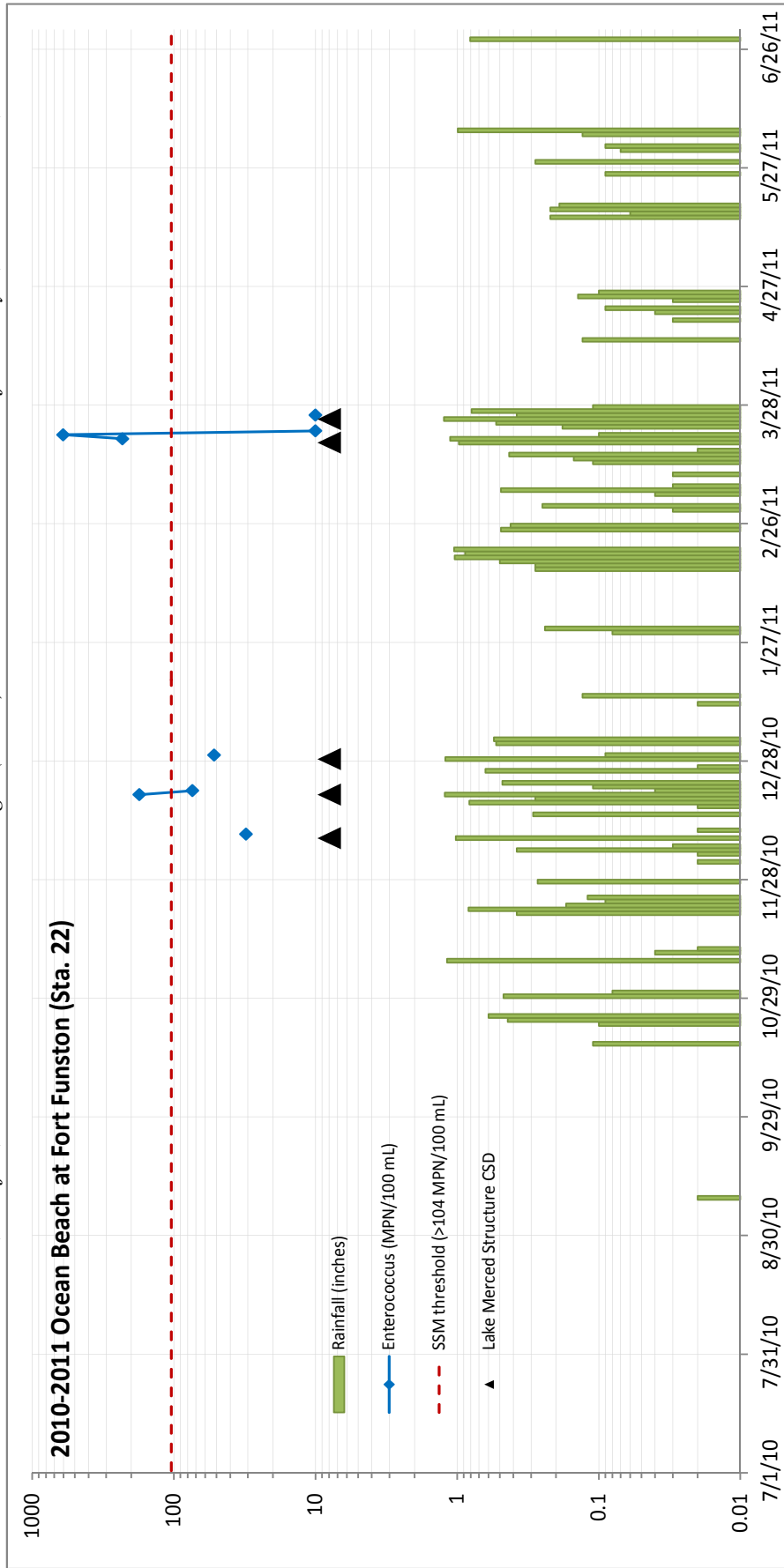


Appendix C-15b

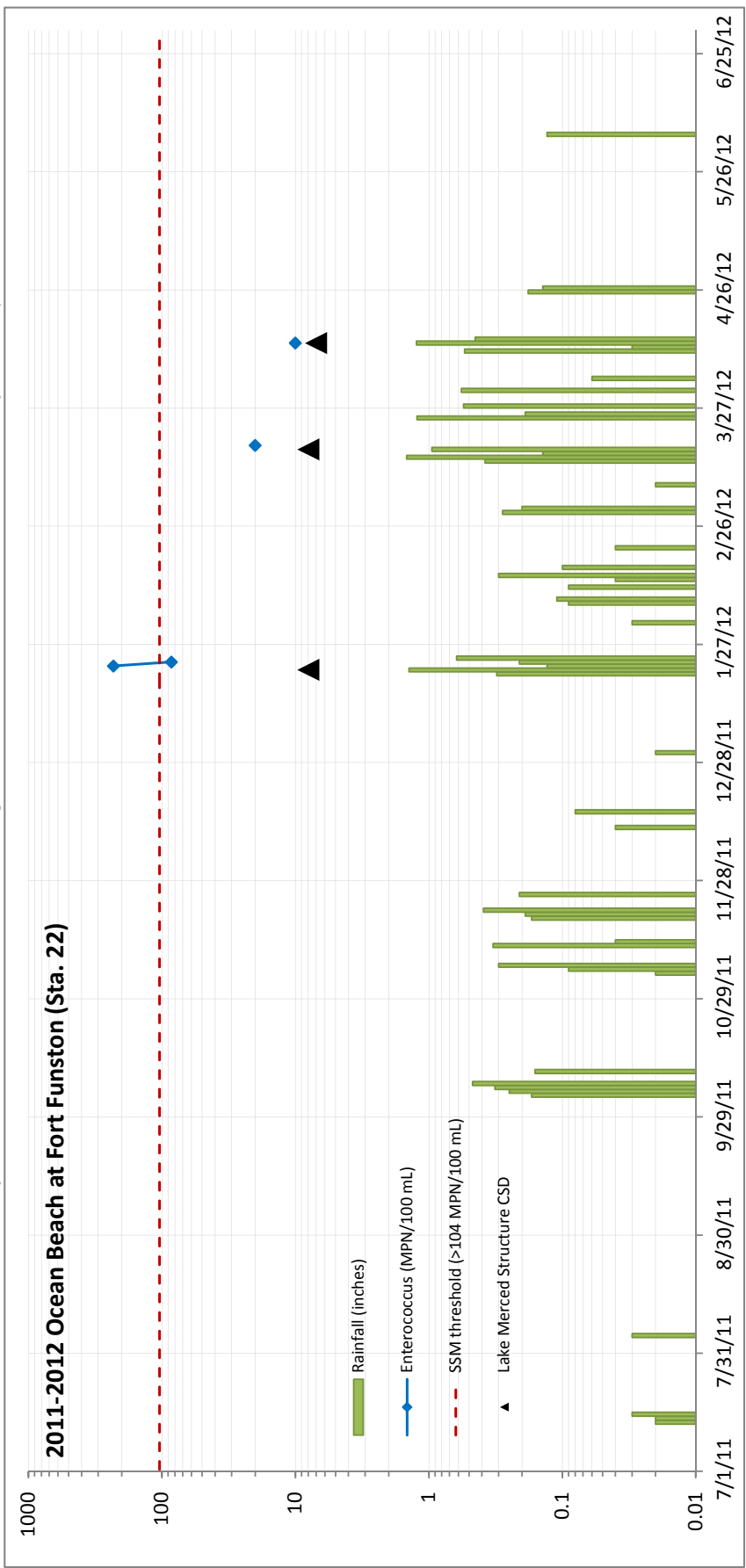
Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 22 from July 1, 2009-June 30, 2010



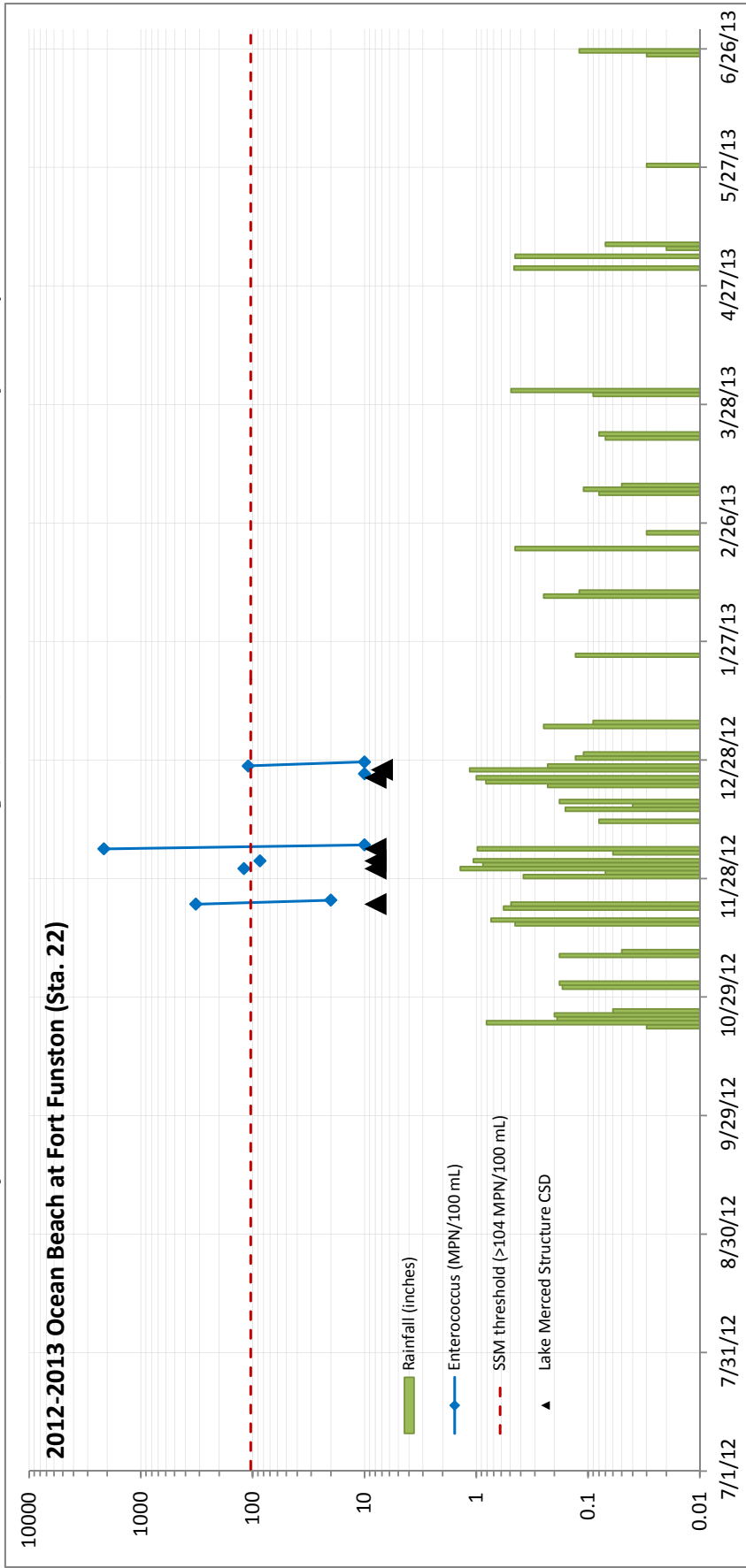
Appendix C-15c  
*Enterococcus* values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 22 from July 1, 2010-June 30, 2011

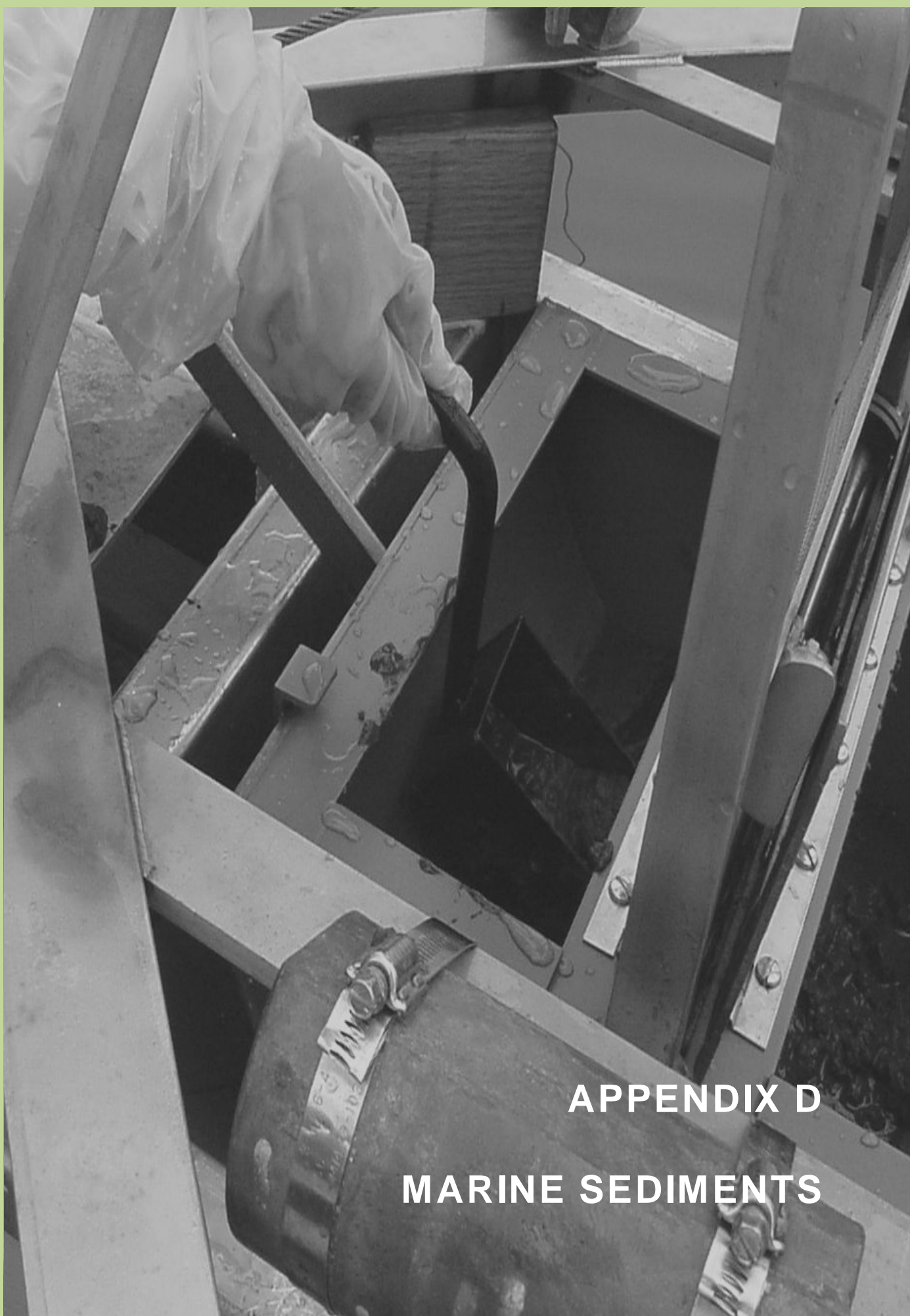


Appendix C-15d  
 Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 22 from July 1, 2011-June 30, 2012



Appendix C-15e  
 Enterococcus values, rainfall, and Combined Sewer Discharges (CSDs) at Ocean Beach Station 22 from July 1, 2012-June 30, 2013





**APPENDIX D**

**MARINE SEDIMENTS**

APPENDIX D  
MARINE SEDIMENTS  
1997 to 2012

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Appendix D-1  
Sediment grain size percentage  
September 2012

Station	Percent Pebble (Gravel) Phi <-2 to -2	Percent Granual (Gravel) Phi >-2 to -1	Percent Very Coarse Sand Phi -1 to 0	Percent Coarse Sand Phi 0 to 1	Percent Medium Sand Phi 1 to 2	Percent Fine Sand Phi 2 to 3	Percent Very Fine Sand Phi 3 to 4	Percent Silt/Clay Phi 4 to > 8	Mean Phi
1	0	0	0	0.1	1.7	52.3	34.3	11.6	2.2
2	0	0	0	1.6	3.6	79.4	11.2	4.2	1.8
4	0	0	0	0.2	2	31.9	50.7	15.2	2.4
6	0	0	0.1	0.9	2.1	45.2	47.1	4.6	2.3
25	0	0	0.1	0.3	3.4	56.3	34.3	5.6	2.1
28	0	0	0	0.1	0.8	18.5	73.4	7.2	2.5
31	0	0	0	0.1	1.3	93.5	4.9	0.2	1.8
32	0	0	0	0.8	6.5	44.6	31.7	16.4	2.4
33	0	0	0	1.5	4.9	38.7	44.8	10.1	2.4
34	0	0	0	1.7	5.2	10.7	52.5	29.9	2.8
35	0	0.8	5.2	7.9	7.6	8.5	29.8	40.2	2.4
36	0	0	0	0.2	2.1	18.8	62.7	16.2	2.7
37	0	0	0	0.7	3.6	50.7	32.5	12.5	2.2
38	0	0	0	0.1	2.3	12.8	72.4	12.4	2.7
39	0	0	0	0.2	1.5	15.8	71.9	10.6	2.5
40	0	0.3	1.7	5.9	8.2	44.1	26.5	13.3	1.9
43	0	0	0	0.1	3.9	94.3	1.6	0.1	1.7
45	0	0	0	0.6	24.4	60.5	11.6	2.9	1.5
47	0	0	0.1	1	14.7	68	14.8	1.4	1.9
48	0	0	0	1.6	29.1	55	10	4.3	1.5
50	0	0	0	0.2	2.5	25.4	63.4	8.5	2.5
51	0	0	0	0.6	8.6	86.1	4.6	0.1	1.7
52	0	0	0	0.2	5.6	90.3	3.8	0.1	1.7
53	0	0	0	0.1	1.1	21.3	63.8	13.7	2.5
54	0	0	0	0.3	1.4	87.2	10.7	0.4	1.8
55	0	0	0	0.7	3.4	76.2	16.4	3.3	2.0
56	0	0	0	0.2	1.4	17.2	54.5	26.7	2.8
57	0	0	0	0.7	2.2	68.2	25.2	3.7	2.1
58	0	0	0	0.1	1.5	20.8	44.8	32.8	2.8
59	0	0	0	0.3	1.8	16.7	61.6	19.6	2.7
60	0	0	0	0.2	0.9	21.1	64.1	13.7	2.5
61	0	0	0	0.3	1.1	25.4	60.6	12.6	2.5
62	0	0	0	0.1	0.7	14.7	72.2	12.3	2.7
63	0	0	0	0.1	0.8	20.7	66.6	11.8	2.5
64	0	0	0	0.1	1.1	22	66.4	10.4	2.5
65	0	0	0	0.1	1.2	34.7	54.8	9.2	2.4
66	0	0	0	0.1	0.8	13.6	69	16.5	2.1
67	0	0	0	0.1	0.8	28.5	63.1	7.5	2.4
68	0	0	0	0.2	1.2	13.8	76.4	8.4	2.7
69	0	0	0	0.2	0.9	21	71.8	6.1	2.9
70	0	0	0	0.1	1.2	36.8	54.1	7.8	2.4
71	0	0	0	0.1	0.6	26.5	66	6.8	2.5
72	0	0	0	0.2	2.2	17.2	66.6	13.8	2.5
73	0	0	0	0.8	2.6	76.4	17.7	2.5	2.0
75	0	0	0	1.4	2.9	72.2	19.1	4.4	2.0
77	0	0	0	0.4	2.3	54.6	36.4	6.3	2.1
78	0	0	0	0.4	1.3	27.7	60.6	10	2.5
79	0	0	0	0.2	1.8	20.8	62.4	14.8	2.5
80	0	1.1	0	1.6	8.6	77.6	11.8	0.4	1.8

Appendix D-2  
*Sediment grain size summary statistics*  
*September 2012*

Station	Mean Phi	Standard Deviation	Skewness	Kurtosis
1	2.2	0.665	0.553	0.854
2	1.8	0.41	0.094	2.561
4	2.4	0.665	-0.207	0.897
6	2.3	0.533	-0.353	0.587
25	2.1	0.616	0.553	0.815
28	2.5	0.564	-0.300	2.316
31	1.8	0.239	0.249	1.47
32	2.4	0.93	0.412	1.061
33	2.4	0.764	-0.278	1.084
34	2.8	0.915	-0.131	1.253
35	2.4	1.548	-0.478	1.046
36	2.7	0.747	-0.036	2.211
37	2.2	0.679	0.549	0.852
38	2.7	0.454	-0.017	2.464
39	2.5	0.577	-0.271	2.404
40	1.9	1.099	0.019	1.411
43	1.7	0.164	0.000	0.738
45	1.5	0.645	-0.240	1.042
47	1.9	0.586	0.244	2.472
48	1.5	0.675	-0.212	0.99
50	2.5	0.62	-0.281	0.951
51	1.7	0.289	-0.288	1.737
52	1.7	0.257	-0.264	1.56
53	2.5	0.62	-0.257	2.20
54	1.8	0.29	0.293	1.781
55	2.0	0.474	0.572	1.785
56	2.8	0.789	0.000	0.968
57	2.1	0.51	0.553	0.668
58	2.8	0.827	-0.027	0.876
59	2.7	0.758	-0.008	2.184
60	2.5	0.618	-0.257	2.210
61	2.5	0.633	-0.251	0.980
62	2.7	0.448	-0.008	2.408
63	2.5	0.606	-0.264	2.258
64	2.5	0.608	-0.271	2.234
65	2.4	0.634	-0.225	0.878
66	2.9	0.582	0.240	2.350
67	2.4	0.607	-0.276	0.914
68	2.7	0.429	-0.029	2.448
69	2.5	0.568	-0.319	2.214
70	2.4	0.626	-0.224	0.855
71	2.5	0.593	-0.293	0.93
72	2.5	0.61	-0.266	2.316
73	2.0	0.471	0.570	1.76
75	2.0	0.501	0.572	1.737
77	2.1	0.623	0.545	0.822
78	2.5	0.628	-0.257	0.939
79	2.5	0.629	-0.257	2.185
80	1.8	0.429	-0.004	2.64



Appendix D-3  
*Sediment chemical analyses*  
*September 2012*

Station	Percent Total Volatile Solids	Percent Total Solids	Total Organic Carbon (mg/Kg)	Organic Carbon Percent	Total Kjeldalh Nitrogen (mg/Kg)
1	1.70	78.8	1427	1.43	241
2	1.76	80.5	1309	1.31	348
4	2.67	70.3	1938	1.94	368
6	2.73	71.8	2040	2.04	406
25	2.57	68.1	1481	1.48	329
28	2.15	66.3	1861	1.86	331
31	1.65	79.0	1019	1.02	182
32	3.05	69.0	6849	6.85	513
33	2.60	65.9	3906	3.91	391
34	3.14	61.8	4732	4.73	487
35	9.40	52.9	8719	8.72	867
36	2.26	67.6	2674	2.67	351
37	2.24	72.9	2043	2.04	286
38	2.43	68.5	2898	2.90	344
39	2.27	70.3	3026	3.03	308
40	3.79	64.0	7186	7.19	548
43	1.15	78.2	701	0.70	116
45	1.32	82.3	1206	1.21	167
47	1.67	74.3	856	0.86	215
48	1.68	74.1	2338	2.34	244
50	2.12	70.7	1920	1.92	305
51	1.18	80.1	522	0.52	112
52	1.23	79.0	635	0.64	124
53	2.21	68.2	2440	2.44	315
54	1.46	74.0	1054	1.05	197
55	1.81	70.5	2443	2.44	349
56	2.59	65.1	3261	3.26	404
57	1.98	67.8	2653	2.65	338
58	2.17	69.3	2953	2.95	321
59	2.10	71.1	3894	3.89	416
60	2.04	71.0	3530	3.53	378
61	1.92	71.8	1687	1.69	366
62	1.87	66.4	1485	1.49	324
63	2.26	67.7	1829	1.83	383
64	1.88	73.3	1935	1.94	367
65	1.18	69.7	1451	1.45	310
66	1.78	71.1	1893	1.89	337
67	1.62	76.0	1739	1.74	271
68	1.48	71.0	1639	1.64	267
69	1.81	67.8	1625	1.63	291
70	1.91	74.4	1882	1.88	309
71	1.79	69.6	1512	1.51	277
72	2.17	69.1	2103	2.10	306
73	1.77	69.3	1526	1.53	311
75	1.90	71.5	2710	2.71	480
77	1.94	71.9	2376	2.38	314
78	2.06	66.9	2181	2.18	405
79	2.12	70.3	4374	4.37	374
80	1.52	76.8	722	0.72	130

## Appendix D-4

Organic pollutants in sediment ( $\mu\text{g}/\text{Kg}$ , dry weight) September 2012

Organic Pollutants	RL*	PCB	Station					
		Congener #	1	2	4	6	25	28
4,4'-DDE	1		ND	ND	ND	ND	ND	ND
4,4'-DDD	1		ND	ND	ND	ND	ND	ND
4,4'-DDT	1		ND	ND	ND	ND	ND	ND
naphthalene	1		1.7	1.7	2.4	1.6	1.6	1.5
acenaphthylene	1		ND	ND	ND	ND	ND	1.9
acenaphthene	1		ND	ND	2.7	ND	6.0	ND
fluorene	1		ND	ND	ND	ND	ND	4.4
phenanthrene	1		18.1	3.6	18.6	1.5	40.5	49.8
anthracene	1		6.9	ND	4.3	ND	5.5	34.2
Fluoranthene	1		28.1	8.0	34.5	2.4	59.2	78.5
Pyrene	1		33.8	9.6	40.3	3.3	66.9	98.0
Benz[a]anthracene	1		10.4	2.8	10.6	ND	13.7	40.0
Chrysene	1		11.8	3.5	10.6	2.0	15.7	43.2
Benzo[b]fluoranthene	1		11.8	4.2	19.1	4.1	16.4	25.8
Benzo[k]fluoranthene	1		10.4	2.8	14.2	2.7	14.4	27.7
Benzo[e]pyrene	1		11.1	3.5	17.0	3.4	14.4	24.5
Perylene	1		8.3	4.2	15.6	3.4	11.6	18.1
Benzo[a]pyrene	1		17.3	4.2	24.1	3.4	25.3	45.2
Indeno[1,2,3-cd]pyrene	1		13.8	4.2	23.4	4.7	19.8	27.7
dibenz[a,h]anthracene	1		2.1	ND	2.8	ND	2.1	3.9
benzo[ghi]perylene	1		14.5	4.9	24.8	4.7	20.5	27.1
2,4'-Dichlorobiphenyl	1	PCB 008	ND	ND	ND	ND	ND	ND
2,2',5'-Trichlorobiphenyl	1	PCB 018	ND	ND	ND	ND	ND	ND
2,4,4'-Trichlorobiphenyl	1	PCB 028	ND	ND	ND	ND	ND	ND
2,2',5,5'-Tetrachlorobiphenyl	1	PCB 052	ND	ND	ND	ND	ND	ND
2,2',3,5'-Tetrachlorobiphenyl	1	PCB 044	ND	ND	ND	ND	ND	ND
2,3',4,4'-Tetrachlorobiphenyl	1	PCB 066	ND	ND	ND	ND	ND	ND
2,2',4,5,5'-Pentachlorobiphenyl	1	PCB 101	ND	ND	ND	ND	ND	ND
3,4,4',5'-Tetrachlorobiphenyl	1	PCB 081	ND	ND	ND	ND	ND	ND
3,3',4,4'-Tetrachlorobiphenyl	1	PCB 077	ND	ND	ND	ND	ND	ND
2',3,4,4',5'-Pentachlorobiphenyl	1	PCB 123	ND	ND	ND	ND	ND	ND
2,3',4,4',5'-Pentachlorobiphenyl	1	PCB 118	ND	ND	ND	ND	ND	ND
2,2',4,4',5,5'-Hexachlorobiphenyl	1	PCB 153	ND	ND	ND	ND	ND	ND
3,3',4,5,5'-Pentachlorobiphenyl	1	PCB 127	ND	ND	ND	ND	ND	ND
2,3,3',4,4'-Pentachlorobiphenyl	1	PCB 105	ND	ND	ND	ND	ND	ND
2,2',3,4,4',5'-Hexachlorobiphenyl	1	PCB 137	ND	ND	ND	ND	ND	ND
3,3',4,4',5'-Pentachlorobiphenyl	1	PCB 126	ND	ND	ND	ND	ND	ND
2,2',3,4',5,5',6-Heptachlorobiphenyl	1	PCB 187	ND	ND	ND	ND	ND	ND
2,2',3,3',4,4'-Hexachlorobiphenyl	1	PCB 128	ND	ND	ND	ND	ND	ND
2,3,3',4,4',5'-Hexachlorobiphenyl	1	PCB 157	ND	ND	ND	ND	ND	ND
2,2',3,4,4',5,5'-Heptachlorobiphenyl	1	PCB 180	ND	ND	ND	ND	ND	ND
2,2',3,3',4,4',5-Heptachlorobiphenyl	1	PCB 170	ND	ND	ND	ND	ND	ND
2,3,3',4,4',5,5'-Heptachlorobiphenyl	1	PCB 189	ND	ND	ND	ND	ND	ND
2,2',3,3',4,4',5,6-Octachlorobiphenyl	1	PCB 195	ND	ND	ND	ND	ND	ND
2,2',3,3',4,4',5,5',6-Nonachlorobiphenyl	1	PCB 206	ND	ND	ND	ND	ND	ND
Decachlorobiphenyl	1	PCB 209	ND	ND	ND	ND	ND	ND
2,4',5'-Trichlorobiphenyl	1	PCB 31	ND	ND	ND	ND	ND	ND
2',3,4'-Trichlorobiphenyl	1	PCB 33	ND	ND	ND	ND	ND	ND
2,2',4,5'-Tetrachlorobiphenyl	1	PCB 49	ND	ND	ND	ND	ND	ND
2,4,4',5'-Tetrachlorobiphenyl	1	PCB 74	ND	ND	ND	ND	ND	ND
2,3',4',5'-Tetrachlorobiphenyl	1	PCB 70	ND	ND	ND	ND	ND	ND
2,2',3,5',6-Pentachlorobiphenyl	1	PCB 95	ND	ND	ND	ND	ND	ND
2,3,3',4'-Tetrachlorobiphenyl	1	PCB 56	ND	ND	ND	ND	ND	ND
2,3,4,4'-Tetrachlorobiphenyl	1	PCB 60	ND	ND	ND	ND	ND	ND
2,2',4,4',5-Pentachlorobiphenyl	1	PCB 99	ND	ND	ND	ND	ND	ND
2,2',3',4,5-Pentachlorobiphenyl	1	PCB 97	ND	ND	ND	ND	ND	ND
2,2',3,4,5'-Pentachlorobiphenyl	1	PCB 87	ND	ND	ND	ND	ND	ND
2,3,3',4',6-Pentachlorobiphenyl	1	PCB 110	ND	ND	ND	ND	ND	ND
2,3,4,4',5-Pentachlorobiphenyl	1	PCB 114	ND	ND	ND	ND	ND	ND
2,2',3,5,5',6-Hexachlorobiphenyl	1	PCB 151	ND	ND	ND	ND	ND	ND
2,2',3,4',5',6-Hexachlorobiphenyl	1	PCB 149	ND	ND	ND	ND	ND	ND
2,2',3,3',4,6'-Hexachlorobiphenyl	1	PCB 132	ND	ND	ND	ND	ND	ND
2,2',3,4,5,5'-Hexachlorobiphenyl	1	PCB 141	ND	ND	ND	ND	ND	ND
2,3,3',4,4',6-Hexachlorobiphenyl	1	PCB 158	ND	ND	ND	ND	ND	ND
2,2',3,4,4',5',6-Heptachlorobiphenyl	1	PCB 183	ND	ND	ND	ND	ND	ND
2,3',4,4',5,5'-Hexachlorobiphenyl	1	PCB 167	ND	ND	ND	ND	ND	ND
2,2',3,3',4,5,6'-Heptachlorobiphenyl	1	PCB 174	ND	ND	ND	ND	ND	ND
2,2',3,3',4',5,6-Heptachlorobiphenyl	1	PCB 177	ND	ND	ND	ND	ND	ND
2,3,3',4,4',5-Hexachlorobiphenyl	1	PCB 156	ND	ND	ND	ND	ND	ND
3,3',4,4',5,5'-Hexachlorobiphenyl	1	PCB 169	ND	ND	ND	ND	ND	ND
2,2',3,3',4,5,5',6-Octachlorobiphenyl	1	PCB 201	ND	ND	ND	ND	ND	ND
2,2',3,4,4',5,5',6-Octachlorobiphenyl	1	PCB 203	ND	ND	ND	ND	ND	ND
2,2',3,3',4,4',5,5'-Octachlorobiphenyl	1	PCB 194	ND	ND	ND	ND	ND	ND
<b>Sum of organics</b>			200.0	56.8	265.0	37.3	333.6	551.6

\*RL = Reporting Limit

ND = Not Detected















## Appendix D-4

Organic pollutants in sediment ( $\mu\text{g}/\text{Kg}$ , dry weight) September 2012

Organic Pollutants	RL*	PCB	Station			
		Congener #	77	78	79	80
4,4'-DDE	1		ND	ND	ND	ND
4,4'-DDD	1		ND	ND	ND	ND
4,4'-DDT	1		ND	ND	ND	ND
naphthalene	1		ND	ND	1.6	ND
acenaphthylene	1		ND	ND	2.3	ND
acenaphthene	1		ND	ND	ND	ND
fluorene	1		ND	ND	1.3	ND
phenanthrene	1		5.5	2.2	32.8	ND
anthracene	1		1.4	ND	26.1	ND
Fluoranthene	1		19.7	5.4	81.5	ND
Pyrene	1		24.3	7.2	107.3	ND
Benz[a]anthracene	1		5.9	1.7	41.1	ND
Chrysene	1		5.9	2.2	40.6	ND
Benzo[b]fluoranthene	1		7.7	3.0	21.9	ND
Benzo[k]fluoranthene	1		6.8	2.2	28.0	ND
Benzo[e]pyrene	1		7.7	2.6	20.5	ND
Perylene	1		3.6	2.6	10.7	ND
Benzo[a]pyrene	1		9.5	3.0	41.6	ND
Indeno[1,2,3-cd]pyrene	1		12.2	3.9	27.1	ND
dibenz[a,h]anthracene	1		1.4	ND	5.1	ND
benzo[ghi]perylene	1		13.5	4.3	24.7	ND
2,4'-Dichlorobiphenyl	1	PCB 008	ND	ND	ND	ND
2,2',5'-Trichlorobiphenyl	1	PCB 018	ND	ND	ND	ND
2,4,4'-Trichlorobiphenyl	1	PCB 028	ND	ND	ND	ND
2,2',5,5'-Tetrachlorobiphenyl	1	PCB 052	ND	ND	ND	ND
2,2',3,5'-Tetrachlorobiphenyl	1	PCB 044	ND	ND	ND	ND
2,3',4,4'-Tetrachlorobiphenyl	1	PCB 066	ND	ND	ND	ND
2,2',4,5,5'-Pentachlorobiphenyl	1	PCB 101	ND	ND	ND	ND
3,4,4',5'-Tetrachlorobiphenyl	1	PCB 081	ND	ND	ND	ND
3,3',4,4'-Tetrachlorobiphenyl	1	PCB 077	ND	ND	ND	ND
2',3,4,4',5-Pentachlorobiphenyl	1	PCB 123	ND	ND	ND	ND
2,3',4,4',5-Pentachlorobiphenyl	1	PCB 118	ND	ND	ND	ND
2,2',4,4',5,5'-Hexachlorobiphenyl	1	PCB 153	ND	ND	ND	ND
3,3',4,5,5'-Pentachlorobiphenyl	1	PCB 127	ND	ND	ND	ND
2,3,3',4,4'-Pentachlorobiphenyl	1	PCB 105	ND	ND	ND	ND
2,2',3,4,4',5-Hexachlorobiphenyl	1	PCB 137	ND	ND	ND	ND
3,3',4,4',5-Pentachlorobiphenyl	1	PCB 126	ND	ND	ND	ND
2,2',3,4',5,5',6-Heptachlorobiphenyl	1	PCB 187	ND	ND	ND	ND
2,2',3,3',4,4'-Hexachlorobiphenyl	1	PCB 128	ND	ND	ND	ND
2,3,3',4,4',5'-Hexachlorobiphenyl	1	PCB 157	ND	ND	ND	ND
2,2',3,4,4',5,5'-Heptachlorobiphenyl	1	PCB 180	ND	ND	ND	ND
2,2',3,3',4,4',5-Heptachlorobiphenyl	1	PCB 170	ND	ND	ND	ND
2,3,3',4,4',5,5'-Heptachlorobiphenyl	1	PCB 189	ND	ND	ND	ND
2,2',3,3',4,4',5,6-Octachlorobiphenyl	1	PCB 195	ND	ND	ND	ND
2,2',3,3',4,4',5,5',6-Nonachlorobiphenyl	1	PCB 206	ND	ND	ND	ND
Decachlorobiphenyl	1	PCB 209	ND	ND	ND	ND
2,4',5-Trichlorobiphenyl	1	PCB 31	ND	ND	ND	ND
2',3,4-Trichlorobiphenyl	1	PCB 33	ND	ND	ND	ND
2,2',4,5-Tetrachlorobiphenyl	1	PCB 49	ND	ND	ND	ND
2,4,4',5-Tetrachlorobiphenyl	1	PCB 74	ND	ND	ND	ND
2,3',4',5-Tetrachlorobiphenyl	1	PCB 70	ND	ND	ND	ND
2,2',3,5',6-Pentachlorobiphenyl	1	PCB 95	ND	ND	ND	ND
2,3,3',4'-Tetrachlorobiphenyl	1	PCB 56	ND	ND	ND	ND
2,3,4,4'-Tetrachlorobiphenyl	1	PCB 60	ND	ND	ND	ND
2,2',4,4',5-Pentachlorobiphenyl	1	PCB 99	ND	ND	ND	ND
2,2',3',4,5-Pentachlorobiphenyl	1	PCB 97	ND	ND	ND	ND
2,2',3,4,5'-Pentachlorobiphenyl	1	PCB 87	ND	ND	ND	ND
2,3,3',4',6-Pentachlorobiphenyl	1	PCB 110	ND	ND	ND	ND
2,3,4,4',5-Pentachlorobiphenyl	1	PCB 114	ND	ND	ND	ND
2,2',3,5,5',6-Hexachlorobiphenyl	1	PCB 151	ND	ND	ND	ND
2,2',3,4',5',6-Hexachlorobiphenyl	1	PCB 149	ND	ND	ND	ND
2,2',3,3',4,6'-Hexachlorobiphenyl	1	PCB 132	ND	ND	ND	ND
2,2',3,4,5,5'-Hexachlorobiphenyl	1	PCB 141	ND	ND	ND	ND
2,3,3',4,4',6-Hexachlorobiphenyl	1	PCB 158	ND	ND	ND	ND
2,2',3,4,4',5',6-Heptachlorobiphenyl	1	PCB 183	ND	ND	ND	ND
2,3',4,4',5,5'-Hexachlorobiphenyl	1	PCB 167	ND	ND	ND	ND
2,2',3,3',4,5,6'-Heptachlorobiphenyl	1	PCB 174	ND	ND	ND	ND
2,2',3,3',4',5,6-Heptachlorobiphenyl	1	PCB 177	ND	ND	ND	ND
2,3,3',4,4',5-Hexachlorobiphenyl	1	PCB 156	ND	ND	ND	ND
3,3',4,4',5,5'-Hexachlorobiphenyl	1	PCB 169	ND	ND	ND	ND
2,2',3,3',4,5,5',6'-Octachlorobiphenyl	1	PCB 201	ND	ND	ND	ND
2,2',3,4,4',5,5',6-Octachlorobiphenyl	1	PCB 203	ND	ND	ND	ND
2,2',3,3',4,4',5,5'-Octachlorobiphenyl	1	PCB 194	ND	ND	ND	ND
<b>Sum of organics</b>			124.8	40.4	514.5	0.0

\*RL = Reporting Limit  
ND = Not Detected

Appendix D-5  
*Oceanside Water Pollution Control Plant Polycyclic Aromatic Hydrocarbons (PAHs)  
on Final Effluent*

2001 (µg/L)	Value	MDL
Acenaphthene	< 0.03	0.03
Acenaphthylene	< 0.14	0.14
Anthracene	< 0.01	0.01
Benzo[a]anthracene	< 0.01	0.01
Benzo[a]pyrene	< 0.01	0.01
Benzo[b]fluoranthene	< 0.01	0.01
Benzo[ghi]perylene	< 0.01	0.01
Benzo[k]fluoranthene	< 0.01	0.01
Chrysene	< 0.02	0.02
Dibenzo-[a,h]-anthracene	< 0.01	0.01
Fuoranthene	< 0.04	0.04
Fluorene	< 0.02	0.02
Indeno[1,2,3-cd]pyrene	< 0.02	0.02
Naphthalene	< 0.06	0.06
Phenanthrene	< 0.06	0.06
Pyrene	< 0.03	0.03
<b>Total PAH's</b>	<b>0.49</b>	

2002 (µg/L)	Value	MDL
Acenaphthene	< 2.24	2.24
Acenaphthylene	< 2.34	2.34
Anthracene	< 2.18	2.18
Benzo[a]anthracene	< 2.02	2.02
Benzo[a]pyrene	< 2.8	2.8
Benzo[b]fluoranthene	< 2.22	2.22
Benzo[ghi]perylene	< 3.3	3.3
Benzo[k]fluoranthene	< 2.28	2.28
Chrysene	< 2.02	2.02
Dibenzo-[a,h]-anthracene	< 2.82	2.82
Fuoranthene	< 2.08	2.08
Fluorene	< 2.42	2.42
Indeno[1,2,3-cd]pyrene	< 2.7	2.7
Naphthalene	< 1.86	1.86
Phenanthrene	< 2.24	2.24
Pyrene	< 1.94	1.94
<b>Total PAH's</b>	<b>37.46</b>	

2003 (µg/L)	Value	MDL
Acenaphthene	< 0.376	0.376
Acenaphthylene	0.2425 DNQ	0.2425
Anthracene	< 0.033	0.033
Benzo[a]anthracene	< 0.066	0.066
Benzo[a]pyrene	< 0.068	0.068
Benzo[b]fluoranthene	< 0.068	0.068
Benzo[ghi]perylene	< 0.072	0.072
Benzo[k]fluoranthene	< 0.131	0.131
Chrysene	< 0.094	0.094
Dibenzo-[a,h]-anthracene	< 0.065	0.094
Fuoranthene	< 0.189	0.189
Fluorene	< 0.097	0.097
Indeno[1,2,3-cd]pyrene	< 0.095	0.095
Naphthalene	< 0.277	0.277
Phenanthrene	< 0.096	0.096
Pyrene(DNQ)	< 0.183	0.183
<b>Total PAH's</b>	<b>2.15</b>	

2004 (µg/L)	Value	MDL
Acenaphthene	< 0.44	0.44
Acenaphthylene	< 0.28	0.28
Anthracene	< 0.04	0.04
Benzo[a]anthracene	< 0.08	0.08
Benzo[a]pyrene	< 0.08	0.08
Benzo[b]fluoranthene	< 0.08	0.08
Benzo[ghi]perylene	< 0.08	0.08
Benzo[k]fluoranthene	< 0.12	0.12
Chrysene	< 0.12	0.12
Dibenzo-[a,h]-anthracene	< 0.08	0.08
Fuoranthene	< 0.24	0.24
Fluorene	< 0.12	0.12
Indeno[1,2,3-cd]pyrene	< 0.12	0.12
Naphthalene	< 0.32	0.32
Phenanthrene	< 0.12	0.12
Pyrene	< 0.24	0.24
<b>Total PAH's</b>	<b>2.56</b>	

2005 (µg/L)	Value	MDL
Acenaphthene	< 0.33	0.33
Acenaphthylene	< 0.21	0.21
Anthracene	< 0.03	0.03
Benzo[a]anthracene	< 0.06	0.06
Benzo[a]pyrene	< 0.06	0.06
Benzo[b]fluoranthene	< 0.06	0.06
Benzo[ghi]perylene	< 0.06	0.06
Benzo[k]fluoranthene	< 0.09	0.09
Chrysene	< 0.09	0.09
Dibenzo-[a,h]-anthracene	< 0.06	0.06
Fuoranthene	< 0.18	0.18
Fluorene	< 0.09	0.09
Indeno[1,2,3-cd]pyrene	< 0.09	0.09
Naphthalene	< 0.24	0.24
Phenanthrene	< 0.09	0.09
Pyrene	< 0.18	0.18
<b>Total PAH's</b>	<b>1.92</b>	

2006 (µg/L)	Value	MDL
Acenaphthene	< 0.44	0.44
Acenaphthylene	< 0.20	0.07
Anthracene	< 0.04	0.04
Benzo[a]anthracene	< 0.08	0.08
Benzo[a]pyrene	< 0.08	0.08
Benzo[b]fluoranthene	< 0.08	0.08
Benzo[ghi]perylene	< 0.08	0.08
Benzo[k]fluoranthene	< 0.12	0.12
Chrysene	< 0.12	0.12
Dibenzo-[a,h]-anthracene	< 0.06	0.06
Fuoranthene	< 0.24	0.06
Fluorene	< 0.12	0.12
Indeno[1,2,3-cd]pyrene	< 0.12	0.12
Naphthalene	< 0.32	0.32
Phenanthrene	< 0.12	0.12
Pyrene	< 0.24	0.24
<b>Total PAH's</b>	<b>2.46</b>	

DNQ = detected but not quantified

Appendix D-5 (cont.)

Oceanside Water Pollution Control Plant Polycyclic Aromatic Hydrocarbons (PAHs)  
on Final Effluent

2007 (µg/L)	Value	MDL
Acenaphthene	< 0.356	0.356
Acenaphthylene	< 0.162	0.162
Anthracene	< 0.033	0.033
Benzo[a]anthracene	0.002 DNQ	0.002
Benzo[a]pyrene	0.002 DNQ	0.002
Benzo[b]fluoranthene	< 0.074	0.01
Benzo[ghi]perylene	0.005 DNQ	0.004
Benzo[k]fluoranthene	0.003 DNQ	0.002
Chrysene	0.003 DNQ	0.002
Dibenzo-[a,h]-anthracene	0.002 DNQ	0.001
Fuoranthene	< 0.174	0.174
Fluorene	< 0.092	0.092
Indeno[1,2,3-cd]pyrene	< 0.086	0.086
Naphthalene	0.039 DNQ	0.101
Phenanthrene	< 0.102	0.055
Pyrene	0.015 DNQ	0.012
<b>Total PAH's</b>	<b>1.15</b>	

2008 (µg/L)	Value	MDL
Acenaphthene	< 0.132	0.0132
Acenaphthylene	< 0.176	0.176
Anthracene	0.002 DNQ	0.0015
Benzo[a]anthracene	< 0.008	0.008
Benzo[a]pyrene	< 0.008	0.008
Benzo[b]fluoranthene	< 0.008	0.008
Benzo[ghi]perylene	< 0.008	0.008
Benzo[k]fluoranthene	< 0.004	0.004
Chrysene	< 0.008	0.008
Dibenzo-[a,h]-anthracene	< 0.004	0.004
Fuoranthene	< 0.048	0.048
Fluorene	< 0.044	0.044
Indeno[1,2,3-cd]pyrene	< 0.032	0.032
Naphthalene	< 0.099	0.099
Phenanthrene	< 0.044	0.044
Pyrene	< 0.024	0.024
<b>Total PAH's</b>	<b>0.649</b>	

2009 (µg/L)	Value	MDL
Acenaphthene	< 0.0674	0.0674
Acenaphthylene	< 0.1924	0.01924
Anthracene	0.006	0.0009
Benzo[a]anthracene	0.0136 DNQ	0.0008
Benzo[a]pyrene	0.0151 DNQ	0.0034
Benzo[b]fluoranthene	0.0151 DNQ	0.0017
Benzo[ghi]perylene	0.006 DNQ	0.002
Benzo[k]fluoranthene	0.0151 DNQ	0.002
Chrysene	0.0072 DNQ	0.001
Dibenzo-[a,h]-anthracene	0.032 DNQ	0.001
Fuoranthene	0.0176 DNQ	0.0092
Fluorene	< 0.0134	0.0134
Indeno[1,2,3-cd]pyrene	0.0143 DNQ	0.0032
Naphthalene	0.0727 DNQ	0.0545
Phenanthrene	0.4711 DNQ	0.169
Pyrene	0.0398 DNQ	0.0182
<b>Total PAH's</b>	<b>0.9854</b>	

2010 (µg/L)	Value	MDL
Acenaphthene	< 0.018	0.018
Acenaphthylene	< 0.051	0.051
Anthracene	0.001 DNQ	0.001
Benzo[a]anthracene	0.003 DNQ	0.0005
Benzo[a]pyrene	< 0.002	0.002
Benzo[b]fluoranthene	0.004 DNQ	0.002
Benzo[ghi]perylene	< 0.002	0.002
Benzo[k]fluoranthene	< 0.007	0.007
Chrysene	0.002 DNQ	0.001
Dibenzo-[a,h]-anthracene	< 0.001	0.001
Fuoranthene	< 0.016	0.016
Fluorene	0.005 DNQ	0.003
Indeno[1,2,3-cd]pyrene	< 0.002	0.002
Naphthalene	0.061 DNQ	0.017
Phenanthrene	0.004 DNQ	0.002
Pyrene	0.006 DNQ	0.005
<b>Total PAH's</b>	<b>0.186</b>	

2011 (µg/L)	Value	MDL
Acenaphthene	< 0.018	0.018
Acenaphthylene	< 0.051	0.051
Anthracene	< 0.001	0.001
Benzo[a]anthracene	< 0.0005	0.0005
Benzo[a]pyrene	< 0.002	0.002
Benzo[b]fluoranthene	< 0.002	0.002
Benzo[ghi]perylene	< 0.002	0.002
Benzo[k]fluoranthene	< 0.007	0.007
Chrysene	< 0.001	0.001
Dibenzo-[a,h]-anthracene	< 0.001	0.001
Fuoranthene	< 0.016	0.016
Fluorene	< 0.003	0.003
Indeno[1,2,3-cd]pyrene	< 0.002	0.002
Naphthalene	0.027 DNQ	0.017
Phenanthrene	< 0.002	0.002
Pyrene	< 0.005	0.005
<b>Total PAH's</b>	<b>0.14</b>	

2012 (µg/L)	Value	MDL
Acenaphthene	< 0.61	0.61
Acenaphthylene	< 0.74	0.74
Anthracene	< 0.41	0.41
Benzo[a]anthracene	< 0.28	0.28
Benzo[a]pyrene	< 0.32	0.32
Benzo[b]fluoranthene	< 0.31	0.31
Benzo[ghi]perylene	< 0.62	0.62
Benzo[k]fluoranthene	< 0.31	0.31
Chrysene	< 0.31	0.31
Dibenzo-[a,h]-anthracene	< 0.73	0.73
Fuoranthene	< 0.54	0.54
Fluorene	< 0.71	0.71
Indeno[1,2,3-cd]pyrene	< 0.67	0.67
Naphthalene	< 0.48	0.48
Phenanthrene	< 0.48	0.48
Pyrene	< 0.4	0.4
<b>Total PAH's</b>	<b>7.92</b>	

DNQ = detected but not quantified



Appendix D-7  
*Analytical techniques and detection limits for  
 sediment metals analysis (mg/Kg, dry weight)  
 September 2012*

Element		Method	Method Detection Limit (MDL)
Aluminum	Al	ICP-AES	9.46
Arsenic	As	ICP-AES	0.10
Cadmium	Cd	ICP-AES	0.02
Chromium	Cr	ICP-AES	0.04
Copper	Cu	ICP-AES	0.06
Iron	Fe	ICP-AES	7.15
Lead	Pb	ICP-AES	0.01
Manganese	Mn	ICP-AES	0.02
Mercury	Hg	CVAAS	0.02
Nickel	Ni	ICP-AES	0.10
Selenium	Se	HGAAS	0.04
Silver	Ag	FAAS	0.01
Zinc	Zn	ICP-AES	0.32

ICP-AES = inductively coupled plasma atomic emission spectroscopy

CVAAS = cold vapor atomic absorption spectroscopy

GF-AAS = graphite furnace atomic absorption spectroscopy

FAAS = flame atomic absorption spectroscopy

HGAAS = hydride generation atomic absorption spectroscopy

Appendix D-8  
*Oceanside Water Pollution Control Plant*  
*Metals (mg/Kg) from final effluent*  
*2001 to 2012*

Metal (mg/Kg)	2001	2001 MDL	2002	2002 MDL	2003	2003 MDL
Silver, Ag	0.95	1.5	9.77	2.81	3.54 DNQ	1.21
Arsenic, As	5.91	5.7	10.82	6.91	12.28 DNQ	5.76
Cadmium, Cd <	0.33	0.4	1.52	0.67	0.58 DNQ	0.33
Chromium, Cr <	1.31	0.6	12.18	4.87	10.98 DNQ	3.05
Copper, Cu	32.24	14.7	219.61	4.16	17.00 DNQ	0.59
Mercury, Hg <	0.02	0.001	0.31	0.01	0.21	0.01
Nickel, Ni	6.57	2.1	35.28	6.97	18.65 DNQ	4.48
Lead, Pb	4.80	4.3	28.36	7.91	13.57 DNQ	8.90
Zinc, Zn	119.85	22.2	760.19	9.38	483.27	8.04

Metal (mg/Kg)	2004	2004 MDL	2005	2005 MDL	2006	2006 MDL
Silver, Ag	3.71 DNQ	1.41	1.75 DNQ	0.45	1.42 DNQ	0.34
Arsenic, As	17.55 DNQ	8.06	8.79 DNQ	3.71	8.45 DNQ	5.12
Cadmium, Cd	1.42 DNQ	0.67	1.37 DNQ	0.40	0.33 DNQ	0.12
Chromium, Cr	7.28 DNQ	1.83	8.20 DNQ	2.27	6.30 DNQ	1.90
Copper, Cu	222.34 DNQ	3.56	14.71 DNQ	3.52	182.87	41.93
Mercury, Hg	0.28	0.01	0.16	0.01	0.16	0.01
Nickel, Ni	26.16 DNQ	6.48	38.55 DNQ	3.79	33.04 DNQ	2.41
Lead, Pb	19.82 DNQ	6.92	14.02 DNQ	7.43	19.00 DNQ	5.95
Zinc, Zn	527.73	8.04	437.06	6.30	410.60	5.47

DNQ = Detected but not quantified

Appendix D-8 (cont.)  
*Oceanside Water Pollution Control Plant*  
*Metals (mg/Kg) from final effluent*  
*2001 to 2012 (cont.)*

Metal (mg/Kg)	2007	2007 MDL	2008	2008 MDL	2009	2009 MDL
Silver, Ag	8.2 DNQ	0.54	1.6 DNQ	0.29	2.3 DNQ	0.245
Arsenic, As	4.0 DNQ	2.56	16.6	16.64	22.4 DNQ	5.166
Cadmium, Cd	3.1 DNQ	0.35	2.2 DNQ	0.35	0.9 DNQ	0.038
Chromium, Cr	3.6 DNQ	0.71	3.0 DNQ	0.95	11.0 DNQ	1.593
Copper, Cu	249.6	45.15	241.5	41.93	174.4	17.213
Mercury, Hg	15.4	0.01	0.1	0.01	0.1 DNQ	0.001
Nickel, Ni	26.2 DNQ	2.41	28.0 DNQ	2.41	19.8 DNQ	2.560
Lead, Pb	22.5 DNQ	5.95	14.5 DNQ	6.70	7.4 DNQ	3.696
Zinc, Zn	510.8	5.89	522.9	5.47	298.8	3.995

Metal (mg/Kg)	2010	2010 MDL	2011	2011 MDL	2012	2012 MDL
Silver, Ag	1.30 DNQ	0.06	0.95 DNQ	0.04	1.64 DNQ	0.04
Arsenic, As	8.65 DNQ	1.33	10.56 DNQ	0.84	10.76 DNQ	0.84
Cadmium, Cd	1.13 DNQ	0.11	0.77 DNQ	0.07	0.62 DNQ	0.05
Chromium, Cr	7.63 DNQ	0.08	9.22	0.05	11.44	0.05
Copper, Cu	130.12	0.57	144.03	0.36	163.17	0.36
Mercury, Hg	0.11	0.00	0.10	0.00	0.17	0.00
Manganese, Mn					34.95	0.03
Nickel, Ni	26.26	0.38	30.91	0.24	32.20	0.24
Lead, Pb	7.65 DNQ	0.38	8.20 DNQ	0.24	8.53 DNQ	0.24
Selenium, Se	2.84 DNQ	0.51	2.37 DNQ	0.36	2.32 DNQ	0.36
Zinc, Zn	283.88	3.80	313.24	2.40	364.54	2.40

DNQ = Detected but not quantified

Appendix D-9

*Multivariate Analysis of Variance (MANOVA) and Analysis of Variance (ANOVA) of Trace Metals  
1997 to 2012*

**Multivariate Analysis of Variance (MANOVA)**

Source Term	DF	Sum of Squares	Mean Square	F-Ratio	Prob Level	Power (Alpha = 0.05)
A: area	1	1.380235E+07	1.380235E+07	0.43	0.513001	0.100166
S	524	1.687755E+10	3.220906E+07			
Total (adjusted)	525	1.689135E+10				
Total	526					

**Analysis of Variance (ANOVA)**

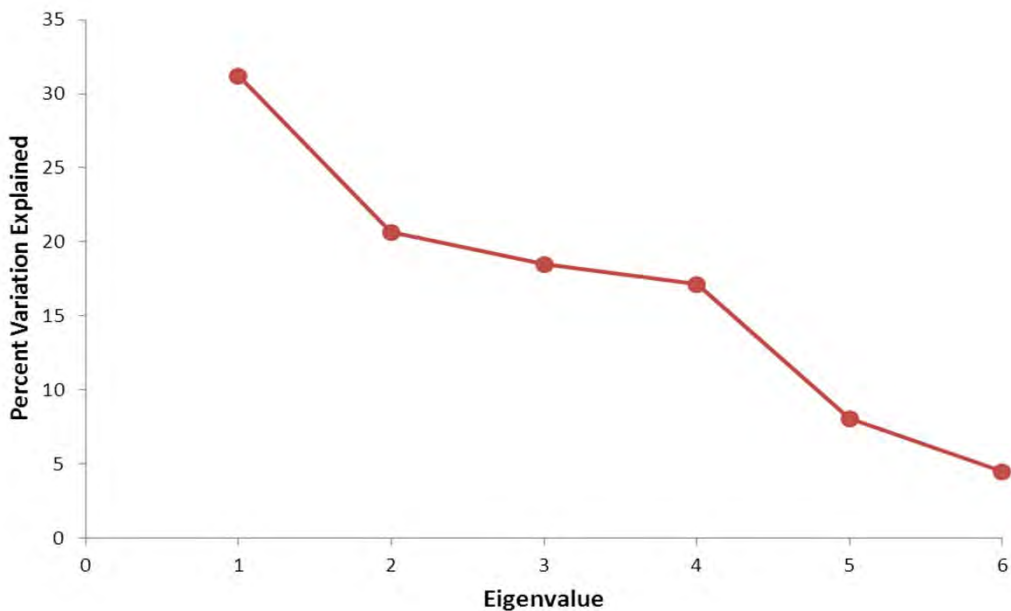
Source Term	DF	Sum of Squares	Mean Square	F-Ratio	Prob Level	Power (Alpha = 0.05)
A: area	1	3940585	3940585	2.49	0.114913	0.350781
B: Analyte	15	1.604365E+10	1.069576E+09	676.97	0*	1.000000
AB	15	3.965750E+07	2643833	1.67	0.052647	0.913093
S	495	7.804987E+08	1579957			
Total Adjusted	525	1.689135E+10				
Total	526					

\*Term significant at alpha = 0.05



Appendix D-10  
*Skree plot of PCA factors, Significant Eigenvalue, and  
 Factor loading after Varimax rotation  
 1997 to 2012*

Skree Plot of PCA Factors

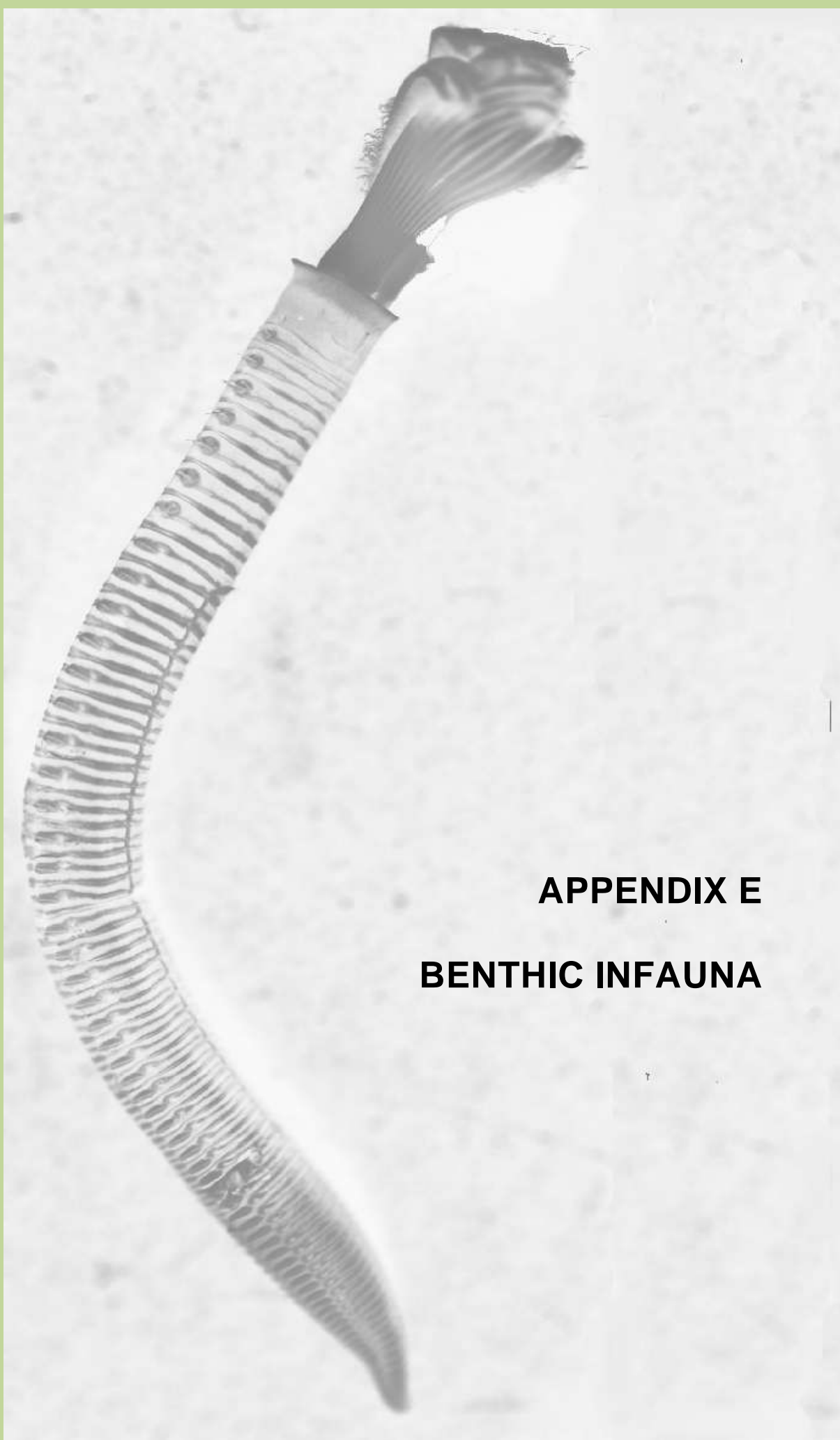


Significant Eigenvalue

Factor	Eigenvalue	Percent of variation explained
1	1.8729	31.21
2	1.2384	20.64
3	1.1084	18.47
4	1.0273	17.12
Total		87.44

Factor Loading after varimax rotation

Variables	Factor1	Factor2	Factor3	Factor4
sMETq	-0.00761	<b>0.968698</b>	-0.02564	0.069492
sORG	-0.195714	0.072695	0.0721	<b>0.975131</b>
SLT-CLY	<b>-0.837399</b>	-0.251329	-0.19431	0.123944
TKN	-0.114009	-0.038322	<b>0.93724</b>	0.058219
TOC	<b>-0.74876</b>	0.294356	0.367601	0.18755
TVS	<b>-0.748103</b>	0.115937	0.425714	0.132874



**APPENDIX E**  
**BENTHIC INFAUNA**

APPENDIX E  
Benthic Infauna data  
1997-2012

<u>Appendix</u>		<u>Page</u>
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E-2	Taxa and abundances of benthic infauna collected by station in 2012	E-10
E-3	Community measures for each station 1997 - 2012	E-53
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Appendix E-1

Traditional classification of benthic infauna collected from 1997 through 2012

CNIDARIA		NASSARIIDAE	
HYDROZOA		<i>Caesia fossatus</i>	
<i>Euphysa</i> spp.		<i>Caesia rhinetes</i>	
ANTHOZOA	<i>Anthozoa</i>	OLIVELLIDAE	
PENNATULACEA		<i>Callianax pycna</i>	
<i>Stylatula</i> spp.		BORSONIIDAE	
<i>Edwardsia juliae</i>		<i>Ophiodermella</i> spp.	
<i>Halcampa decententaculata</i>		MANGELIIDAE	
PLATYHELMINTHES		<i>Kurtziella plumbea</i>	
<i>Turbellaria</i>		<i>Kurtzina beta</i>	
NEMERTEA		TURRIDAE	
<i>Carinoma mutabilis</i>		ACTEONIDAE	
<i>Carinoma</i> sp.		<i>Rictaxis punctocaelatus</i>	
<i>Tubulanidae</i> sp. B		APLUSTRIDAE	
<i>Tubulanus cingulatus</i>		<i>Parvaplustrum</i> sp. A	
<i>Tubulanus nothus</i>		PYRAMIDELLIDAE	
<i>Tubulanus pellucidus</i>		<i>Brachystomia angularis</i>	
<i>Tubulanus</i> spp.		<i>Cyclostremella californica</i>	
Lineidae		<i>Odostomia churchi</i>	
<i>Cerebratulus californiensis</i>		<i>Odostomia</i> spp.	
<i>Cerebratulus</i> spp.		<i>Turbonilla</i> spp.	
<i>Micrura</i> spp. (?)		HETEROBRANCHIA	
<i>Paranemertes californica</i>		<i>Opisthobranchia</i>	
<i>Paranemertes californica</i>		<i>Cephalaspidea</i>	
<i>Monostylifera</i> sp. B		DIAPHANIDAE	
MOLLUSCA		<i>Diaphana californica</i>	
APLACOPHORA		HAMINOEIDAE	
CHAETODERMATIDA		<i>Haminoea virescens</i>	
<i>Falcidens longus</i>		<i>Haminoea</i> sp.	
GASTROPODA		PHILINIDAE	
NATICIDAE		<i>Philine auriformis</i>	
<i>Glossaluax reclusiana</i>		<i>Philine</i> spp.	
<i>Polinices draconis</i>		AGLAJIDAE	
<i>Polinices lewisii</i>		<i>Aglaja ocelligera</i>	
<i>Polinices</i> spp.		<i>Melanochlamys diomedea</i>	
RISSOIDAE		CYLICHNIDAE	
<i>Alvania compacta</i>		<i>Acteocina</i> spp.	
<i>Alvania rosana</i>		<i>Cylichna attonsa</i>	
BARLEEIIDAE		<i>Cylichna</i> spp.	
<i>Barleeia haliotiphila</i>		GASTROPTERIDAE	
EPITONIIDAE		<i>Gastropteron pacificum</i>	
<i>Epitonium</i> spp.		RETUSIDAE	
EULIMIDAE		<i>Volvulella cylindrica</i>	
<i>Balcis</i> spp.		<i>Volvulella panamica</i>	
COLUMBELLIDAE		<i>Volvulella</i> spp.	
<i>Astyris gausapata</i>		NUDIBRANCHIA	
<i>Mitrella</i> spp.		<i>Armina californica</i>	
<i>Mitrella</i> spp.		<i>Dendronotus</i> spp.	
		<i>Dirona</i> spp.	
		<i>Eubranchus misakiensis</i>	

Appendix E-1 (cont.)

Traditional classification of benthic infauna collected from 1997 through 2012

BIVALVIA

NUCULANIDAE

*Saccella penderi*  
*Saccella spp.*  
*Saccella taphria*

YOLDIIDAE

*Yoldia cooperii*  
*Yoldia seminuda*  
*Yoldia spp.*

MYTILIDAE

*Modiolus capax*  
*Modiolus rectus*  
*Modiolus rectus*  
*Modiolus spp.*  
*Mytilus spp.*

PECTINIDAE

*Leptopecten latiauratus*

LUCINIDAE

THYASIRIDAE

*Axinopsida serricata*

LASAEIDAE

*Kellia sp. SF1*  
*Kellia spp.*  
*Kurtiella coani*  
*Kurtiella grippi*  
*Kurtiella pedroana*  
*Kurtiella sp. SF1*  
*Kurtiella tumida*  
*Kurtiella spp.*

NEOLEPTONIDAE

*Neolepton salmoneum*

CARDIIDAE

*Clinocardium nuttallii*  
*Trachycardium quadragenarium*

TELLINIDAE

*Macoma acolasta*  
*Macoma nasuta*  
*Macoma yoldiformis*  
*Macoma spp.*  
*Tellina bodegensis*  
*Tellina modesta*  
*Tellina nuculoides*  
*Tellina spp.*

SOLENIDAE

*Solen rostiformis*  
*Solen sicarius*

PHARIDAE

*Siliqua lucida*  
*Siliqua sp. SF1*  
*Siliqua spp.*

HIATELLIDAE

*Hiatella arctica*  
*Saxicavella nybakkeni*  
*Saxicavella pacifica*

VENERIDAE

*Leukoma staminea*  
*Compsomyax subdiaphana*  
*Nutricola confusa*  
*Nutricola tantilla*  
*Nutricola spp.*  
*Venerupis philippinarum*

PETRICOLIDAE

*Cooperella subdiaphana*

MACTRIDAE

*Mactromeris catilliformis*  
*Simomactra sp.*  
*Tresus spp.*

MYIDAE

*Cryptomya californica*

PANDORIDAE

*Heteroclidus punctatus*  
*Pandora bilirata*

LYONSIIDAE

*Lyonsia californica*

THRACIIDAE

*Asthenothaerus spp.*  
*Thracioidea sp. SF1*

SCAPHOPODA

GADILIDAE

*Gadila aberrans*

SIPUNCULA

*Golfingia sp. A*  
*Siphonosoma ingens*  
*Sipunculus nudus*  
*Themiste spp.*

ECHIURA

*Listriolobus pelodes*

ANNELIDA

POLYCHAETA

CAPITELLIDAE

*Capitella capitata complex*  
*Heteromastus filobranthus*  
*Heteromastus filiformis Cmplx*  
*Heteromastus spp.*  
*Mediomastus acutus*  
*Mediomastus spp.*  
*Notomastus lineatus*  
*Notomastus hemipodus*  
*Notomastus spp.*

Appendix E-1 (cont.)

Traditional classification of benthic infauna collected from 1997 through 2012

COSSURIDAE

*Cossura candida*  
*Cossura spp.*

MALDANIDAE

*Axiothella rubrocincta*  
*Euclymeninae*  
*Euclymeninae sp. SF1*  
*Petaloclymene pacifica*

OPHELIIDAE

*Armandia brevis*  
*Ophelia assimilis*  
*Travisia gigas*

ORBINIIDAE

*Leitoscoloplos pugettensis*  
*Naineris uncinata*  
*Phylo felix*  
*Scoloplos armiger*  
*Scoloplos sp. SF1*  
*Scoloplos spp.*

PARAONIDAE

*Aricidea (Acmira) catherinae*  
*Aricidea (Acmira) horikoshii*  
*Aricidea (Acmira) lopezi*  
*Aricidea (Acmira) spp.*  
*Aricidea (Aedicira) pacifica*  
*Aricidea (Aedicira) sp. A*  
*Aricidea (Aedicira) spp.*  
*Aricidea (Aricidea) wassi*  
*Aricidea (Aricidea) sp. SF1*  
*Aricidea (Aricidea) sp. SF2*  
*Aricidea (Aricidea) sp. SF3*  
*Aricidea spp.*  
*Paraonella platybranchiata*

SCALIBREGMATIDAE

*Scalibregma californicum*

APHRODITIDAE

*Aphrodita refulgida*  
*Aphrodita spp.*

POLYNOIDAE

*Lepidasthenia berkeleyae*  
*Lepidasthenia longicirrata*  
*Lepidasthenia spp.*  
*Halosydna brevisetosa*  
*Halosydna spp.*  
*Arcteobia cf. anticostiensis*  
*Harmothoe imbricata complex*  
*Harmothoe spp.*  
*Hesperonoe laevis*  
*Malmgreniella liei*  
*Malmgreniella spp.*  
*Tenonia priops*

PHOLOIDAE

*Pholoe glabra*  
*Pholoe spp.*  
*Pholoides asperus*

SIGALIONIDAE

*Sigalion spinosus*  
*Sthenelais berkeleyi*  
*Sthenelais tertiaglabra*  
*Sthenelais verruculosa*  
*Sthenelais sp.*  
*Sthenolepis fimbriarum*

PISIONIDAE

*Pisione remota*

CHRYSOPETALIDAE

*Paleanotus bellis*

HESIONIDAE

*Heteropodarke heteromorpha*  
*Microphthalmus spp. complex*  
*Podarke spp.*  
*Podarkeopsis glabrus*

NEREIDIDAE

*Nereis neoneanthes*  
*Nereis spp.*  
*Platynereis bicanaliculata*

PILARGIDAE

*Ancistrostylis groenlandica*  
*Ancistrostylis spp.*  
*Parandalia fauveli*  
*Pilargis berkeleyae*  
*Sigambra sp. SF2*  
*Sigambra spp.*

SYLLIDAE

*Autolytinae*  
*Eusyllis transecta*  
*Exogone lourei*  
*Sphaerosyllis californiensis*  
*Streptosyllis sp. SF1*  
*Syllis (Ehlersia) hyperionii*  
*Syllis spp.*  
*Typosyllis farallonensis*  
*Typosyllis nipponiica*  
*Typosyllis spp.*

GLYCERIDAE

*Glycera americana*  
*Glycera capitata*  
*Glycera macrobranchia*  
*Glycera robusta*  
*Glycera tenuis*  
*Glycera spp.*  
*Hemipodia simplex*

Appendix E-1 (cont.)

Traditional classification of benthic infauna collected from 1997 through 2012

GONIADIDAE

*Glycinde picta*  
*Glycinde sp. SF1*  
*Glycinde spp.*  
*Goniada maculata*  
*Goniada spp.*

NEPHTYIDAE

*Micronephtys cornuta*  
*Nephtys caeca*  
*Nephtys caecoides*  
*Nephtys californiensis*  
*Nephtys ferruginea*  
*Nephtys sp. SF1*  
*Nephtys spp.*

PHYLLODOCIDAE

*Eteone ?californica*  
*Eteone fauchaldi*  
*Eteone longa?*  
*Eteone (Mysta) sp. SF1*  
*Eteone pacifica*  
*Eteone sp. SF3*  
*Eteone sp. SF4*  
*Eteone spp.*  
*Eumida longicornuta*  
*Hesionura coineaui difficilis*  
*Paranaitis sp. SF1*  
*Phyllodoce cuspidata*  
*Phyllodoce groenlandica*  
*Phyllodoce hartmanae*  
*Phyllodoce longipes*  
*Phyllodoce multipapillata*  
*Phyllodoce spp.*  
*Phyllodoce williamsi*  
*Phyllodoce williamsi*

SPHAERODORIDAE

AMPHINOMIDAE

*Pareurythoe californica*

DORVILLEIDAE

*Dorvillea rudolphi*  
*Ophryotrocha sp. SF1*  
*Protodorvillea gracilis*  
*Protodorvillea spp.*

LUMBRINERIDAE

*Lumbrinerides platypygos*  
*Lumbrineris japonica*  
*Lumbrineris californiensis*  
*Lumbrineris limicola*  
*Ninno spp.*  
*Scoletoma luti*  
*Scoletoma spp.*

OENONIDAE

*Arabella iricolor*  
*Drilonereis falcata*  
*Drilonereis spp.*

ONUPHIDAE

*Diopatra ornata*  
*Onuphis sp. A*  
*Onuphis spp.*

OWENIIDAE

*Galathowenia oculata*  
*Owenia collaris*

SABELLARIIDAE

*Neosabellaria cementarium*

SABELLIDAE

*Chone mollis*  
*Chone spp.*  
*Euchone hancocki*  
*Euchone spp.*  
*Paradialychone ecaudata*  
*Paradialychone eiffelturris*  
*Potamethus sp. A*

CIRRATULIDAE

*Aphelochaeta cf. elongata*  
*Aphelochaeta monilaris*  
*Aphelochaeta sp. SF3*  
*Aphelochaeta petersenae*  
*Aphelochaeta spp.*  
*Chaetozone bansei*  
*Chaetozone columbiana*  
*Chaetozone spp.*  
*Tharyx spp.*

FLABELLIGERIDAE

*Pherusa neopapillata*

AMPHARETIDAE

*Ampharete acutifrons*  
*Ampharete finmarchica*  
*Ampharete labrops*  
*Ampharete spp.*  
*Amphicteis scaphobranchiata*  
*Amphicteis spp.*  
*Melinna oculata*  
*Schistocomus sp. A*

PECTINARIIDAE

*Pectinaria californiensis*

TEREBELLIDAE

*Amaeana occidentalis*  
*Amaeana spp.*  
*Polycirrus sp. I*  
*Polycirrus sp. SF1*  
*Polycirrus spp.*

Appendix E-1 (cont.)

Traditional classification of benthic infauna collected from 1997 through 2012

<i>Eupolymnia heterobranchia</i>	<i>Scolelepis occidentalis</i>
<i>Eupolymnia spp.</i>	<i>Scolelepis sp. SF1</i>
<i>Lanassa venusta</i>	<i>Scolelepis sp. SF2</i>
<i>Nicolea sp. SF1</i>	<i>Scolelepis sp. SF3</i>
<i>Pista elongata</i>	<i>Scolelepis spp.</i>
<i>Pista estevanica</i>	<i>Spio butleri</i>
<i>Pista sp. SF1</i>	<i>Spiophanes duplex</i>
<i>Pista wui</i>	<i>Spiophanes norrisi</i>
<i>Pista spp.</i>	<i>Spiophanes spp.</i>
<i>Streblosoma sp. SF1</i>	<i>Streblospio benedicti</i>
<i>Streblosoma spp.</i>	
CHAETOPTERIDAE	ARCHIANNELIDA
<i>Mesochaetopterus taylori</i>	POLYGORDIIDAE
<i>Mesochaetopterus sp. SF1</i>	PROTODRILIDAE
<i>Mesochaetopterus spp.</i>	SACCOCIRRIDAE
<i>Phyllochaetopterus cf. claparedi</i>	HIRUDINEA
<i>Spiochaetopterus costarum</i>	OLIGOCHAETA
MAGELONIDAE	ENCHYTRAEIDAE
<i>Magelona berkeleyi</i>	TUBIFICIDAE
<i>Magelona californica</i>	ARTHROPODA
<i>Magelona hartmanae</i>	PYCNOGONIDA
<i>Magelona pitelkai</i>	<i>Achelia alaskensis</i>
<i>Magelona sacculata</i>	<i>Pycnogonum stearnsi</i>
<i>Magelona spp.</i>	CRUSTACEA
POECILOCHAETIDAE	CIRRIPEDIA
<i>Poecilochaetus johnsoni</i>	<i>Balanus spp.</i>
TROCHOCHAETIDAE	COPEPODA
<i>Trochochaeta franciscanum</i>	HARPACTICOIDA
SPIONIDAE	OSTRACODA
<i>Apoprionospio pygmaea</i>	<i>Leuroleberis sharpei</i>
<i>Boccardia pugettensis</i>	<i>Xenoleberis californica</i>
<i>Boccardia spp.</i>	<i>Ostracoda sp. SF2</i>
<i>Carazziella sp. A</i>	<i>Euphilomedes carcharodonta</i>
<i>Dipolydora brachycephala</i>	<i>Euphilomedes spp.</i>
<i>Dipolydora commensalis</i>	<i>Eusarsiella zostericola</i>
<i>Dipolydora socialis</i>	<i>Podocopida</i>
<i>Dipolydora sp. SF1</i>	MYSIDACEA
<i>Dipolydora magna</i>	<i>Acanthomysis spp.</i>
<i>Dipolydora spp.</i>	<i>Archaeomysis grebnitzkii</i>
<i>Dispio uncinata</i>	<i>Holmesimysis costata</i>
<i>Laonice cirrata</i>	<i>Holmesimysis macropsis</i>
<i>Paraprionospio alata</i>	<i>Holmesimysis sp. A</i>
<i>Polydora cornuta</i>	<i>Holmesimysis spp.</i>
<i>Polydora narica</i>	<i>Neomysis kadiakensis</i>
<i>Polydora spp.</i>	<i>Neomysis spp.</i>
<i>Prionospio lighti</i>	AMPHIPODA
<i>Prionospio steenstrupi</i>	AORIDAE
<i>Prionospio spp.</i>	<i>Aoroides inermis</i>
<i>Scolelepis squamata</i>	<i>Aoroides spinosus</i>
<i>Spiophanes berkeleyorum</i>	<i>Aoroides spp.</i>
	<i>Grandidierella japonica</i>



Appendix E-1 (cont.)

Traditional classification of benthic infauna collected from 1997 through 2012

COROPHOIDEA	<i>Pacifoculodes barnardi</i>
<i>Monocorophium acherusicum</i>	<i>Americhelidium</i> spp.
<i>Monocorophium insidiosum</i>	<i>Westwoodilla tone</i>
<i>Monocorophium</i> sp.	SYNOPIIDAE
<i>Sinocorophium heteroceratum</i>	<i>Tiron biocellata</i>
<i>Cheirimeideia zotea</i>	ARGISSIDAE
<i>Cheirimeideia</i> spp.	<i>Argissa hamatipes</i>
<i>Protomeideia penates</i>	LILJEBORGIIDAE
<i>Protomeideia</i> spp.	<i>Listriella diffusa</i>
CAPRELLIDAE	<i>Listriella goleta</i>
<i>Caprella californica</i>	<i>Listriella</i> spp.
<i>Caprella equilibra</i>	HAUSTORIIDAE
<i>Caprella mendax</i>	<i>Eohaustorius</i> spp.
<i>Caprella natalensis</i>	PHOXOCEPHALIDAE
<i>Caprella</i> spp.	<i>Foxiphalus obtusidens</i>
<i>Metacaprella anomala</i>	<i>Grandifoxus grandis</i>
<i>Tritella pilimana</i>	<i>Grandifoxus</i> cf. <i>grandis</i>
ISAEIDAE	<i>Rhepoxynius abronius</i>
<i>Cheirophotis</i> spp.	<i>Rhepoxynius fatigans</i>
DULICHIIDAE	<i>Rhepoxynius lucubrans</i>
<i>Dyopedos arcticus</i>	<i>Rhepoxynius menziesi</i>
PODOCERIDAE	<i>Rhepoxynius tridentatus</i>
<i>Podocerus spongicolus</i>	<i>Rhepoxynius variatus</i>
ISCHYROCERIDAE	<i>Rhepoxynius vigitegus</i>
<i>Erichthonius brasiliensis</i>	<i>Rhepoxynius</i> spp.
<i>Erichthonius</i> spp.	<i>Mandibulophoxus gilesi</i>
<i>Ischyrocerus anguipes</i>	<i>Eobrolgus spinosus</i>
<i>Ischyrocerus pelagops</i>	<i>Metaphoxus frequens</i>
<i>Ischyrocerus</i> sp. SF2	DEXAMINIDAE
<i>Jassa marmorata</i>	<i>Atylus tridens</i>
<i>Jassa</i> spp.	AMPELISCIDAE
<i>Microjassa barnardi</i>	<i>Ampelisca abdita</i>
PHOTIDAE	<i>Ampelisca agassizi</i>
<i>Gammaropsis</i> sp.	<i>Ampelisca careyi</i>
<i>Photis bifurcata</i>	<i>Ampelisca cristata</i>
<i>Photis brevipes</i>	<i>Ampelisca milleri</i>
<i>Photis californica</i>	<i>Ampelisca</i> spp.
<i>Photis conchicola</i>	MELITIDAE
<i>Photis macinerneyi</i>	<i>Desdimelita desdichada</i>
<i>Photis parvidons</i>	<i>Melita dentata</i>
<i>Photis</i> spp.	MELPHIDIPPIDAE
PLEUSTIDAE	<i>Melphisana</i> sp. SF1
<i>Gnathopleustes pugettensis</i>	MELITIDAE
STENOTHOIDAE	<i>Megamoera subtener</i>
<i>Stenothoides bicoma</i>	MEGALUROPIDAE
<i>Stenothoides burbanki</i>	<i>Gibberosus myersi</i>
<i>Stenothoe</i> spp.	LYSIANASSIDAE
OEDICEROTIDAE	<i>Lysianassidae</i> sp. SF1
<i>Americhelidium shoemakeri</i>	<i>Wecomedon</i> spp.
<i>Americhelidium</i> sp. SD1	

Appendix E-1 (cont.)

Traditional classification of benthic infauna collected from 1997 through 2012

PACHYNIDAE	
<i>Pachynus barnardi</i>	
URISTIDAE	
<i>Anonyx adoxus</i>	
ISOPODA	
ANTHURIDAE	
<i>Haliophasma geminatum</i>	
PARANTHURIDAE	
<i>Paranthura elegans</i>	
IDOTEIDAE	
<i>Edotia sublittoralis</i>	
<i>Edotia spp.</i>	
<i>Synidotea consolidata</i>	
<i>Synidotea laticauda</i>	
<i>Synidotea spp.</i>	
PARAMUNNIDAE	
<i>Munnogonium tillerae</i>	
<i>Munnogonium spp.</i>	
<i>Pleurogonium sp. SF1</i>	
SPHAEROMATIDAE	
<i>Gnorimosphaeroma oregonensis</i>	
<i>Tecticeps convexus</i>	
ANCINIDAE	
<i>Bathycopea daltonae</i>	
JANIRIDAE	
<i>Ianiropsis derjugini</i>	
MUNNIDAE	
<i>Munna spp.</i>	
TANAIDACEA	
PARATANAIDAE	
<i>Leptocheilia dubia</i>	
CUMACEA	
LEUCONIDAE	
<i>Eudorella pacifica</i>	
<i>Leucon spp.</i>	
NANNASTACIDAE	
<i>Campylaspis spp.</i>	
LAMPROPIDAE	
<i>Hemilamprops californicus</i>	
<i>Lamprops carinata</i>	
<i>Lamprops tomalesi</i>	
<i>Lamprops triserratus</i>	
<i>Lamprops spp.</i>	
<i>Mesolamprops dillonensis</i>	
DIASTYLIDAE	
<i>Anchicolurus occidentalis</i>	
<i>Diastylis santamariensis</i>	
<i>Diastylis spp.</i>	
<i>Diastylopsis dawsoni</i>	
<i>Diastylopsis tenuis</i>	
<i>Diastylopsis spp.</i>	
	DECAPODA
	CARIDEA
	HIPPOLYTIDAE
	<i>Heptacarpus stimpsoni</i>
	<i>Heptacarpus spp.</i>
	<i>Heptacarpus spp.</i>
	CRANGONIDAE
	<i>Crangon franciscorum</i>
	<i>Crangon nigricauda</i>
	<i>Crangon nigromaculata</i>
	<i>Crangon sp. SF1</i>
	<i>Crangon spp.</i>
	LISSOCRANGON
	<i>Lissocrangon stylirostris</i>
	ANOMURA
	CALLIANASSIDAE
	<i>Neotrypaea spp.</i>
	<i>Anomura</i>
	DIOGENIDAE
	<i>Isocheles pilosus</i>
	PAGURIDAE
	<i>Pagurus spp.</i>
	PORCELLANIDAE
	BLEPHARIPODIDAE
	<i>Blepharipoda occidentalis</i>
	BRACHYURA
	INACHOIDIDAE
	PARTHENOPIDAE
	<i>Heterocrypta occidentalis</i>
	CANCRIDAE
	<i>Cancer productus</i>
	<i>Metacarcinus gracilis</i>
	<i>Metacarcinus magister</i>
	<i>Romaleon antennarium</i>
	PINNOTHERIDAE
	<i>Fabia subquadrata</i>
	<i>Opisthopus transversus</i>
	<i>Scleroplax granulata</i>
	<i>Pinnixa franciscana</i>
	<i>Pinnixa spp.</i>
	NEMATODA
	ECHINODERMATA
	ASTEROIDEA
	<i>Pisaster brevispinus</i>
	OPHIUROIDEA
	<i>Amphiodia digitata</i>
	<i>Amphiodia spp.</i>
	ECHINOIDEA
	<i>Dendraster excentricus</i>

Appendix E-1 (cont.)

*Traditional classification of benthic infauna collected from 1997 through 2012*

HOLOTHUROIDEA	<i>Dendrochirotida</i>
	<i>Pentamera rigida</i>
	<i>Leptosynapta spp.</i>
	<i>Holothuroidea sp. SF1</i>
	<i>Paracaudina chilensis</i>
PHORONIDA	<i>Phoronis spp.</i>
BRACHIOPODA	<i>Inarticulata</i>
ECTOPROCTA	<i>Filicrisia franciscana</i>
	<i>Tricellaria ternata</i>
HEMICHORDATA	<i>Enteropneusta</i>
UROCHORDATA	
ASCIDIACEA	<i>Molgula manhattensis</i>

Appendix E-2  
Benthic infauna collected in 2012

<b>STATION 01</b>		<i>Amphiodia</i> spp.	5
<i>Spiophanes norrisi</i>	2230	<i>Cylichna attonsa</i>	5
<i>Photis</i> spp.	291	<i>Synidotea consolidata</i>	5
<i>Scoletoma luti</i>	175	<i>Ampelisca careyi</i>	4
<i>Protomedeia penates</i>	148	<i>Amphiodia digitata</i>	4
<i>Owenia collaris</i>	113	<i>Paranemertes californica</i>	4
<i>Photis macinerneyi</i>	80	<i>Scoloplos</i> sp. SF1	4
<i>Callianax pycna</i>	66	<i>Caesia rhinetes</i>	3
<i>Onuphis</i> spp.	40	<i>Modiolus capax</i>	3
<i>Mediomastus</i> spp.	33	<i>Nephtys caecoides</i>	3
<i>Tellina modesta</i>	31	<i>Spiophanes berkeleyorum</i>	3
<i>Pectinaria californiensis</i>	30	<i>Sthenelais verruculosa</i>	3
<i>Onuphis</i> sp. A	22	<i>Edotia sublittoralis</i>	2
<i>Photis parvidons</i>	21	Enteropneusta	2
<i>Glycinde picta</i>	19	<i>Glossaluax reclusiana</i>	2
<i>Bathycopea daltonae</i>	17	<i>Goniada maculata</i>	2
Lineidae	17	<i>Halcampa decententaculata</i>	2
<i>Euphilomedes carcharodonta</i>	15	<i>Hemilamprops californicus</i>	2
<i>Glycera macrobranchia</i>	15	<i>Kurtziella plumbea</i>	2
<i>Leukoma staminea</i>	15	Nassariidae	2
<i>Pacifoculodes barnardi</i>	15	Nemertea	2
<i>Pleurogonium</i> sp. SF1	15	<i>Podarkeopsis glabrus</i>	2
<i>Astyris gausapata</i>	14	<i>Siliqua lucida</i>	2
<i>Ischyrocerus pelagops</i>	14	Tubulanidae sp. B	2
<i>Glycinde</i> spp.	14	<i>Americhelidium shoemakeri</i>	1
<i>Diastylis santamariensis</i>	13	<i>Ampelisca cristata</i>	1
<i>Diastylopsis dawsoni</i>	13	<i>Ampharete acutifrons</i>	1
<i>Carinoma mutabilis</i>	12	<i>Amphicteis scaphobranchiata</i>	1
<i>Macoma</i> spp.	10	Anthozoa	1
<i>Pandora bilirata</i>	10	<i>Apoprionospio pygmaea</i>	1
<i>Tritella pilimana</i>	10	<i>Aricidea (Aedicira)</i> sp. A	1
<i>Clinocardium nuttallii</i>	9	<i>Aricidea (Aricidea)</i> sp. SF3	1
<i>Kurtiella tumida</i>	9	Cardiidae	1
<i>Crangon nigromaculata</i>	8	<i>Cheirimedeia zotea</i>	1
<i>Magelona hartmanae</i>	8	<i>Cooperella subdiaphana</i>	1
<i>Tenonia priops</i>	8	<i>Dendraster excentricus</i>	1
<i>Argissa hamatipes</i>	7	Dendrochirotida	1
Terebellidae	7	<i>Eteone (Mysta)</i> sp. SF1	1
<i>Macoma nasuta</i>	6	<i>Eumida longicornuta</i>	1
<i>Magelona sacculata</i>	6		

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Gastropteron pacificum</i>	1	<i>Amphiodia</i> spp.	14
<i>Haliophasma geminatum</i>	1	<i>Owenia collaris</i>	13
<i>Leitoscoloplos pugettensis</i>	1	<i>Podarkeopsis glabrus</i>	12
<i>Leptopecten latiauratus</i>	1	<i>Tritella pilimana</i>	11
<i>Mesochaetopterus</i> sp. SF1	1	<i>Onuphis</i> sp. A	10
<i>Mesolamprops dillonensis</i>	1	<i>Pacifoculodes barnardi</i>	9
<i>Micronephtys cornuta</i>	1	<i>Kurtiella tumida</i>	8
Nematoda	1	<i>Onuphis</i> spp.	7
<i>Odostomia</i> spp.	1	<i>Magelona hartmanae</i>	6
Ostracoda sp. SF2	1	<i>Nephtys caecoides</i>	6
<i>Paradialychone eiffelturris</i>	1	<i>Foxiphalus obtusidens</i>	5
<i>Pentamera rigida</i>	1	<i>Odostomia</i> spp.	5
<i>Photis brevipes</i>	1	<i>Tenonia priops</i>	5
<i>Phyllodoce williamsi</i>	1	<i>Cylichna</i> spp.	4
<i>Pinnixa franciscana</i>	1	<i>Eumida longicornuta</i>	4
Polynoidae	1	<i>Glycera macrobranchia</i>	4
<i>Sigambra</i> sp. SF2	1	<i>Leitoscoloplos pugettensis</i>	4
<i>Tecticeps convexus</i>	1	<i>Sthenelais verruculosa</i>	4
<i>Tresus</i> spp.	1	<i>Euphilomedes carcharodonta</i>	3
<i>Typosyllis farallonensis</i>	1	<i>Gastropteron pacificum</i>	3
<b>STATION 02</b>		<i>Modiolus capax</i>	3
<i>Spiophanes norrisi</i>	4237	<i>Bathycopea daltonae</i>	2
<i>Photis</i> spp.	1099	<i>Crangon nigromaculata</i>	2
<i>Photis macinerneyi</i>	412	<i>Diastylopsis dawsoni</i>	2
<i>Scoletoma luti</i>	170	<i>Edotia sublittoralis</i>	2
<i>Protomedeia penates</i>	99	<i>Kurtziella plumbea</i>	2
<i>Glycinde picta</i>	90	<i>Mactromeris catilliformis</i>	2
<i>Ischyrocerus pelagops</i>	69	<i>Micronephtys cornuta</i>	2
<i>Carinoma mutabilis</i>	59	<i>Pectinaria californiensis</i>	2
<i>Tellina modesta</i>	57	<i>Pherusa neopapillata</i>	2
<i>Pleurogonium</i> sp. SF1	48	<i>Phyllodoce williamsi</i>	2
<i>Callianax pycna</i>	34	<i>Phylo felix</i>	2
<i>Tresus</i> spp.	33	<i>Rhepoxynius fatigans</i>	2
<i>Leukoma staminea</i>	29	<i>Sipuncula</i>	2
<i>Astyris gausapata</i>	25	<i>Stylatula</i> spp.	2
<i>Mediomastus</i> spp.	24	<i>Typosyllis farallonensis</i>	2
<i>Clinocardium nuttallii</i>	22	<i>Amaeana occidentalis</i>	1
<i>Diastylis santamariensis</i>	17	<i>Ampharete</i> spp.	1
<i>Glycinde</i> spp.	17	Anthozoa	1
<i>Tiron biocellata</i>	17	<i>Armandia brevis</i>	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Compsomyax subdiaphana</i>	1	<i>Macoma nasuta</i>	19
<i>Dendraster excentricus</i>	1	<i>Pleurogonium</i> sp. SF1	18
<i>Eteone ?californica</i>	1	<i>Onuphis</i> sp. A	17
<i>Eudorella pacifica</i>	1	<i>Pandora bilirata</i>	17
<i>Halcampa decemtentaculata</i>	1	<i>Diastylis santamariensis</i>	16
<i>Kurtiella coani</i>	1	<i>Magelona sacculata</i>	16
<i>Leptochelia dubia</i>	1	<i>Micronephtys cornuta</i>	16
<i>Macoma</i> spp.	1	<i>Photis parvidons</i>	12
<i>Magelona sacculata</i>	1	<i>Macoma</i> spp.	11
<i>Maldanidae</i>	1	<i>Onuphidae</i>	11
<i>Monostylifera</i> sp. B	1	<i>Podarkeopsis glabrus</i>	11
<i>Nassariidae</i>	1	<i>Tritella pilimana</i>	10
<i>Onuphidae</i>	1	<i>Dendrochirotida</i>	9
<i>Pentamera rigida</i>	1	<i>Euphilomedes carcharodonta</i>	9
<i>Photis parvidons</i>	1	<i>Magelona hartmanae</i>	9
<i>Phyllodoce hartmanae</i>	1	<i>Spiophanes berkeleyorum</i>	9
<i>Phyllodoce williamsi</i>	1	<i>Pacifoculodes barnardi</i>	8
<i>Prionospio lighti</i>	1	<i>Phylo felix</i>	8
<i>Scolecopsis</i> sp. SF1	1	<i>Caesia rhinetes</i>	7
<i>Scolecopsis squamata</i>	1	<i>Ischyrocerus pelagops</i>	7
<i>Terebellidae</i>	1	<i>Scoloplos</i> sp. SF1	7
<b>STATION 04</b>		<i>Tenonia priops</i>	7
<i>Spiophanes norrisi</i>	2168	<i>Bathycopea daltonae</i>	6
<i>Photis</i> spp.	322	<i>Glycera macrobranchia</i>	6
<i>Callianax pycna</i>	235	<i>Amphiodia</i> spp.	5
<i>Protomedeia penates</i>	226	<i>Odostomia</i> spp.	5
<i>Owenia collaris</i>	178	<i>Phyllodoce hartmanae</i>	5
<i>Scoletoma luti</i>	167	<i>Terebellidae</i>	5
<i>Photis macinerneyi</i>	143	<i>Ampelisca careyi</i>	4
<i>Pectinaria californiensis</i>	135	<i>Cylichna</i> spp.	4
<i>Mediomastus</i> spp.	69	<i>Eteone</i> sp. SF4	4
<i>Tellina modesta</i>	69	<i>Halcampa decemtentaculata</i>	4
<i>Glycinde</i> spp.	52	<i>Paradialychone eiffelturris</i>	4
<i>Glycinde picta</i>	33	<i>Amphiodia digitata</i>	3
<i>Apoprionospio pygmaea</i>	28	<i>Eumida longicornuta</i>	3
<i>Lineidae</i>	28	<i>Modiolus capax</i>	3
<i>Diastylopsis dawsoni</i>	25	<i>Rhepoxynius fatigans</i>	3
<i>Onuphis</i> spp.	23	<i>Sthenelais verruculosa</i>	3
<i>Leukoma staminea</i>	22	<i>Americhelidium shoemakeri</i>	2
<i>Nephtys caecoides</i>	20	<i>Ampelisca cristata</i>	2

Appendix E-2 (cont.)  
*Benthic infauna collected in 2012*

<i>Aphelochaeta petersenae</i>	2	<b>STATION 06</b>	
<i>Aricidea (Acmira) catherinae</i>	2	<i>Spiophanes norrisi</i>	7145
<i>Axinopsida serricata</i>	2	<i>Photis</i> spp.	771
<i>Clinocardium nuttallii</i>	2	<i>Photis macinerneyi</i>	373
<i>Crangon nigromaculata</i>	2	<i>Scoletoma luti</i>	119
<i>Kurtziella plumbea</i>	2	<i>Callianax pycna</i>	81
<i>Macoma acolasta</i>	2	<i>Ischyrocerus pelagops</i>	76
Maldanidae	2	<i>Glycinde picta</i>	64
<i>Microphthalmus</i> spp. complex	2	<i>Protomedea penates</i>	53
Nemertea	2	<i>Eumida longicornuta</i>	39
<i>Neotrypaea</i> spp.	2	<i>Pleurogonium</i> sp. SF1	31
<i>Paranemertes californica</i>	2	<i>Carinoma mutabilis</i>	27
<i>Synidotea consolidata</i>	2	<i>Diastylopsis dawsoni</i>	17
<i>Tresus</i> spp.	2	<i>Glycinde</i> spp.	17
<i>Yoldia cooperii</i>	2	<i>Photis parvidons</i>	15
Ampharetidae	1	<i>Magelona hartmanae</i>	11
Anthozoa	1	<i>Mediomastus</i> spp.	10
<i>Argissa hamatipes</i>	1	<i>Onuphis</i> sp. A	10
<i>Astyris gausapata</i>	1	<i>Pacifocolodes barnardi</i>	10
<i>Carinoma mutabilis</i>	1	<i>Onuphis</i> spp.	9
<i>Dendraster excentricus</i>	1	<i>Amphiodia</i> spp.	8
<i>Diopatra ornata</i>	1	<i>Magelona sacculata</i>	8
<i>Enteropneusta</i>	1	<i>Tritella pilimana</i>	6
<i>Eteone fauchaldi</i>	1	<i>Podarkeopsis glabrus</i>	5
<i>Euchone hancocki</i>	1	<i>Tenonia priops</i>	5
<i>Euclymeninae</i> sp. SF1	1	<i>Diastylis santamariensis</i>	4
<i>Gastropteron pacificum</i>	1	<i>Euphilomedes carcharodonta</i>	4
<i>Goniada maculata</i>	1	<i>Glycera macrobranchia</i>	4
<i>Heteroclidus punctatus</i>	1	<i>Nephtys caecoides</i>	4
Holothuroidea	1	<i>Leitoscoloplos pugettensis</i>	3
<i>Kurtiella tumida</i>	1	<i>Lumbrineris californiensis</i>	3
<i>Lanassa venusta</i>	1	<i>Phyllodoce hartmanae</i>	3
<i>Leitoscoloplos pugettensis</i>	1	<i>Phylo felix</i>	3
Lumbrineridae	1	<i>Rhepoxynius fatigans</i>	3
<i>Pinnixa</i> spp.	1	<i>Tellina modesta</i>	3
Polynoidae	1	<i>Americhelidium shoemakeri</i>	2
<i>Spiochaetopterus costarum</i>	1	<i>Gastropteron pacificum</i>	2
<i>Tubulanus</i> spp.	1	<i>Spiochaetopterus costarum</i>	2
Turbellaria	1	Terebellidae	2
		<i>Amphiodia digitata</i>	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Apoprionospio pygmaea</i>	1	<i>Mediomastus</i> spp.	23
<i>Argissa hamatipes</i>	1	<i>Diastylopsis dawsoni</i>	21
<i>Aricidea</i> ( <i>Aedicira</i> ) sp. A	1	<i>Onuphis</i> sp. A	17
<i>Aricidea</i> ( <i>Aricidea</i> ) sp. SF2	1	<i>Tritella pilimana</i>	15
<i>Aricidea</i> ( <i>Aricidea</i> ) sp. SF3	1	<i>Glycera macrobranchia</i>	14
<i>Armandia brevis</i>	1	<i>Tenonia priops</i>	13
<i>Astyris gausapata</i>	1	Onuphidae	12
<i>Crangon nigromaculata</i>	1	<i>Pacifoculodes barnardi</i>	12
<i>Cylichna</i> spp.	1	<i>Pandora bilirata</i>	12
<i>Eteone fauchaldi</i>	1	<i>Kurtiella tumida</i>	11
<i>Glycinde</i> sp. SF1	1	<i>Nephtys caecoides</i>	11
<i>Kurtiella tumida</i>	1	Lineidae	10
<i>Kurtziella plumbea</i>	1	<i>Magelona hartmanae</i>	10
<i>Neotrypaea</i> spp.	1	<i>Bathycopea daltonae</i>	9
<i>Nephtys</i> spp.	1	<i>Paranemertes californica</i>	8
<i>Phyllodoce williamsi</i>	1	<i>Euphilomedes carcharodonta</i>	7
Polynoidae	1	<i>Leitoscoloplos pugettensis</i>	7
Sipuncula	1	Nemertea	7
<i>Tubulanidae</i> sp. B	1	<i>Cylichna</i> spp.	6
<i>Tubulanus pellucidus</i>	1	<i>Eumida longicornuta</i>	6
<b>STATION 25</b>		<i>Leukoma staminea</i>	6
<i>Spiophanes norrisi</i>	3867	<i>Clinocardium nuttallii</i>	5
<i>Photis</i> spp.	1504	<i>Odostomia</i> spp.	5
<i>Photis macinerneyi</i>	613	<i>Synidotea consolidata</i>	5
<i>Protomedea penates</i>	495	<i>Americhelidium shoemakeri</i>	4
<i>Scoletoma luti</i>	147	<i>Carinoma mutabilis</i>	4
<i>Owenia collaris</i>	123	<i>Halcampa decemtentaculata</i>	4
<i>Glycinde</i> spp.	93	<i>Magelona sacculata</i>	4
<i>Tellina modesta</i>	80	<i>Dendraster excentricus</i>	3
<i>Pectinaria californiensis</i>	76	<i>Dendrochirotida</i>	3
<i>Callianax pycna</i>	56	<i>Gastropteron pacificum</i>	3
<i>Ischyrocerus pelagops</i>	49	Mactridae	3
<i>Pleurogonium</i> sp. SF1	47	<i>Micronephtys cornuta</i>	3
<i>Diastylis santamariensis</i>	41	<i>Pista wui</i>	3
<i>Glycinde picta</i>	36	<i>Rhepoxynius fatigans</i>	3
<i>Onuphis</i> spp.	27	<i>Amphiodia digitata</i>	2
<i>Amphiodia</i> spp.	26	Cardiidae	2
<i>Photis parvidons</i>	26	<i>Edotia sublittoralis</i>	2
<i>Astyris gausapata</i>	25	<i>Eteone fauchaldi</i>	2
<i>Macoma nasuta</i>	25	<i>Glycinde</i> sp. SF1	2



Appendix E-2 (cont.)  
Benthic infauna collected in 2012

Lumbrineridae	2	<b>STATION 28</b>	
<i>Mesolamprops dillonensis</i>	2	<i>Spiophanes norrisi</i>	576
<i>Paradialychone eiffelturris</i>	2	<i>Owenia collaris</i>	293
<i>Podarkeopsis glabrus</i>	2	<i>Mediomastus</i> spp.	193
Polynoidae	2	<i>Scoletoma luti</i>	168
<i>Sthenelais verruculosa</i>	2	<i>Pectinaria californiensis</i>	130
Terebellidae	2	<i>Apoprionospio pygmaea</i>	93
<i>Ampelisca cristata</i>	1	<i>Tellina modesta</i>	77
<i>Ampharete acutifrons</i>	1	<i>Magelona sacculata</i>	66
Anthozoa	1	<i>Protomedeia penates</i>	57
<i>Apoprionospio pygmaea</i>	1	<i>Photis</i> spp.	45
<i>Argissa hamatipes</i>	1	<i>Diastylopsis dawsoni</i>	43
Bivalvia	1	<i>Callianax pycna</i>	35
<i>Chaetozone columbiana</i>	1	<i>Glycinde</i> spp.	25
<i>Diaphana californica</i>	1	<i>Photis macinerneyi</i>	25
Gastropoda	1	Terebellidae	24
<i>Goniada maculata</i>	1	<i>Micronephtys cornuta</i>	23
<i>Mactromeris catilliformis</i>	1	<i>Onuphis</i> sp. A	20
Maldanidae	1	<i>Pacifocolodes barnardi</i>	20
<i>Mesochaetopterus</i> sp. SF1	1	<i>Rhepoxynius lucubrans</i>	20
<i>Modiolus capax</i>	1	<i>Glycinde picta</i>	19
<i>Neotrypaea</i> spp.	1	<i>Amphiodia</i> spp.	18
Nereididae	1	<i>Macoma nasuta</i>	18
<i>Paracaudina chilensis</i>	1	<i>Magelona hartmanae</i>	16
<i>Phyllodoce groenlandica</i>	1	<i>Diastylis santamariensis</i>	15
<i>Phyllodoce longipes</i>	1	<i>Leitoscoloplos pugettensis</i>	15
<i>Phylo felix</i>	1	<i>Paradialychone eiffelturris</i>	15
<i>Rhepoxynius lucubrans</i>	1	<i>Prionospio lighti</i>	14
<i>Sigalion spinosus</i>	1	<i>Nephtys caecoides</i>	12
<i>Solen sicarius</i>	1	<i>Spiophanes berkeleyorum</i>	12
<i>Spiochaetopterus costarum</i>	1	Dendrochirotida	10
<i>Stylatula</i> spp.	1	Lumbrineridae	10
<i>Tiron biocellata</i>	1	Nemertea	10
<i>Tresus</i> spp.	1	Onuphidae	10
Tubulanidae sp. B	1	<i>Rhepoxynius fatigans</i>	10
<i>Tubulanus nothus</i>	1	<i>Kurtiella tumida</i>	9
<i>Tubulanus pellucidus</i>	1	<i>Leukoma staminea</i>	9
<i>Typosyllis farallonensis</i>	1	<i>Pleurogonium</i> sp. SF1	9
		<i>Cylichna attonsa</i>	8
		<i>Modiolus capax</i>	8

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Scoloplos</i> spp.	2	<i>Pleurogonium</i> sp. SF1	3
<i>Spiophanes norrisi</i>	2	<i>Prionospio lighti</i>	3
<i>Aricidea</i> ( <i>Aricidea</i> ) sp. SF3	1	<i>Apoprionospio pygmaea</i>	2
Cardiidae	1	<i>Carinoma mutabilis</i>	2
Lineidae	1	Cirratulidae	2
<i>Mandibulophoxus gilesi</i>	1	Enteropneusta	2
Orbiniidae	1	Euclymeninae	2
Paraonidae	1	<i>Glycinde</i> spp.	2
<i>Phoronis</i> spp.	1	<i>Listriella goleta</i>	2
<i>Photis macinerneyi</i>	1	<i>Macoma nasuta</i>	2
<i>Tecticeps convexus</i>	1	<i>Nephtys caecoides</i>	2
<b>STATION 32</b>		Paraonidae	2
<i>Rhepoxynius fatigans</i>	147	<i>Rhepoxynius lucubrans</i>	2
<i>Mediomastus</i> spp.	118	<i>Spiophanes norrisi</i>	2
<i>Magelona sacculata</i>	41	<i>Ampharete acutifrons</i>	1
<i>Protomedeia penates</i>	21	<i>Amphiodia</i> spp.	1
<i>Leitoscoloplos pugettensis</i>	19	<i>Callianax pycna</i>	1
<i>Tellina modesta</i>	16	<i>Chone mollis</i>	1
<i>Mactromeris catilliformis</i>	13	Corophoidea	1
<i>Pectinaria californiensis</i>	12	<i>Diastylopsis dawsoni</i>	1
<i>Scoletoma luti</i>	12	<i>Glossaluax reclusiana</i>	1
<i>Spiophanes berkeleyorum</i>	12	<i>Hesperonoe laevis</i>	1
<i>Magelona hartmanae</i>	11	Maldanidae	1
<i>Onuphis</i> sp. A	11	<i>Malmgreniella</i> spp.	1
Nemertea	10	Onuphidae	1
<i>Kurtiella coani</i>	8	<i>Phoronis</i> spp.	1
<i>Kurtiella tumida</i>	7	<i>Phyllodoce hartmanae</i>	1
<i>Ampelisca cristata</i>	6	<i>Phylo felix</i>	1
<i>Glycinde picta</i>	6	Polynoidae	1
<i>Amaeana occidentalis</i>	5	<i>Scoloplos</i> sp. SF1	1
<i>Aphelochaeta petersenae</i>	5	<i>Spiophanes</i> spp.	1
<i>Neotrypaea</i> spp.	5	Terebellidae	1
<i>Americhelidium shoemakeri</i>	4	<i>Tubulanus pellucidus</i>	1
<i>Cylichna attonsa</i>	4	<i>Turbonilla</i> spp.	1
Lineidae	4	<b>STATION 33</b>	
<i>Pacifoculodes barnardi</i>	4	<i>Spiophanes norrisi</i>	807
<i>Paraprionospio alata</i>	4	<i>Mediomastus</i> spp.	146
<i>Podarkeopsis glabrus</i>	4	<i>Protomedeia penates</i>	140
<i>Glycera macrobranchia</i>	3	<i>Rhepoxynius fatigans</i>	88
<i>Odostomia</i> spp.	3	<i>Stylatula</i> spp.	76

Appendix E-2 (cont.)  
*Benthic infauna collected in 2012*

<i>Scoletoma luti</i>	53	<i>Americhelidium shoemakeri</i>	2
<i>Photis</i> spp.	37	<i>Apoprionospio pygmaea</i>	2
<i>Mactromeris catilliformis</i>	33	Cardiidae	2
Pennatulacea	31	<i>Cylichna attonsa</i>	2
<i>Pleurogonium</i> sp. SF1	25	<i>Dendrochirotida</i>	2
<i>Tellina modesta</i>	25	<i>Euchone hancocki</i>	2
<i>Glycinde</i> spp.	18	<i>Ischyrocerus pelagops</i>	2
<i>Onuphis</i> sp. A	17	<i>Macoma nasuta</i>	2
<i>Diastylopsis dawsoni</i>	16	<i>Mediomastus acutus</i>	2
<i>Leukoma staminea</i>	15	<i>Mesochaetopterus</i> sp. SF1	2
<i>Pectinaria californiensis</i>	15	Nemertea	2
<i>Onuphis</i> spp.	14	<i>Neotrypaea</i> spp.	2
<i>Glycinde picta</i>	13	<i>Photis parvidons</i>	2
<i>Photis macinerneyi</i>	13	<i>Podarkeopsis glabrus</i>	2
<i>Aphelochaeta petersenae</i>	11	Sabellidae	2
<i>Clinocardium nuttallii</i>	11	<i>Tubulanus pellucidus</i>	2
<i>Phoronis</i> spp.	11	<i>Ampharete labrops</i>	1
<i>Nephtys caecoides</i>	10	Autolytinae	1
Enteropneusta	9	<i>Carinoma mutabilis</i>	1
<i>Diastylis santamariensis</i>	8	<i>Cheirimedeia zotea</i>	1
<i>Bathycopea daltonae</i>	7	<i>Euclymeninae</i> sp. SF1	1
Lineidae	7	<i>Eumida longicornuta</i>	1
<i>Amphiodia</i> spp.	6	<i>Glossaluax reclusiana</i>	1
<i>Magelona sacculata</i>	6	<i>Glycera macrobranchia</i>	1
<i>Amaeana occidentalis</i>	5	<i>Heteromastus</i> spp.	1
<i>Callianax pycna</i>	5	<i>Kurtziella plumbea</i>	1
<i>Odostomia</i> spp.	5	Lumbrineridae	1
<i>Scoloplos</i> sp. SF1	5	<i>Lumbrineris californiensis</i>	1
<i>Ampelisca cristata</i>	4	<i>Macoma acolasta</i>	1
<i>Ampharete acutifrons</i>	4	<i>Magelona hartmanae</i>	1
Maldanidae	4	<i>Malmgreniella liei</i>	1
<i>Pandora bilirata</i>	4	<i>Malmgreniella</i> spp.	1
<i>Compsomyx subdiaphana</i>	3	<i>Mesolamprops dillonensis</i>	1
<i>Dendraster excentricus</i>	3	<i>Metacarcinus gracilis</i>	1
<i>Goniada maculata</i>	3	<i>Micronephtys cornuta</i>	1
<i>Lanassa venusta</i>	3	Nassariidae	1
<i>Leitoscoloplos pugettensis</i>	3	Onuphidae	1
<i>Macoma</i> spp.	3	<i>Paradialychone eiffelturris</i>	1
<i>Solen sicarius</i>	3	<i>Polycirrus</i> sp. I	1
<i>Tenonia priops</i>	3	Polynoidae	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Sthenelais verruculosa</i>	1	<i>Phylo felix</i>	5
<i>Synidotea consolidata</i>	1	<i>Streblosoma</i> sp. SF1	5
Terebellidae	1	Terebellidae	5
<i>Yoldia cooperii</i>	1	<i>Ampelisca cristata</i>	4
<b>STATION 34</b>		<i>Astyris gausapata</i>	4
<i>Mediomastus</i> spp.	307	<i>Leukoma staminea</i>	4
<i>Protomedeia penates</i>	148	<i>Pandora bilirata</i>	4
<i>Stylatula</i> spp.	78	<i>Paraprionospio alata</i>	4
<i>Macoma nasuta</i>	52	<i>Pholoe glabra</i>	4
<i>Carazziella</i> sp. A	50	<i>Pista wui</i>	4
<i>Kurtiella tumida</i>	49	<i>Podarkeopsis glabrus</i>	4
<i>Amaeana occidentalis</i>	43	<i>Saccella</i> spp.	4
<i>Scoletoma luti</i>	38	<i>Callianax pycna</i>	3
<i>Rhepoxynius lucubrans</i>	33	<i>Clinocardium nuttallii</i>	3
<i>Aphelochaeta petersenae</i>	26	<i>Dipolydora</i> spp.	3
<i>Rhepoxynius fatigans</i>	24	<i>Lepidasthenia berkeleyae</i>	3
<i>Spiophanes berkeleyorum</i>	23	<i>Lumbrineris californiensis</i>	3
<i>Amphiodia</i> spp.	22	<i>Odostomia</i> spp.	3
<i>Tellina modesta</i>	22	<i>Prionospio lighti</i>	3
<i>Leitoscoloplos pugettensis</i>	19	<i>Sigambra</i> sp. SF2	3
Polynoidae	19	<i>Ampelisca careyi</i>	2
<i>Glycinde</i> spp.	17	<i>Ampharete</i> spp.	2
Nematoda	16	<i>Apoprionospio pygmaea</i>	2
<i>Pectinaria californiensis</i>	16	<i>Axiothella rubrocincta</i>	2
<i>Amphiteis scaphobranchiata</i>	13	<i>Crangon nigromaculata</i>	2
Corophiidae	13	<i>Eteone ?californica</i>	2
Dendrochirotida	13	<i>Kurtiella coani</i>	2
<i>Diastylopsis dawsoni</i>	12	<i>Malmgreniella</i> spp.	2
<i>Micronephtys cornuta</i>	11	<i>Modiolus capax</i>	2
<i>Pleurogonium</i> sp. SF1	10	<i>Nereis neoneanthes</i>	2
<i>Magelona hartmanae</i>	9	<i>Paradialychone eiffelturris</i>	2
Nemertea	9	Paraonidae	2
Euclymeninae sp. SF1	8	<i>Philine auriformis</i>	2
<i>Onuphis</i> sp. A	8	<i>Pinnixa franciscana</i>	2
<i>Ampharete acutifrons</i>	6	<i>Scoloplos</i> sp. SF1	2
<i>Glycinde picta</i>	6	<i>Solen sicarius</i>	2
<i>Tenonia priops</i>	6	<i>Sthenelais verruculosa</i>	2
<i>Cylichna attonsa</i>	5	<i>Tubulanus nothus</i>	2
<i>Pacifoculodes barnardi</i>	5	<i>Aricidea (Aricidea)</i> sp. SF3	1
<i>Photis</i> spp.	5	<i>Axinopsida serricata</i>	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Cooperella subdiaphana</i>	1	<i>Paraprionospio alata</i>	9
<i>Edwardsia juliae</i>	1	<i>Kurtiella tumida</i>	8
Enteropneusta	1	<i>Magelona hartmanae</i>	8
Gastropoda	1	<i>Modiolus rectus</i>	7
<i>Glossaluax reclusiana</i>	1	<i>Ischyrocerus pelagops</i>	6
<i>Haliophasma geminatum</i>	1	<i>Leitoscoloplos pugettensis</i>	6
<i>Ischyrocerus pelagops</i>	1	<i>Macoma nasuta</i>	6
<i>Kurtzina beta</i>	1	Nematoda	6
Lineidae	1	Dendrochirotida	5
<i>Mactromeris catilliformis</i>	1	<i>Micronephtys cornuta</i>	5
<i>Neotrypaea</i> spp.	1	Nemertea	5
<i>Nephtys caecoides</i>	1	<i>Nutricula confusa</i>	5
Onuphidae	1	<i>Onuphis</i> spp.	5
<i>Onuphis</i> spp.	1	<i>Cheirimeidia zotea</i>	4
<i>Pachynus barnardi</i>	1	<i>Dipolydora magna</i>	4
<i>Petaloclymene pacifica</i>	1	<i>Magelona sacculata</i>	4
<i>Pherusa neopapillata</i>	1	<i>Onuphis</i> sp. A	4
<i>Poecilochaetus johnsoni</i>	1	<i>Photis macinerneyi</i>	4
<i>Turbonilla</i> spp.	1	<i>Amphiteis scaphobranchiata</i>	3
<i>Yoldia cooperii</i>	1	<i>Apoprionospio pygmaea</i>	3
<b>STATION 35</b>		<i>Axinopsida serricata</i>	3
<i>Diastylopsis dawsoni</i>	290	<i>Munnogonium tillerae</i>	3
<i>Mediomastus</i> spp.	165	Onuphidae	3
<i>Owenia collaris</i>	67	<i>Americhelidium shoemakeri</i>	2
<i>Ampelisca cristata</i>	60	<i>Ampharete acutifrons</i>	2
<i>Protomedeia penates</i>	49	<i>Aricidea (Acmira) catherinae</i>	2
<i>Pectinaria californiensis</i>	33	Bivalvia	2
<i>Spiophanes berkeleyorum</i>	33	<i>Clinocardium nuttallii</i>	2
<i>Amphiodia</i> spp.	32	<i>Cylichna</i> spp.	2
<i>Callianax pycna</i>	24	Euclymeninae sp. SF1	2
<i>Tellina modesta</i>	21	Lineidae	2
<i>Sigambra</i> sp. SF2	20	Maldanidae	2
<i>Diastylis santamariensis</i>	17	<i>Neotrypaea</i> spp.	2
<i>Stylatula</i> spp.	17	<i>Nephtys caecoides</i>	2
<i>Aphelochaeta petersenae</i>	14	<i>Odostomia</i> spp.	2
<i>Pleurogonium</i> sp. SF1	14	Ostracoda	2
<i>Glycinde</i> spp.	13	<i>Pholoe glabra</i>	2
<i>Photis</i> spp.	13	<i>Prionospio lighti</i>	2
<i>Astyris gausapata</i>	12	<i>Scoletoma luti</i>	2
<i>Glycinde picta</i>	10	<i>Scoloplos</i> sp. SF1	2

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Siliqua lucida</i>	2	<i>Pleurogonium</i> sp. SF1	11
<i>Ampharete labrops</i>	1	<i>Callianax pycna</i>	10
<i>Caesia rhinetes</i>	1	<i>Photis</i> spp.	9
<i>Carinoma</i> sp.	1	<i>Apoprionospio pygmaea</i>	8
<i>Cylichna attonsa</i>	1	<i>Aphelochaeta petersenae</i>	7
Echiura	1	<i>Leitoscoloplos pugettensis</i>	7
Enteropneusta	1	<i>Stylatula</i> spp.	6
<i>Glossaluax reclusiana</i>	1	Dendrochirotida	5
<i>Listriella diffusa</i>	1	<i>Leukoma staminea</i>	5
Lumbrineridae	1	<i>Macoma nasuta</i>	5
<i>Mactromeris catilliformis</i>	1	<i>Malmgreniella</i> spp.	5
<i>Magelona berkeleyi</i>	1	<i>Kurtiella tumida</i>	4
<i>Mesolamprops dillonensis</i>	1	Nemertea	4
<i>Modiolus</i> spp.	1	<i>Bathycopea daltonae</i>	3
<i>Monocorophium acherusicum</i>	1	Cardiidae	3
<i>Pachynus barnardi</i>	1	<i>Dendraster excentricus</i>	3
<i>Paranemertes californica</i>	1	Onuphidae	3
<i>Phoronis</i> spp.	1	<i>Photis macinerneyi</i>	3
<i>Rhepoxynius fatigans</i>	1	<i>Amaeana occidentalis</i>	2
<i>Saccella taphria</i>	1	<i>Ampharete acutifrons</i>	2
<i>Sthenelais verruculosa</i>	1	<i>Carinoma mutabilis</i>	2
<i>Tenonia priops</i>	1	<i>Cheirimeдея zotea</i>	2
Terebellidae	1	Isaeidae	2
<b>STATION 36</b>		<i>Magelona hartmanae</i>	2
<i>Spiophanes norrisi</i>	567	<i>Modiolus capax</i>	2
<i>Protomeдея penates</i>	172	<i>Odostomia</i> spp.	2
<i>Scoletoma luti</i>	143	<i>Pacifocolodes barnardi</i>	2
<i>Tellina modesta</i>	78	<i>Scoloplos</i> sp. SF1	2
<i>Mactromeris catilliformis</i>	63	<i>Streblosoma</i> spp.	2
<i>Pectinaria californiensis</i>	42	<i>Synidotea consolidata</i>	2
<i>Mediomastus</i> spp.	38	Ampharetidae	1
<i>Glycinde</i> spp.	28	<i>Amphiodia digitata</i>	1
<i>Onuphis</i> sp. A	27	<i>Amphiodia</i> spp.	1
<i>Rhepoxynius fatigans</i>	21	<i>Aricidea (Aedicira) pacifica</i>	1
<i>Glycinde picta</i>	16	<i>Axinopsida serricata</i>	1
<i>Owenia collaris</i>	14	Bivalvia	1
<i>Diastylopsis dawsoni</i>	12	<i>Cheirophotis</i> spp.	1
<i>Magelona sacculata</i>	12	<i>Diastylis santamariensis</i>	1
Enteropneusta	11	<i>Eteone ?californica</i>	1
<i>Nephtys caecoides</i>	11	<i>Glossaluax reclusiana</i>	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Hemilamprops californicus</i>	1	<i>Kurtiella tumida</i>	6
<i>Ischyrocerus pelagops</i>	1	<i>Amphiodia digitata</i>	5
<i>Kurtziella plumbea</i>	1	<i>Nephtys caecoides</i>	5
<i>Lanassa venusta</i>	1	<i>Stylatula</i> spp.	5
<i>Lepidasthenia berkeleyae</i>	1	<i>Americhelidium shoemakeri</i>	4
<i>Macoma acolasta</i>	1	<i>Aphelochaeta petersenae</i>	4
<i>Magelona berkeleyi</i>	1	<i>Apoprionospio pygmaea</i>	4
<i>Micronephtys cornuta</i>	1	<i>Cylichna attonsa</i>	4
Mysidacea	1	<i>Diastylopsis dawsoni</i>	4
<i>Pandora bilirata</i>	1	<i>Mactromeris catilliformis</i>	4
<i>Paradialychone eiffelturris</i>	1	<i>Onuphis</i> sp. A	4
<i>Photis brevipes</i>	1	<i>Tenonia priops</i>	4
<i>Photis parvidons</i>	1	<i>Cerebratulus californiensis</i>	3
<i>Phyllodoce</i> spp.	1	<i>Cylichna</i> spp.	3
Polynoidae	1	<i>Dendraster excentricus</i>	3
<i>Prionospio lighti</i>	1	<i>Kurtiella coani</i>	3
<i>Rictaxis punctocaelatus</i>	1	<i>Pleurogonium</i> sp. SF1	3
Sabellidae	1	<i>Glycera macrobranchia</i>	2
<i>Streblosoma</i> sp. SF1	1	<i>Harmothoe</i> spp.	2
<i>Tenonia priops</i>	1	<i>Magelona hartmanae</i>	2
Terebellidae	1	<i>Onuphis</i> spp.	2
<i>Tubulanus pellucidus</i>	1	<i>Pacificolodes barnardi</i>	2
<b>STATION 37</b>		<i>Phylo felix</i>	2
<i>Spiophanes norrisi</i>	488	<i>Amaeana occidentalis</i>	1
<i>Protomedeia penates</i>	98	<i>Ampharete acutifrons</i>	1
<i>Mediomastus</i> spp.	74	Ampharetidae	1
<i>Callianax pycna</i>	53	<i>Aricidea</i> spp.	1
<i>Scoletoma luti</i>	53	Bivalvia	1
<i>Tellina modesta</i>	25	<i>Caesia rhinetes</i>	1
<i>Carinoma mutabilis</i>	17	Cirratulidae	1
<i>Glycinde picta</i>	13	<i>Clinocardium nuttallii</i>	1
<i>Glycinde</i> spp.	12	Corophoidea	1
<i>Owenia collaris</i>	12	<i>Eteone fauchaldi</i>	1
<i>Axiothella rubrocincta</i>	11	<i>Euphilomedes carcharodonta</i>	1
<i>Leitoscoloplos pugettensis</i>	10	<i>Kurtzina beta</i>	1
<i>Magelona sacculata</i>	7	Lineidae	1
<i>Pectinaria californiensis</i>	7	<i>Macoma nasuta</i>	1
<i>Photis</i> spp.	7	<i>Modiolus capax</i>	1
<i>Bathycopea daltonae</i>	6	<i>Modiolus rectus</i>	1
Enteropneusta	6	<i>Neotrypaea</i> spp.	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Odostomia</i> spp.	1	<i>Carinoma mutabilis</i>	5
Onuphidae	1	Lineidae	5
<i>Paraonella platybranchiata</i>	1	<i>Micronephlys cornuta</i>	5
<i>Phoronis</i> spp.	1	<i>Odostomia</i> spp.	5
<i>Photis macinerneyi</i>	1	<i>Cylichna</i> spp.	4
<i>Podarkeopsis glabrus</i>	1	<i>Glycinde picta</i>	4
<i>Sinocorophium heteroceratum</i>	1	<i>Pista wui</i>	4
<i>Spiophanes berkeleyorum</i>	1	<i>Cheirimeidia zotea</i>	3
<i>Synidotea consolidata</i>	1	<i>Halcampa decementaculata</i>	3
Terebellidae	1	<i>Kurtiella tumida</i>	3
<b>STATION 38</b>		<i>Leitoscoloplos pugettensis</i>	3
<i>Spiophanes norrisi</i>	191	Maldanidae	3
<i>Protomeidia penates</i>	132	<i>Pandora bilirata</i>	3
<i>Diastylopsis dawsoni</i>	87	<i>Photis brevipes</i>	3
<i>Scoletoma luti</i>	58	<i>Photis parvidons</i>	3
<i>Mediomastus</i> spp.	53	<i>Phyllodoce hartmanae</i>	3
<i>Callianax pycna</i>	42	<i>Pista</i> spp.	3
<i>Tellina modesta</i>	35	<i>Spiophanes</i> spp.	3
<i>Glycinde</i> spp.	28	<i>Sthenelais verruculosa</i>	3
<i>Photis</i> spp.	28	<i>Tenonia priops</i>	3
<i>Rhepoxynius fatigans</i>	26	<i>Tubulanus pellucidus</i>	3
<i>Stylatula</i> spp.	19	<i>Amaeana occidentalis</i>	2
<i>Onuphis</i> sp. A	18	<i>Americhelidium shoemakeri</i>	2
<i>Amphiodia</i> spp.	17	<i>Axinopsida serricata</i>	2
<i>Onuphis</i> spp.	17	<i>Diaphana californica</i>	2
<i>Astyris gausapata</i>	13	<i>Diastylis santamariensis</i>	2
Dendrochirotida	13	<i>Edwardsia juliae</i>	2
<i>Aphelochaeta petersenae</i>	11	Enteropneusta	2
<i>Scoloplos</i> sp. SF1	10	<i>Ischyrocerus pelagops</i>	2
<i>Ampelisca careyi</i>	8	<i>Kurtziella plumbea</i>	2
<i>Apoprionospio pygmaea</i>	8	<i>Leukoma staminea</i>	2
<i>Pleurogonium</i> sp. SF1	8	<i>Magelona hartmanae</i>	2
<i>Ampelisca cristata</i>	7	<i>Magelona</i> spp.	2
<i>Pectinaria californiensis</i>	7	Onuphidae	2
Pennatulacea	7	<i>Onuphis</i> spp.	2
Terebellidae	7	<i>Owenia collaris</i>	2
<i>Clinocardium nuttallii</i>	6	<i>Paraprionospio alata</i>	2
<i>Lanassa venusta</i>	6	Polynoidae	2
<i>Nephtys caecoides</i>	6	<i>Rictaxis punctocaelatus</i>	2
<i>Amphicteis scaphobranchiata</i>	5	<i>Spiophanes berkeleyorum</i>	2



Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Ampharetidae</i>	1	<i>Aphelochaeta</i> sp. SF3	7
<i>Aricidea</i> ( <i>Aricidea</i> ) sp. SF3	1	<i>Glycinde picta</i>	7
<i>Caesia rhinetes</i>	1	Maldanidae	7
<i>Chaetozone columbiana</i>	1	<i>Nephtys caecoides</i>	7
<i>Cylichna attonsa</i>	1	<i>Odostomia</i> spp.	7
<i>Gastropteron pacificum</i>	1	<i>Onuphis</i> sp. A	7
<i>Gnathopleustes pugettensis</i>	1	<i>Onuphis</i> spp.	7
<i>Goniada maculata</i>	1	<i>Ampelisca careyi</i>	5
<i>Macoma acolasta</i>	1	<i>Apoprionospio pygmaea</i>	5
<i>Mactromeris catilliformis</i>	1	<i>Hemilamprops californicus</i>	5
Mangeliidae	1	<i>Leukoma staminea</i>	5
<i>Nassarius</i> spp.	1	<i>Photis</i> spp.	5
<i>Neotrypaea</i> spp.	1	<i>Tenonia priops</i>	5
<i>Phyllodoce longipes</i>	1	<i>Cylichna</i> spp.	4
<i>Phylo felix</i>	1	<i>Goniada maculata</i>	4
<i>Tritella pilimana</i>	1	<i>Kurtiella tumida</i>	4
<b>STATION 39</b>		<i>Macoma nasuta</i>	4
<i>Mediomastus</i> spp.	133	<i>Phylo felix</i>	4
<i>Scoletoma luti</i>	82	<i>Amphiteis scaphobranchiata</i>	3
<i>Diastylopsis dawsoni</i>	63	<i>Euphilomedes carcharodonta</i>	3
<i>Spiophanes norrisi</i>	48	<i>Glycera macrobranchia</i>	3
<i>Glycinde</i> spp.	46	Lineidae	3
<i>Tellina modesta</i>	45	<i>Solen sicarius</i>	3
<i>Pista wui</i>	39	<i>Aricidea</i> ( <i>Aricidea</i> ) sp. SF3	2
<i>Protomedeia penates</i>	36	<i>Carinoma mutabilis</i>	2
<i>Callianax pycna</i>	32	<i>Cheirimedeia zotea</i>	2
<i>Rhepoxynius fatigans</i>	28	<i>Dyopedos arcticus</i>	2
<i>Astyris gausapata</i>	24	<i>Gastropteron pacificum</i>	2
Dendrochirotida	16	<i>Halcampa decententaculata</i>	2
<i>Lanassa venusta</i>	16	Inarticulata	2
<i>Amphiodia</i> spp.	15	<i>Leitoscoloplos pugettensis</i>	2
<i>Magelona sacculata</i>	12	<i>Modiolus capax</i>	2
<i>Clinocardium nuttallii</i>	10	Nematoda	2
<i>Magelona hartmanae</i>	10	Nemertea	2
Enteropneusta	9	<i>Neotrypaea</i> spp.	2
<i>Micronephtys cornuta</i>	9	<i>Pleurogonium</i> sp. SF1	2
<i>Pectinaria californiensis</i>	9	Polynoidae	2
<i>Stylatula</i> spp.	9	<i>Prionospio lighti</i>	2
<i>Chaetozone columbiana</i>	8	<i>Scoloplos</i> sp. SF1	2
<i>Ampelisca cristata</i>	7	Terebellidae	2

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Ampharete acutifrons</i>	1	<i>Mediomastus</i> spp.	26
<i>Ampharete</i> spp.	1	<i>Malmgreniella</i> spp.	16
Autolytinae	1	<i>Diastylopsis dawsoni</i>	15
<i>Axiothella rubrocincta</i>	1	Nemertea	13
<i>Compsomyax subdiaphana</i>	1	<i>Leitoscoloplos pugettensis</i>	11
<i>Cylichna attonsa</i>	1	<i>Kurtiella coani</i>	10
<i>Dendraster excentricus</i>	1	<i>Carinoma mutabilis</i>	9
<i>Diaphana californica</i>	1	<i>Modiolus rectus</i>	8
<i>Diastylis santamariensis</i>	1	<i>Onuphis</i> sp. A	8
<i>Diastylis</i> spp.	1	<i>Bathycopea daltonae</i>	7
<i>Edwardsia juliae</i>	1	<i>Protomeдея penates</i>	7
<i>Eteone ?californica</i>	1	<i>Glycera macrobranchia</i>	5
<i>Euchone hancocki</i>	1	<i>Magelona sacculata</i>	5
Flabelligeridae	1	<i>Neotrypaea</i> spp.	5
<i>Kurtziella plumbea</i>	1	<i>Edotia sublittoralis</i>	3
<i>Lepidasthenia berkeleyae</i>	1	<i>Owenia collaris</i>	3
Lumbrineridae	1	<i>Rhepoxynius vigitegus</i>	3
<i>Magelona berkeleyi</i>	1	<i>Americhelidium shoemakeri</i>	2
<i>Magelona</i> spp.	1	<i>Cerebratulus californiensis</i>	2
Nassariidae	1	<i>Kurtiella tumida</i>	2
<i>Pacifoculodes barnardi</i>	1	<i>Macoma nasuta</i>	2
<i>Pandora bilirata</i>	1	<i>Pacifoculodes barnardi</i>	2
<i>Paraprionospio alata</i>	1	<i>Pectinaria californiensis</i>	2
<i>Philine auriformis</i>	1	<i>Amphiodia</i> spp.	1
<i>Photis brevipes</i>	1	Anthozoa	1
<i>Scolecopsis</i> spp.	1	<i>Aphelochaeta petersenae</i>	1
<i>Spiochaetopterus costarum</i>	1	<i>Aphelochaeta</i> spp.	1
<i>Spiophanes</i> spp.	1	<i>Aricidea (Aedicira) pacifica</i>	1
<i>Sthenelais verruculosa</i>	1	<i>Dyopedos arcticus</i>	1
<i>Yoldia cooperii</i>	1	<i>Eobrolgus spinosus</i>	1
<b>STATION 40</b>		<i>Eteone ?californica</i>	1
<i>Spiophanes norrisi</i>	681	<i>Eteone</i> sp. SF3	1
<i>Callianax pycna</i>	194	<i>Ischyrocercus pelagops</i>	1
<i>Scoletoma luti</i>	76	<i>Isocheles pilosus</i>	1
<i>Photis macinerneyi</i>	39	<i>Mesochaetopterus</i> spp.	1
<i>Photis</i> spp.	39	<i>Microphthalmus</i> spp. complex	1
<i>Glycinde picta</i>	32	Nematoda	1
<i>Amaeana occidentalis</i>	31	<i>Nephtys caecoides</i>	1
<i>Tellina modesta</i>	30	<i>Nereis neoneanthes</i>	1
<i>Glycinde</i> spp.	26	<i>Paracaudina chilensis</i>	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Paranemertes californica</i>	1	<i>Apoprionospio pygmaea</i>	1
<i>Paraonella platybranchiata</i>	1	<i>Bathycopea daltonae</i>	1
<i>Podarkeopsis glabrus</i>	1	<i>Dendraster excentricus</i>	1
<i>Polycirrus</i> sp. I	1	<i>Diastylis santamariensis</i>	1
<i>Stylatula</i> spp.	1	Enteropneusta	1
<i>Tiron biocellata</i>	1	<i>Kurtiella tumida</i>	1
<i>Tubulanus pellucidus</i>	1	<i>Magelona sacculata</i>	1
<b>STATION 43</b>		<i>Malmgreniella</i> spp.	1
<i>Mandibulophoxus gilesi</i>	133	<i>Mandibulophoxus gilesi</i>	1
<i>Eohaustorius</i> spp.	38	Nematoda	1
<i>Americhelidium</i> sp. SD1	9	<i>Neotrypaea</i> spp.	1
<i>Kurtiella tumida</i>	9	<i>Nephtys caecoides</i>	1
<i>Rhepoxynius menziesi</i>	8	<i>Onuphis</i> sp. A	1
<i>Rhepoxynius vigitegus</i>	5	<i>Pectinaria californiensis</i>	1
<i>Callianax pycna</i>	4	Phoxocephalidae	1
<i>Carinoma mutabilis</i>	2	<i>Rhepoxynius fatigans</i>	1
<i>Caesia rhinetes</i>	1	<i>Scoletoma luti</i>	1
<i>Foxiphalus obtusidens</i>	1	<i>Sinocorophium heteroceratum</i>	1
<i>Magelona</i> spp.	1	<i>Tellina modesta</i>	1
<i>Nephtys caecoides</i>	1	<b>STATION 47</b>	
<i>Odostomia</i> spp.	1	<i>Callianax pycna</i>	584
<i>Pacifoculodes barnardi</i>	1	<i>Carinoma mutabilis</i>	25
<i>Scolelepis squamata</i>	1	<i>Diastylopsis dawsoni</i>	25
Syllidae	1	<i>Pacifoculodes barnardi</i>	16
<i>Tellina nukuloides</i>	1	<i>Mediomastus</i> spp.	12
<b>STATION 45</b>		<i>Rhepoxynius lucubrans</i>	12
<i>Callianax pycna</i>	1253	<i>Scoloplos armiger</i>	12
<i>Diastylopsis dawsoni</i>	44	<i>Eohaustorius</i> spp.	10
<i>Carinoma mutabilis</i>	40	<i>Phylo felix</i>	7
<i>Pacifoculodes barnardi</i>	17	<i>Tellina modesta</i>	5
<i>Spiophanes norrisi</i>	8	<i>Chaetozone columbiana</i>	4
<i>Chaetozone bansei</i>	4	<i>Tecticeps convexus</i>	4
<i>Mesolamprops dillonensis</i>	4	<i>Amaeana occidentalis</i>	3
<i>Protomedeia penates</i>	4	<i>Kurtiella</i> sp. SF1	3
<i>Scoloplos armiger</i>	4	<i>Rhepoxynius vigitegus</i>	3
<i>Photis</i> spp.	3	<i>Scoletoma luti</i>	3
<i>Amaeana occidentalis</i>	2	<i>Spiophanes norrisi</i>	3
<i>Glycinde</i> spp.	2	<i>Capitella capitata</i> complex	2
<i>Mediomastus</i> spp.	2	<i>Kurtiella coani</i>	2
<i>Tellina nukuloides</i>	2	<i>Magelona sacculata</i>	2

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Onuphis</i> sp. A	2	<i>Americhelidium shoemakeri</i>	1
<i>Photis</i> spp.	2	<i>Ampelisca careyi</i>	1
<i>Caesia fossatus</i>	1	<i>Cerebratulus</i> spp.	1
<i>Caesia rhinetes</i>	1	<i>Chaetozone bansei</i>	1
<i>Dendraster excentricus</i>	1	<i>Cooperella subdiaphana</i>	1
<i>Diastylis santamariensis</i>	1	Enteropneusta	1
<i>Nephtys caecoides</i>	1	<i>Glycera macrobranchia</i>	1
<i>Nereis neoneanthes</i>	1	<i>Kurtziella plumbea</i>	1
<i>Pagurus</i> spp.	1	Lineidae	1
Polynoidae	1	<i>Mesolamprops dillonensis</i>	1
<b>STATION 48</b>		Mysidacea	1
<i>Spiophanes norrisi</i>	957	<i>Neotrypaea</i> spp.	1
<i>Photis</i> spp.	176	<i>Photis brevipes</i>	1
<i>Diastylopsis dawsoni</i>	99	<i>Tenonia priops</i>	1
<i>Callianax pycna</i>	88	<i>Travisia gigas</i>	1
<i>Photis macinerneyi</i>	85	<i>Tritella pilimana</i>	1
<i>Protomedeia penates</i>	28	<b>STATION 50</b>	
<i>Glycinde</i> spp.	26	<i>Spiophanes norrisi</i>	585
<i>Pectinaria californiensis</i>	22	<i>Protomedeia penates</i>	241
<i>Glycinde picta</i>	19	<i>Diastylopsis dawsoni</i>	143
<i>Ischyrocerus pelagops</i>	11	<i>Rhepoxynius fatigans</i>	107
<i>Carinoma mutabilis</i>	10	<i>Callianax pycna</i>	68
<i>Mediomastus</i> spp.	7	<i>Glycinde</i> spp.	66
<i>Scoletoma luti</i>	7	Nematoda	64
Nemertea	6	<i>Tellina modesta</i>	58
<i>Edotia sublittoralis</i>	4	<i>Scoletoma luti</i>	54
Isaeidae	4	<i>Photis</i> spp.	49
<i>Pacifoculodes barnardi</i>	4	<i>Mediomastus</i> spp.	48
<i>Amphiodia</i> spp.	3	<i>Rhepoxynius abronius</i>	36
<i>Pleurogonium</i> sp. SF1	3	<i>Onuphis</i> sp. A	34
<i>Tellina nuculoides</i>	3	<i>Stylatula</i> spp.	21
<i>Apoprionospio pygmaea</i>	2	<i>Euphilomedes carcharodonta</i>	16
<i>Caesia rhinetes</i>	2	<i>Nephtys caecoides</i>	16
<i>Dendraster excentricus</i>	2	Dendrochirotida	14
Hydrozoa	2	<i>Magelona hartmanae</i>	14
<i>Kurtiella coani</i>	2	<i>Kurtiella tumida</i>	13
<i>Mesochaetopterus</i> sp. SF1	2	<i>Macoma nasuta</i>	13
<i>Synidotea consolidata</i>	2	<i>Glycinde picta</i>	12
<i>Tellina modesta</i>	2	<i>Micronephtys cornuta</i>	10
<i>Amaeana occidentalis</i>	1	<i>Ampelisca careyi</i>	9

Appendix E-2 (cont.)  
*Benthic infauna collected in 2012*

<i>Pectinaria californiensis</i>	9	<i>Pleurogonium</i> sp. SF1	2
<i>Photis macinerneyi</i>	8	<i>Tenonia priops</i>	2
<i>Amphiodia</i> spp.	7	<i>Ampelisca cristata</i>	1
<i>Apoprionospio pygmaea</i>	7	<i>Argissa hamatipes</i>	1
<i>Leukoma staminea</i>	7	<i>Aricidea (Acmira) horikoshii</i>	1
Nemertea	7	<i>Aricidea (Aedicira) pacifica</i>	1
<i>Cylichna</i> spp.	6	<i>Aricidea (Aricidea) sp. SF3</i>	1
<i>Leitoscoloplos pugettensis</i>	6	<i>Axinopsida serricata</i>	1
<i>Magelona sacculata</i>	6	<i>Caesia rhinetes</i>	1
<i>Podarkeopsis glabrus</i>	6	Cardiidae	1
<i>Tresus</i> spp.	6	<i>Cylichna attonsa</i>	1
<i>Carinoma mutabilis</i>	5	<i>Dendraster excentricus</i>	1
Enteropneusta	5	<i>Eohaustorius</i> spp.	1
<i>Aricidea (Aricidea) sp. SF2</i>	4	<i>Euchone hancocki</i>	1
<i>Astyris gausapata</i>	4	<i>Gadila aberrans</i>	1
<i>Kurtziella plumbea</i>	4	<i>Glycera macrobranchia</i>	1
Lineidae	4	Holothuroidea	1
<i>Photis brevipes</i>	4	<i>Kurtziella coani</i>	1
<i>Phyllodoce hartmanae</i>	4	<i>Lepidasthenia longicirrata</i>	1
<i>Prionospio lighti</i>	4	<i>Lumbrineris californiensis</i>	1
Anthozoa	3	<i>Macoma acolasta</i>	1
Lumbrineridae	3	<i>Malmgreniella</i> spp.	1
Paraonidae	3	<i>Mandibulophoxus gilesi</i>	1
<i>Phyllodoce longipes</i>	3	Mangeliidae	1
<i>Pista wui</i>	3	<i>Mediomastus acutus</i>	1
<i>Rhepoxynius</i> spp.	3	<i>Odostomia</i> spp.	1
<i>Spiophanes berkeleyorum</i>	3	<i>Owenia collaris</i>	1
<i>Amaeana occidentalis</i>	2	<i>Pacifoculodes barnardi</i>	1
<i>Aphelochaeta petersenae</i>	2	<i>Pandora bilirata</i>	1
<i>Aricidea</i> spp.	2	<i>Paranemertes californica</i>	1
<i>Bathycopea daltonae</i>	2	<i>Pinnixa franciscana</i>	1
<i>Cheirimeдея zotea</i>	2	<i>Scolecopsis</i> sp. SF2	1
<i>Diastylis santamariensis</i>	2	<i>Scolecopsis</i> spp.	1
<i>Diastylopsis</i> spp.	2	<i>Scoloplos</i> sp. SF1	1
<i>Goniada maculata</i>	2	<i>Spiochaetopterus costarum</i>	1
<i>Lanassa venusta</i>	2	<i>Sthenelais verruculosa</i>	1
<i>Mesolamprops dillonensis</i>	2	<i>Streblosoma</i> sp. SF1	1
<i>Micrura</i> spp. (?)	2	Terebellidae	1
<i>Neotrypaea</i> spp.	2	Tubulanidae sp. B	1
Phyllodocidae	2	<i>Turbonilla</i> spp.	1
		<i>Yoldia cooperii</i>	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<b>STATION 51</b>		Lineidae	2
<i>Callianax pycna</i>	772	<i>Rhepoxynius menziesi</i>	2
<i>Spiophanes norrisi</i>	29	<i>Scoloplos armiger</i>	2
<i>Pacifoculodes barnardi</i>	25	<i>Americhelidium shoemakeri</i>	1
<i>Mandibulophoxus gilesi</i>	17	<i>Apoprionospio pygmaea</i>	1
<i>Diastylopsis dawsoni</i>	14	<i>Caesia rhinetes</i>	1
<i>Rhepoxynius menziesi</i>	8	<i>Eumida longicornuta</i>	1
<i>Eohaustorius</i> spp.	6	<i>Leitoscoloplos pugettensis</i>	1
<i>Glycera macrobranchia</i>	4	<i>Onuphis</i> sp. A	1
<i>Photis macinerneyi</i>	4	<i>Photis macinerneyi</i>	1
Nemertea	3	<i>Rhepoxynius lucubrans</i>	1
<i>Onuphis</i> sp. A	3	<i>Rhepoxynius</i> spp.	1
<i>Photis</i> spp.	3	<i>Rhepoxynius vigitegus</i>	1
<i>Americhelidium shoemakeri</i>	2	<i>Scoletoma luti</i>	1
<i>Kurtiella tumida</i>	2	<i>Sigalion spinosus</i>	1
<i>Rhepoxynius vigitegus</i>	2	<i>Spiophanes</i> spp.	1
<i>Scoloplos armiger</i>	2	<b>STATION 53</b>	
<i>Tellina modesta</i>	2	<i>Spiophanes norrisi</i>	1672
<i>Caesia rhinetes</i>	1	<i>Protomedeia penates</i>	273
<i>Carinoma mutabilis</i>	1	<i>Photis</i> spp.	73
<i>Clinocardium nuttallii</i>	1	<i>Onuphis</i> sp. A	65
<i>Foxiphalus obtusidens</i>	1	<i>Rhepoxynius lucubrans</i>	55
<i>Leitoscoloplos pugettensis</i>	1	<i>Glycinde</i> spp.	46
<i>Nephtys caecoides</i>	1	<i>Scoletoma luti</i>	37
<i>Nephtys</i> spp.	1	<i>Amphiodia</i> spp.	31
<i>Phoronis</i> spp.	1	<i>Onuphis</i> spp.	31
<i>Protomedeia penates</i>	1	<i>Euphilomedes carcharodonta</i>	29
<i>Scoletoma luti</i>	1	<i>Callianax pycna</i>	28
<b>STATION 52</b>		Dendrochirotida	27
<i>Spiophanes norrisi</i>	111	<i>Tellina modesta</i>	27
<i>Callianax pycna</i>	68	<i>Glycinde picta</i>	26
<i>Chaetozone bansei</i>	12	<i>Nephtys caecoides</i>	25
<i>Carinoma mutabilis</i>	11	<i>Mesolamprops dillonensis</i>	22
<i>Glycera tenuis</i>	9	<i>Photis macinerneyi</i>	19
<i>Pacifoculodes barnardi</i>	9	<i>Rhepoxynius fatigans</i>	18
<i>Nephtys caecoides</i>	8	<i>Mediomastus</i> spp.	17
<i>Diastylopsis dawsoni</i>	4	<i>Pista wui</i>	16
<i>Eohaustorius</i> spp.	4	<i>Diastylopsis dawsoni</i>	15
<i>Tellina modesta</i>	3	<i>Lanassa venusta</i>	15
<i>Heteropodarke heteromorpha</i>	2	<i>Diastylis santamariensis</i>	12

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Carinoma mutabilis</i>	9	<i>Phyllodoce</i> spp.	2
<i>Photis parvidons</i>	9	<i>Phylo felix</i>	2
Enteropneusta	8	<i>Saccella</i> spp.	2
<i>Halcampa decententaculata</i>	8	<i>Scolelepis</i> sp. SF2	2
<i>Ampelisca cristata</i>	7	<i>Scoloplos</i> sp. SF1	2
<i>Kurtziella plumbea</i>	7	<i>Ampharete acutifrons</i>	1
<i>Pandora bilirata</i>	7	<i>Caesia rhinetes</i>	1
<i>Ampelisca careyi</i>	6	Cardiidae	1
<i>Apoprionospio pygmaea</i>	6	<i>Euchone hancocki</i>	1
<i>Astyris gausapata</i>	6	<i>Ischyrocerus pelagops</i>	1
<i>Micronephthys cornuta</i>	6	<i>Kurtiella tumida</i>	1
<i>Bathycopea daltonae</i>	5	<i>Macoma acolasta</i>	1
<i>Magelona hartmanae</i>	5	<i>Modiolus capax</i>	1
<i>Odostomia</i> spp.	5	Nassariidae	1
<i>Cylichna</i> spp.	4	Nemertea	1
Nematoda	4	<i>Nereis neoneanthes</i>	1
<i>Neotrypaea</i> spp.	4	<i>Owenia collaris</i>	1
<i>Pectinaria californiensis</i>	4	<i>Paranemertes californica</i>	1
<i>Aphelochaeta petersenae</i>	3	<i>Paraprionospio alata</i>	1
<i>Cheirimedeia zotea</i>	3	<i>Phyllodoce longipes</i>	1
<i>Cylichna attonsa</i>	3	Polynoidae	1
<i>Leukoma staminea</i>	3	<i>Sigalion spinosus</i>	1
Lineidae	3	<i>Solen sicarius</i>	1
<i>Mactromeris catilliformis</i>	3	<i>Spiophanes</i> spp.	1
<i>Magelona sacculata</i>	3	<i>Sthenelais verruculosa</i>	1
<i>Pacifoculodes barnardi</i>	3	<i>Xenolebris californica</i>	1
<i>Pleurogonium</i> sp. SF1	3	<b>STATION 54</b>	
<i>Stylatula</i> spp.	3	<i>Spiophanes norrisi</i>	381
<i>Tenonia priops</i>	3	<i>Photis</i> spp.	100
Terebellidae	3	<i>Photis macinerneyi</i>	98
<i>Amphicteis scaphobranchiata</i>	2	<i>Ischyrocerus pelagops</i>	54
<i>Amphiodia digitata</i>	2	<i>Carinoma mutabilis</i>	45
Anthozoa	2	<i>Diastylopsis dawsoni</i>	34
<i>Argissa hamatipes</i>	2	<i>Pacifoculodes barnardi</i>	27
<i>Crangon nigromaculata</i>	2	<i>Scoletoma luti</i>	26
<i>Gastropteron pacificum</i>	2	<i>Glycinde picta</i>	13
<i>Leitoscoloplos pugettensis</i>	2	<i>Callianax pycna</i>	8
<i>Macoma nasuta</i>	2	<i>Tellina modesta</i>	5
Paraonidae	2	<i>Glycinde</i> spp.	4
<i>Phyllodoce hartmanae</i>	2	<i>Rhepoxynius</i> spp.	4

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Onuphis</i> sp. A	3	Lineidae	19
<i>Eobrolgus spinosus</i>	2	<i>Euphilomedes carcharodonta</i>	17
<i>Glycera macrobranchia</i>	2	<i>Onuphis</i> sp. A	17
Nematoda	2	<i>Pleurogonium</i> sp. SF1	17
<i>Pleurogonium</i> sp. SF1	2	<i>Rhepoxynius fatigans</i>	15
<i>Rhepoxynius abronius</i>	2	<i>Diastylis santamariensis</i>	14
<i>Rhepoxynius lucubrans</i>	2	<i>Pacifocolodes barnardi</i>	12
<i>Tecticeps convexus</i>	2	<i>Odostomia</i> spp.	11
<i>Americhelidium shoemakeri</i>	1	<i>Spiophanes</i> spp.	11
Amphipoda	1	<i>Tenonia priops</i>	11
<i>Armandia brevis</i>	1	<i>Eumida longicornuta</i>	9
<i>Cylichna</i> spp.	1	<i>Macoma nasuta</i>	8
<i>Eohaustorius</i> spp.	1	<i>Paradialychone eiffelturris</i>	8
<i>Eusyllis transecta</i>	1	<i>Apoprionospio pygmaea</i>	6
Isaeidae	1	<i>Ischyrocerus pelagops</i>	6
<i>Leitoscoloplos pugettensis</i>	1	<i>Scoloplos</i> sp. SF1	6
Nemertea	1	<i>Magelona sacculata</i>	5
<i>Nephtys caecoides</i>	1	<i>Argissa hamatipes</i>	4
<i>Tenonia priops</i>	1	<i>Cylichna attonsa</i>	4
<i>Tresus</i> spp.	1	<i>Sithenelais verruculosa</i>	4
<b>STATION 56</b>		Terebellidae	4
<i>Spiophanes norrisi</i>	6150	<i>Ampelisca careyi</i>	3
<i>Protomedeia penates</i>	290	<i>Aricidea (Acmira) catherinae</i>	3
<i>Owenia collaris</i>	249	<i>Bathycopea daltonae</i>	3
<i>Photis</i> spp.	199	<i>Caesia rhinetes</i>	3
<i>Scoletoma luti</i>	90	<i>Kurtziella tumida</i>	3
<i>Glycinde picta</i>	85	<i>Kurtziella plumbea</i>	3
<i>Callianax pycna</i>	72	<i>Leukoma staminea</i>	3
<i>Photis macinerneyi</i>	64	<i>Macoma</i> spp.	3
<i>Onuphis</i> spp.	60	<i>Phyllodoce hartmanae</i>	3
<i>Glycinde</i> spp.	59	<i>Podarkeopsis glabrus</i>	3
<i>Mediomastus</i> spp.	48	<i>Spiochaetopterus costarum</i>	3
<i>Tellina modesta</i>	42	<i>Stylatula</i> spp.	3
<i>Diastylopsis dawsoni</i>	41	<i>Tresus</i> spp.	3
<i>Pectinaria californiensis</i>	35	<i>Americhelidium shoemakeri</i>	2
<i>Micronephtys cornuta</i>	33	<i>Ampharete acutifrons</i>	2
<i>Nephtys caecoides</i>	27	<i>Cheirimeдея zotea</i>	2
<i>Pandora bilirata</i>	21	<i>Cooperella subdiaphana</i>	2
<i>Magelona hartmanae</i>	20	<i>Crangon nigromaculata</i>	2
<i>Amphiodia</i> spp.	19	<i>Goniada maculata</i>	2



Appendix E-2 (cont.)  
Benthic infauna collected in 2012

Maldanidae	2	<i>Tellina modesta</i>	147
<i>Mesolamprops dillonensis</i>	2	<i>Pectinaria californiensis</i>	135
<i>Neotrypaea</i> spp.	2	<i>Magelona sacculata</i>	77
<i>Nereis neoneanthes</i>	2	<i>Callianax pycna</i>	67
<i>Ophiidermella</i> spp.	2	<i>Apoprionospio pygmaea</i>	64
<i>Photis parvidons</i>	2	<i>Diastylopsis dawsoni</i>	56
<i>Spiophanes berkeleyorum</i>	2	<i>Magelona hartmanae</i>	37
<i>Synidotea consolidata</i>	2	<i>Macoma nasuta</i>	29
<i>Tritella pilimana</i>	2	<i>Leukoma staminea</i>	24
Tubulanidae sp. B	2	<i>Mactromeris catilliformis</i>	24
<i>Ampelisca milleri</i>	1	<i>Onuphis</i> sp. A	24
Amphipoda	1	<i>Modiolus capax</i>	23
<i>Aphelochaeta petersenae</i>	1	<i>Leitoscoloplos pugettensis</i>	22
<i>Axinopsida serricata</i>	1	<i>Protomedeia penates</i>	22
Bivalvia	1	<i>Micronephtys cornuta</i>	21
<i>Edwardsia juliae</i>	1	<i>Spiophanes berkeleyorum</i>	19
Enteropneusta	1	<i>Photis</i> spp.	15
<i>Eteone fauchaldi</i>	1	Onuphidae	14
Euclymeninae sp. SF1	1	Nemertea	13
<i>Gastropteron pacificum</i>	1	<i>Amphicteis scaphobranchiata</i>	12
<i>Glycera macrobranchia</i>	1	Lineidae	11
<i>Halcapa decemtentaculata</i>	1	<i>Pista wui</i>	11
Inarticulata	1	<i>Amphiodia</i> spp.	8
<i>Leitoscoloplos pugettensis</i>	1	<i>Kurtiella tumida</i>	8
Lumbrineridae	1	<i>Pacifoculodes barnardi</i>	8
<i>Magelona berkeleyi</i>	1	<i>Carinoma mutabilis</i>	7
<i>Magelona</i> spp.	1	<i>Ischyrocerus pelagops</i>	7
<i>Microphthalmus</i> spp. complex	1	<i>Nephtys caecoides</i>	7
Nemertea	1	<i>Photis macinerneyi</i>	7
Onuphidae	1	<i>Argissa hamatipes</i>	6
<i>Paraprionospio alata</i>	1	<i>Glycinde picta</i>	6
<i>Pista</i> spp.	1	<i>Rhepoxynius fatigans</i>	6
<i>Pista wui</i>	1	<i>Streblosoma</i> spp.	6
<i>Siliqua lucida</i>	1	Dendrochirotida	5
<i>Yoldia cooperii</i>	1	<i>Glossaluax reclusiana</i>	5
<b>STATION 58</b>		Nematoda	5
<i>Owenia collaris</i>	301	<i>Pandora bilirata</i>	5
<i>Spiophanes norrisi</i>	174	<i>Cylichna</i> spp.	4
<i>Scoletoma luti</i>	153	<i>Diastylis santamariensis</i>	4
<i>Mediomastus</i> spp.	151	<i>Sthenelais verruculosa</i>	4

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Tubulanus pellucidus</i>	4	<i>Pholoe glabra</i>	1
<i>Aphelochaeta petersenae</i>	3	<i>Pinnixa franciscana</i>	1
<i>Astyris gausapata</i>	3	<i>Pleurogonium</i> sp. SF1	1
<i>Caesia rhinetes</i>	3	<i>Poecilochaetus johnsoni</i>	1
Enteropneusta	3	<i>Prionospio lighti</i>	1
<i>Kurtziella plumbea</i>	3	Sabellidae	1
<i>Paradialychone eiffelturris</i>	3	<i>Stylatula</i> spp.	1
Paraonidae	3	<i>Tenonia priops</i>	1
<i>Ampharete</i> spp.	2	<b>STATION 59</b>	
Anthozoa	2	<i>Spiophanes norrisi</i>	5417
Holothuroidea	2	<i>Photis</i> spp.	938
<i>Malmgreniella</i> spp.	2	<i>Photis macinerneyi</i>	559
<i>Odostomia</i> spp.	2	<i>Callianax pycna</i>	135
<i>Photis brevipes</i>	2	<i>Ischyrocerus pelagops</i>	130
<i>Podarkeopsis glabrus</i>	2	<i>Pleurogonium</i> sp. SF1	128
<i>Sigambra</i> sp. SF2	2	<i>Protomedeia penates</i>	127
<i>Siliqua lucida</i>	2	<i>Glycinde picta</i>	112
<i>Spiochaetopterus costarum</i>	2	<i>Scoletoma luti</i>	109
Terebellidae	2	<i>Glycinde</i> spp.	57
<i>Amaeana occidentalis</i>	1	<i>Pectinaria californiensis</i>	55
<i>Ampharete acutifrons</i>	1	<i>Owenia collaris</i>	45
<i>Aricidea (Aedicira) pacifica</i>	1	<i>Onuphis</i> spp.	37
<i>Balcis</i> spp.	1	<i>Macoma</i> spp.	33
<i>Cossura</i> spp.	1	<i>Diastylis santamariensis</i>	30
<i>Crangon</i> spp.	1	<i>Amphiodia</i> spp.	21
<i>Dendraster excentricus</i>	1	<i>Micronephtys cornuta</i>	20
<i>Dipolydora magna</i>	1	<i>Onuphis</i> sp. A	19
<i>Edotia sublittoralis</i>	1	<i>Diastylopsis dawsoni</i>	17
<i>Edwardsia juliae</i>	1	<i>Nephtys caecoides</i>	16
<i>Eteone fauchaldi</i>	1	<i>Tellina modesta</i>	16
<i>Euphilomedes carcharodonta</i>	1	<i>Clinocardium nuttallii</i>	13
<i>Glycera americana</i>	1	<i>Eumida longicornuta</i>	13
<i>Glycera macrobranchia</i>	1	<i>Macoma nasuta</i>	13
<i>Glycinde</i> spp.	1	<i>Magelona hartmanae</i>	13
<i>Macoma acolasta</i>	1	<i>Glycinde</i> spp.	9
Nassariidae	1	<i>Pacifoculodes barnardi</i>	9
<i>Neotrypaea</i> spp.	1	<i>Tenonia priops</i>	9
<i>Nereis neoneanthes</i>	1	<i>Apoprionospio pygmaea</i>	8
<i>Paracaudina chilensis</i>	1	<i>Leitoscoloplos pugettensis</i>	8
<i>Pentamera rigida</i>	1	<i>Odostomia</i> spp.	8

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Podarkeopsis glabrus</i>	8	Caprellidae	1
<i>Rhepoxynius fatigans</i>	8	<i>Cooperella subdiaphana</i>	1
<i>Mediomastus</i> spp.	7	<i>Dendraster excentricus</i>	1
Cardiidae	6	<i>Dendrochirotida</i>	1
<i>Gastropteron pacificum</i>	6	<i>Edwardsia juliae</i>	1
Lumbrineridae	6	<i>Eteone fauchaldi</i>	1
<i>Modiolus capax</i>	5	<i>Glycera macrobranchia</i>	1
<i>Pandora bilirata</i>	5	<i>Glycinde</i> sp. SF1	1
<i>Tritella pilimana</i>	5	<i>Halcampa decententaculata</i>	1
<i>Crangon nigromaculata</i>	4	Holothuroidea	1
<i>Phyllodoce hartmanae</i>	4	<i>Kurtziella tumida</i>	1
<i>Scolelepis</i> sp. SF1	4	<i>Lumbrineris californiensis</i>	1
<i>Aricidea (Aricidea)</i> sp. SF3	3	<i>Magelona</i> spp.	1
<i>Ischyrocerus anguipes</i>	3	<i>Mesochaetopterus</i> sp. SF1	1
<i>Kurtziella plumbea</i>	3	<i>Modiolus rectus</i>	1
<i>Paradialychone eiffelturris</i>	3	<i>Monocorophium acherusicum</i>	1
<i>Paranemertes californica</i>	3	Nematoda	1
<i>Prionospio lighti</i>	3	<i>Neomysis</i> spp.	1
<i>Spiophanes berkeleyorum</i>	3	Onuphidae	1
<i>Americhelidium shoemakeri</i>	2	<i>Opisthopus transversus</i>	1
<i>Ampharete acutifrons</i>	2	<i>Phyllodoce longipes</i>	1
<i>Aricidea (Aedicira) pacifica</i>	2	<i>Pinnixa franciscana</i>	1
<i>Bathycopea daltonae</i>	2	<i>Rhepoxynius</i> spp.	1
<i>Carinoma mutabilis</i>	2	<i>Scoloplos</i> sp. SF1	1
<i>Cylichna</i> spp.	2	<i>Siliqua lucida</i>	1
<i>Diaphana californica</i>	2	<i>Sinocorophium heteroceratum</i>	1
<i>Euphilomedes carcharodonta</i>	2	<i>Solen sicarius</i>	1
<i>Glycera americana</i>	2	<i>Synidotea consolidata</i>	1
<i>Leukoma staminea</i>	2	Terebellidae	1
Lineidae	2	<i>Tiron biocellata</i>	1
<i>Magelona sacculata</i>	2	<i>Tresus</i> spp.	1
<i>Microphthalmus</i> spp. complex	2	Tubulanidae sp. B	1
Nemertea	2	<i>Tubulanus pellucidus</i>	1
<i>Neotrypaea</i> spp.	2	<i>Typosyllis farallonensis</i>	1
<i>Photis brevipes</i>	2	<i>Yoldia cooperii</i>	1
<i>Photis parvidons</i>	2	<b>STATION 60</b>	
<i>Sthenelais verruculosa</i>	2	<i>Spiophanes norrisi</i>	4566
Anthozoa	1	<i>Protomedeia penates</i>	289
<i>Aphelochaeta</i> spp.	1	<i>Photis</i> spp.	247
<i>Astyris gausapata</i>	1	<i>Rhepoxynius fatigans</i>	100

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Photis macinerneyi</i>	87	<i>Caesia rhinetes</i>	2
<i>Glycinde picta</i>	70	<i>Cerebratulus californiensis</i>	2
<i>Callianax pycna</i>	67	<i>Cheirimeдея zotea</i>	2
<i>Scoletoma luti</i>	67	<i>Clinocardium nuttallii</i>	2
<i>Glycinde</i> spp.	36	Enteropneusta	2
<i>Euphilomedes carcharodonta</i>	29	<i>Eteone ?californica</i>	2
<i>Mediomastus</i> spp.	24	<i>Gastropteron pacificum</i>	2
<i>Tellina modesta</i>	24	<i>Kurtiella tumida</i>	2
<i>Pleurogonium</i> sp. SF1	21	<i>Kurtziella plumbea</i>	2
<i>Micronephtys cornuta</i>	19	<i>Leukoma staminea</i>	2
<i>Carinoma mutabilis</i>	15	Onuphidae	2
<i>Onuphis</i> spp.	15	<i>Owenia collaris</i>	2
<i>Ampelisca cristata</i>	14	<i>Paraprionospio alata</i>	2
<i>Amphiodia</i> spp.	13	<i>Pectinaria californiensis</i>	2
<i>Tenonia priops</i>	13	<i>Phylo felix</i>	2
<i>Onuphis</i> sp. A	11	Polynoidae	2
<i>Diastylis santamariensis</i>	10	<i>Sthenelais verruculosa</i>	2
<i>Pacifocolodes barnardi</i>	10	<i>Ampelisca</i> spp.	1
<i>Kurtzina beta</i>	9	<i>Argissa hamatipes</i>	1
<i>Eumida longicornuta</i>	8	Cephalaspidea	1
<i>Nephtys caecoides</i>	8	<i>Cylichna</i> spp.	1
Nematoda	7	<i>Diaphana californica</i>	1
Terebellidae	7	Gammaridea	1
<i>Crangon nigromaculata</i>	6	<i>Goniada maculata</i>	1
<i>Diastylopsis dawsoni</i>	6	<i>Leptochelia dubia</i>	1
<i>Odostomia</i> spp.	6	<i>Leptosynapta</i> spp.	1
<i>Ampelisca careyi</i>	5	Lineidae	1
<i>Dyopedos arcticus</i>	5	<i>Lumbrineris californiensis</i>	1
<i>Ischyrocerus pelagops</i>	5	<i>Macoma nasuta</i>	1
<i>Kurtiella coani</i>	5	<i>Macoma</i> spp.	1
<i>Glycinde</i> sp. SF1	4	Mactridae	1
<i>Leitoscoloplos pugettensis</i>	4	Maldanidae	1
<i>Magelona hartmanae</i>	4	<i>Neotrypaea</i> spp.	1
Nemertea	4	Opheliidae	1
<i>Stylatula</i> spp.	4	<i>Pandora bilirata</i>	1
<i>Amphiodia digitata</i>	3	<i>Paranemertes californica</i>	1
<i>Podarkeopsis glabrus</i>	3	<i>Photis brevipes</i>	1
<i>Rhepoxynius lucubrans</i>	3	<i>Phyllodoce hartmanae</i>	1
<i>Americhelidium shoemakeri</i>	2	<i>Prionospio lighti</i>	1
<i>Astyris gausapata</i>	2	<i>Scolelepis squamata</i>	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Scoloplos armiger</i>	1	<i>Paranemertes californica</i>	4
<i>Sigalion spinosus</i>	1	<i>Podarkeopsis glabrus</i>	4
<i>Spiochaetopterus costarum</i>	1	Terebellidae	4
<i>Tritella pilimana</i>	1	<i>Aricidea (Acmira) catherinae</i>	3
<b>STATION 61</b>		<i>Astyris gausapata</i>	3
<i>Spiophanes norrisi</i>	5881	<i>Cylichna</i> spp.	3
<i>Photis</i> spp.	754	<i>Pinnixa franciscana</i>	3
<i>Photis macinerneyi</i>	364	<i>Tenonia priops</i>	3
<i>Protomedeia penates</i>	241	<i>Ampelisca careyi</i>	2
<i>Scoletoma luti</i>	159	<i>Argissa hamatipes</i>	2
<i>Callianax pycna</i>	75	<i>Dyopodos arcticus</i>	2
<i>Glycinde picta</i>	64	<i>Goniada maculata</i>	2
<i>Glycinde</i> spp.	52	<i>Leukoma staminea</i>	2
<i>Mediomastus</i> spp.	28	Lumbrineridae	2
<i>Eumida longicornuta</i>	27	<i>Nereis neoneanthes</i>	2
<i>Ischyrocerus pelagops</i>	27	<i>Pandora bilirata</i>	2
<i>Pleurogonium</i> sp. SF1	21	<i>Paradialychone eiffelturris</i>	2
<i>Micronephtys cornuta</i>	17	<i>Spiochaetopterus costarum</i>	2
<i>Onuphis</i> sp. A	17	<i>Spiophanes berkeleyorum</i>	2
<i>Owenia collaris</i>	16	<i>Tritella pilimana</i>	2
<i>Rhepoxynius fatigans</i>	15	<i>Typosyllis farallonensis</i>	2
<i>Magelona hartmanae</i>	13	<i>Americhelidium shoemakeri</i>	1
<i>Amphiodia</i> spp.	12	<i>Ampharete</i> spp.	1
<i>Nephtys caecoides</i>	12	Amphipoda	1
<i>Diastylis santamariensis</i>	11	<i>Aphelochaeta petersenae</i>	1
<i>Euphilomedes carcharodonta</i>	11	<i>Aphelochaeta</i> spp.	1
<i>Diastylopsis dawsoni</i>	10	<i>Apoprionospio pygmaea</i>	1
Lineidae	9	<i>Armandia brevis</i>	1
<i>Tellina modesta</i>	9	Dendrochirotida	1
<i>Onuphis</i> spp.	8	<i>Edotia sublittoralis</i>	1
<i>Photis parvidons</i>	8	Euclymeninae sp. SF1	1
Nemertea	7	<i>Kurtziella plumbea</i>	1
<i>Pacifoculodes barnardi</i>	7	<i>Lissocrangon stylirostris</i>	1
<i>Scoloplos</i> sp. SF1	7	<i>Mesochaetopterus</i> sp. SF1	1
<i>Crangon nigromaculata</i>	6	<i>Microphthalmus</i> spp. complex	1
<i>Glycinde</i> sp. SF1	6	Nematoda	1
<i>Leitoscoloplos pugettensis</i>	5	<i>Neotrypaea</i> spp.	1
<i>Macoma</i> spp.	5	<i>Odostomia</i> spp.	1
<i>Pectinaria californiensis</i>	5	Ostracoda sp. SF2	1
<i>Glycera macrobranchia</i>	4	Paraonidae	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Siliqua lucida</i>	1	<i>Magelona hartmanae</i>	4
<i>Tubulanus pellucidus</i>	1	<i>Pleurogonium</i> sp. SF1	4
<b>STATION 62</b>		<i>Stylatula</i> spp.	4
<i>Spiophanes norrisi</i>	2672	<i>Yoldia cooperii</i>	4
<i>Protomedeia penates</i>	300	<i>Ampharete labrops</i>	3
<i>Callianax pycna</i>	156	<i>Apoprionospio pygmaea</i>	3
<i>Rhepoxynius fatigans</i>	119	<i>Bathycopea daltonae</i>	3
<i>Glycinde</i> spp.	90	<i>Caesia rhinetes</i>	3
<i>Mediomastus</i> spp.	90	<i>Clinocardium nuttallii</i>	3
Nematoda	89	<i>Cylichna</i> spp.	3
<i>Photis</i> spp.	69	<i>Eumida longicornuta</i>	3
<i>Glycinde picta</i>	58	<i>Pagurus</i> spp.	3
<i>Euphilomedes carcharodonta</i>	50	<i>Paranemertes californica</i>	3
<i>Rhepoxynius lucubrans</i>	47	<i>Phyllodoce</i> spp.	3
<i>Scoletoma luti</i>	37	<i>Pista wui</i>	3
<i>Kurtiella tumida</i>	34	Polynoidae	3
<i>Tellina modesta</i>	25	<i>Scolelepis</i> sp. SF2	3
<i>Ampelisca careyi</i>	20	<i>Amphiodia digitata</i>	2
<i>Mesolamprops dillonensis</i>	18	<i>Aricidea (Acmira) catherinae</i>	2
<i>Onuphis</i> sp. A	16	<i>Glycera macrobranchia</i>	2
<i>Nephtys caecoides</i>	14	<i>Haliophasma geminatum</i>	2
<i>Photis macinerneyi</i>	13	<i>Lanassa venusta</i>	2
<i>Micronephtys cornuta</i>	12	<i>Leitoscoloplos pugettensis</i>	2
<i>Amphiodia</i> spp.	11	<i>Neotrypaea</i> spp.	2
<i>Diastylopsis dawsoni</i>	10	<i>Owenia collaris</i>	2
<i>Halcampa decemtentaculata</i>	10	<i>Photis parvidons</i>	2
<i>Pacifoculodes barnardi</i>	10	<i>Phyllodoce hartmanae</i>	2
<i>Ampelisca cristata</i>	9	<i>Rhepoxynius</i> spp.	2
<i>Carinoma mutabilis</i>	8	<i>Spiochaetopterus costarum</i>	2
<i>Astyris gausapata</i>	7	<i>Anonyx adoxus</i>	1
Lineidae	7	<i>Argissa hamatipes</i>	1
Nemertea	7	<i>Axinopsida serricata</i>	1
<i>Diastylis santamariensis</i>	6	<i>Chaetozone columbiana</i>	1
<i>Onuphis</i> spp.	6	<i>Crangon nigromaculata</i>	1
<i>Tenonia priops</i>	6	<i>Dendraster excentricus</i>	1
Dendrochirotida	5	<i>Epitonium</i> spp.	1
<i>Macoma nasuta</i>	5	<i>Kurtziella plumbea</i>	1
<i>Pectinaria californiensis</i>	5	<i>Macoma acolasta</i>	1
<i>Podarkeopsis glabrus</i>	5	<i>Magelona sacculata</i>	1
<i>Spiophanes berkeleyorum</i>	5	<i>Magelona</i> spp.	1

Appendix E-2 (cont.)  
*Benthic infauna collected in 2012*

Maldanidae	1	<i>Photis brevipes</i>	8
<i>Neomysis kadiakensis</i>	1	<i>Bathycopea daltonae</i>	7
Nudibranchia	1	<i>Magelona hartmanae</i>	7
<i>Odostomia</i> spp.	1	<i>Pectinaria californiensis</i>	6
Paraonidae	1	<i>Diastylis santamariensis</i>	5
<i>Paraprionospio alata</i>	1	<i>Micronephtys cornuta</i>	5
<i>Scolelepis</i> spp.	1	<i>Americhelidium shoemakeri</i>	4
<i>Scoloplos</i> sp. SF1	1	<i>Clinocardium nuttallii</i>	4
<i>Sigalion spinosus</i>	1	<i>Heptacarpus</i> spp.	4
<i>Sthenelais verruculosa</i>	1	<i>Ampharete acutifrons</i>	3
Terebellidae	1	Nematoda	3
Tubulanidae sp. B	1	Nemertea	3
<b>STATION 63</b>		<i>Pandora bilirata</i>	3
<i>Spiophanes norrisi</i>	2829	<i>Phyllodoce hartmanae</i>	3
<i>Callianax pycna</i>	146	<i>Tenonia priops</i>	3
<i>Photis</i> spp.	145	<i>Tritella pilimana</i>	3
<i>Ischyrocerus pelagops</i>	125	<i>Astyris gausapata</i>	2
<i>Scoletoma luti</i>	93	<i>Axinopsida serricata</i>	2
<i>Rhepoxynius fatigans</i>	78	<i>Glycera macrobranchia</i>	2
<i>Photis macinerneyi</i>	77	<i>Glycinde</i> sp. SF1	2
<i>Onuphis</i> sp. A	59	<i>Halcampa decententaculata</i>	2
<i>Protomedeia penates</i>	52	<i>Leukoma staminea</i>	2
<i>Glycinde picta</i>	51	<i>Magelona sacculata</i>	2
<i>Glycinde</i> spp.	48	<i>Mesolamprops dillonensis</i>	2
<i>Onuphis</i> spp.	31	<i>Paranemertes californica</i>	2
<i>Eumida longicornuta</i>	28	<i>Scoloplos</i> sp. SF1	2
<i>Pacifoculodes barnardi</i>	28	<i>Sthenelais verruculosa</i>	2
<i>Pleurogonium</i> sp. SF1	20	Terebellidae	2
<i>Tellina modesta</i>	20	Amphipoda	1
<i>Euphilomedes carcharodonta</i>	16	<i>Argissa hamatipes</i>	1
<i>Amphiodia</i> spp.	11	<i>Aricidea (Aricidea)</i> sp. SF3	1
<i>Nephtys caecoides</i>	11	Bivalvia	1
<i>Carinoma mutabilis</i>	10	<i>Caesia rhinetes</i>	1
<i>Ampelisca careyi</i>	9	<i>Crangon nigromaculata</i>	1
Lineidae	9	<i>Cylichna</i> spp.	1
<i>Mediomastus</i> spp.	9	Dendrochirotida	1
<i>Diastylopsis dawsoni</i>	8	Enteropneusta	1
<i>Kurtziella plumbea</i>	8	<i>Gastropteron pacificum</i>	1
<i>Macoma</i> spp.	8	<i>Hemilamprops californicus</i>	1
<i>Odostomia</i> spp.	8	Holothuroidea	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Kurtiella tumida</i>	1	<i>Spiophanes berkeleyorum</i>	11
<i>Leitoscoloplos pugettensis</i>	1	<i>Pandora bilirata</i>	10
<i>Magelona berkeleyi</i>	1	<i>Cylichna</i> spp.	8
<i>Metacarcinus gracilis</i>	1	<i>Nephtys caecoides</i>	8
<i>Microphthalmus</i> spp. complex	1	<i>Rhepoxynius fatigans</i>	8
Nassariidae	1	Lumbrineridae	7
Onuphidae	1	<i>Diastylis santamariensis</i>	5
<i>Owenia collaris</i>	1	<i>Pectinaria californiensis</i>	5
<i>Phyllodoce</i> spp.	1	<i>Astyris gausapata</i>	4
<i>Podarkeopsis glabrus</i>	1	<i>Kurtzina beta</i>	4
Polynoidae	1	<i>Micronephtys cornuta</i>	4
<i>Sigalion spinosus</i>	1	<i>Ampharete acutifrons</i>	3
<i>Stylatula</i> spp.	1	<i>Leukoma staminea</i>	3
<i>Tubulanus</i> spp.	1	Nemertea	3
<i>Yoldia cooperii</i>	1	<i>Odostomia</i> spp.	3
<b>STATION 64</b>		Onuphidae	3
<i>Spiophanes norrisi</i>	548	<i>Scoloplos armiger</i>	3
<i>Photis</i> spp.	353	<i>Tenonia priops</i>	3
<i>Photis macinerneyi</i>	179	<i>Americhelidium shoemakeri</i>	2
<i>Diastylopsis dawsoni</i>	114	<i>Argissa hamatipes</i>	2
<i>Callianax pycna</i>	111	<i>Glycera macrobranchia</i>	2
<i>Scoletoma luti</i>	72	<i>Leitoscoloplos pugettensis</i>	2
<i>Tellina modesta</i>	61	<i>Magelona berkeleyi</i>	2
<i>Apoprionospio pygmaea</i>	45	<i>Melanochlamys diomedea</i>	2
<i>Macoma nasuta</i>	45	Nassariidae	2
<i>Magelona sacculata</i>	43	Nematoda	2
<i>Onuphis</i> spp.	35	<i>Podarkeopsis glabrus</i>	2
<i>Pacifoculodes barnardi</i>	31	<i>Spiochaetopterus costarum</i>	2
<i>Magelona hartmanae</i>	25	Ampharetidae	1
<i>Onuphis</i> sp. A	24	<i>Aphelochaeta petersenae</i>	1
<i>Pleurogonium</i> sp. SF1	23	<i>Bathycopea daltonae</i>	1
<i>Owenia collaris</i>	22	<i>Clinocardium nuttallii</i>	1
<i>Ischyrocerus anguipes</i>	18	<i>Eteone</i> spp.	1
<i>Mediomastus</i> spp.	14	<i>Eumida longicornuta</i>	1
<i>Glycinde picta</i>	13	<i>Euphilomedes carcharodonta</i>	1
<i>Protomedeia penates</i>	13	<i>Kurtziella plumbea</i>	1
<i>Amphiodia</i> spp.	12	Lineidae	1
<i>Paradialychone eiffelturris</i>	12	<i>Lumbrineris californiensis</i>	1
<i>Carinoma mutabilis</i>	11	Maldanidae	1
<i>Glycinde</i> spp.	11	<i>Modiolus</i> spp.	1



Appendix E-2 (cont.)  
*Benthic infauna collected in 2012*

Nuculanidae	1	Onuphidae	9
Orbiniidae	1	<i>Owenia collaris</i>	9
Paraonidae	1	<i>Nephtys caecoides</i>	6
<i>Photis parvidons</i>	1	<i>Kurtziella plumbea</i>	5
<i>Phylo felix</i>	1	Maldanidae	5
<i>Polinices draconis</i>	1	<i>Micronephtys cornuta</i>	5
<i>Prionospio lighti</i>	1	Ampharetidae	4
Sabellidae	1	<i>Leitoscoloplos pugettensis</i>	4
<i>Tritella pilimana</i>	1	<i>Scoloplos</i> sp. SF1	4
<b>STATION 65</b>		<i>Americhelidium shoemakeri</i>	3
<i>Spiophanes norrisi</i>	646	<i>Ampharete acutifrons</i>	3
<i>Scoletoma luti</i>	149	<i>Aphelochaeta petersenae</i>	3
Lumbrineridae	111	<i>Armandia brevis</i>	3
<i>Photis</i> spp.	99	<i>Cylichna attonsa</i>	3
<i>Apoprionospio pygmaea</i>	84	<i>Mesolamprops dillonensis</i>	3
<i>Magelona sacculata</i>	76	<i>Argissa hamatipes</i>	2
<i>Pectinaria californiensis</i>	61	<i>Carinoma mutabilis</i>	2
<i>Photis macinerneyi</i>	52	<i>Eumida longicornuta</i>	2
<i>Mediomastus</i> spp.	49	<i>Gastropteron pacificum</i>	2
<i>Diastylopsis dawsoni</i>	39	<i>Glycinde</i> sp. SF1	2
<i>Glycinde</i> spp.	36	<i>Magelona</i> spp.	2
<i>Tellina modesta</i>	28	<i>Rhepoxynius fatigans</i>	2
<i>Macoma</i> spp.	23	<i>Spiochaetopterus costarum</i>	2
<i>Callianax pycna</i>	21	<i>Spiophanes</i> spp.	2
<i>Pandora bilirata</i>	19	<i>Tenonia priops</i>	2
<i>Onuphis</i> spp.	18	Terebellidae	2
<i>Pacifoculodes barnardi</i>	18	<i>Yoldia cooperii</i>	2
<i>Pleurogonium</i> sp. SF1	18	<i>Ampharete</i> spp.	1
Lineidae	16	<i>Amphicteis scaphobranchiata</i>	1
<i>Magelona hartmanae</i>	15	<i>Aricidea (Acmira) catherinae</i>	1
<i>Protomedeia penates</i>	15	<i>Astyris gausapata</i>	1
<i>Glycera macrobranchia</i>	14	<i>Caesia rhinetes</i>	1
<i>Ischyrocerus pelagops</i>	12	<i>Cheirimeidia zotea</i>	1
Nematoda	12	<i>Crangon nigromaculata</i>	1
<i>Onuphis</i> sp. A	12	<i>Cylichna</i> spp.	1
<i>Paradialychone eiffelturris</i>	11	<i>Euclymeninae</i> sp. SF1	1
<i>Spiophanes berkeleyorum</i>	11	<i>Euphilomedes carcharodonta</i>	1
<i>Amphiodia</i> spp.	10	<i>Galathowenia oculata</i>	1
<i>Macoma nasuta</i>	10	<i>Glycera americana</i>	1
<i>Glycinde picta</i>	9	<i>Hemilamprops californicus</i>	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Kurtiella tumida</i>	1	Nemertea	5
<i>Leukoma staminea</i>	1	<i>Odostomia</i> spp.	5
<i>Microphthalmus</i> spp. complex	1	<i>Paranemertes californica</i>	5
Nassariidae	1	<i>Pleurogonium</i> sp. SF1	5
<i>Pista wui</i>	1	<i>Tenonia priops</i>	5
<i>Solen sicarius</i>	1	<i>Aricidea (Acmira) catherinae</i>	4
<i>Spiophanes duplex</i>	1	<i>Axinopsida serricata</i>	4
<b>STATION 66</b>		<i>Bathycopea daltonae</i>	4
<i>Spiophanes norrisi</i>	1035	<i>Carinoma mutabilis</i>	4
<i>Protomedeia penates</i>	115	<i>Paraprionospio alata</i>	4
<i>Rhepoxynius fatigans</i>	112	<i>Phyllodoce hartmanae</i>	4
<i>Mediomastus</i> spp.	74	<i>Pista wui</i>	4
<i>Glycinde</i> spp.	45	<i>Cheirimedeia zotea</i>	3
<i>Photis</i> spp.	36	<i>Glycinde</i> sp. SF1	3
<i>Scoletoma luti</i>	32	<i>Halcampa decententaculata</i>	3
<i>Kurtiella tumida</i>	31	Lineidae	3
Nematoda	29	<i>Onuphis</i> spp.	3
<i>Ampelisca careyi</i>	22	<i>Pagurus</i> spp.	3
<i>Magelona hartmanae</i>	21	<i>Amphicteis scaphobranchiata</i>	2
<i>Ampelisca cristata</i>	20	<i>Caesia rhinetes</i>	2
<i>Micronephtys cornuta</i>	19	<i>Caprella californica</i>	2
<i>Onuphis</i> sp. A	19	<i>Crangon nigromaculata</i>	2
<i>Glycinde picta</i>	17	<i>Eumida longicornuta</i>	2
<i>Diastylopsis dawsoni</i>	16	<i>Glycera macrobranchia</i>	2
<i>Tellina modesta</i>	15	<i>Ischyrocerus pelagops</i>	2
<i>Spiophanes berkeleyorum</i>	14	<i>Kurtziella plumbea</i>	2
<i>Magelona sacculata</i>	13	<i>Lanassa venusta</i>	2
<i>Mesolamprops dillonensis</i>	13	<i>Paracaudina chilensis</i>	2
<i>Pacifoculodes barnardi</i>	11	<i>Saccella</i> spp.	2
<i>Apoprionospio pygmaea</i>	9	<i>Scoloplos armiger</i>	2
<i>Euphilomedes carcharodonta</i>	9	<i>Spiochaetopterus costarum</i>	2
<i>Nephtys caecoides</i>	8	<i>Turbonilla</i> spp.	2
<i>Rhepoxynius lucubrans</i>	8	<i>Ampharete acutifrons</i>	1
<i>Spiophanes</i> spp.	8	<i>Amphiodia digitata</i>	1
<i>Argissa hamatipes</i>	7	Anthozoa	1
<i>Astyris gausapata</i>	7	<i>Aoroides inermis</i>	1
<i>Pectinaria californiensis</i>	7	<i>Aphelocheata petersenae</i>	1
<i>Callianax pycna</i>	6	Caprellidae	1
<i>Photis brevipes</i>	6	<i>Chaetozone columbiana</i>	1
<i>Amphiodia</i> spp.	5	<i>Compsomyax subdiaphana</i>	1

Appendix E-2 (cont.)  
*Benthic infauna collected in 2012*

<i>Cylichna</i> spp.	1	<i>Amphiodia</i> spp.	9
<i>Dendrochirotida</i>	1	<i>Carinoma mutabilis</i>	8
<i>Diastylis santamariensis</i>	1	<i>Gastropterion pacificum</i>	6
<i>Dipolydora magna</i>	1	<i>Pleurogonium</i> sp. SF1	6
<i>Dipolydora</i> spp.	1	<i>Micronephtys cornuta</i>	5
Enteropneusta	1	<i>Diastylis santamariensis</i>	4
Euclymeninae sp. SF1	1	<i>Eumida longicornuta</i>	4
<i>Glossaluax reclusiana</i>	1	<i>Kurtziella plumbea</i>	4
<i>Glycera</i> spp.	1	<i>Leitoscoloplos pugettensis</i>	4
<i>Goniada maculata</i>	1	<i>Leukoma staminea</i>	4
<i>Kurtiella coani</i>	1	<i>Macoma</i> spp.	4
<i>Leitoscoloplos pugettensis</i>	1	<i>Pandora bilirata</i>	4
<i>Macoma nasuta</i>	1	<i>Tenonia priops</i>	4
<i>Modiolus capax</i>	1	<i>Americhelidium shoemakeri</i>	3
<i>Photis macinerneyi</i>	1	<i>Apoprionospio pygmaea</i>	3
<i>Podarkeopsis glabrus</i>	1	<i>Bathycopea daltonae</i>	3
Polynoidae	1	Lineidae	3
<i>Sthenelais verruculosa</i>	1	<i>Magelona sacculata</i>	3
<i>Tiron biocellata</i>	1	Nematoda	3
<i>Tubulanus pellucidus</i>	1	Terebellidae	3
<b>STATION 67</b>		<i>Tritella pilimana</i>	3
<i>Spiophanes norrisi</i>	1574	<i>Ampelisca cristata</i>	2
<i>Photis</i> spp.	341	<i>Amphiodia digitata</i>	2
<i>Photis macinerneyi</i>	124	<i>Axinopsida serricata</i>	2
<i>Rhepoxynius fatigans</i>	115	<i>Cylichna</i> spp.	2
<i>Scoletoma luti</i>	75	<i>Glycera macrobranchia</i>	2
<i>Protomedeia penates</i>	66	<i>Magelona hartmanae</i>	2
<i>Onuphis</i> sp. A	40	<i>Odostomia</i> spp.	2
<i>Onuphis</i> spp.	32	Onuphidae	2
<i>Callianax pycna</i>	31	<i>Pectinaria californiensis</i>	2
<i>Glycinde</i> spp.	27	<i>Phyllodoce longipes</i>	2
<i>Ampelisca careyi</i>	22	Polynoidae	2
<i>Tellina modesta</i>	18	<i>Sthenelais verruculosa</i>	2
<i>Euphilomedes carcharodonta</i>	17	<i>Ampharete acutifrons</i>	1
<i>Pacifoculodes barnardi</i>	15	<i>Argissa hamatipes</i>	1
<i>Diastylopsis dawsoni</i>	12	<i>Astyris gausapata</i>	1
<i>Nephtys caecoides</i>	12	<i>Cerebratulus californiensis</i>	1
<i>Glycinde picta</i>	11	<i>Crangon nigromaculata</i>	1
<i>Hemilamprops californicus</i>	10	<i>Diaphana californica</i>	1
<i>Photis parvidons</i>	10	Enteropneusta	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Glycinde</i> sp. SF1	1	<i>Apoprionospio pygmaea</i>	7
<i>Kurtiella tumida</i>	1	<i>Bathycopea daltonae</i>	7
Maldanidae	1	<i>Stylatula</i> spp.	7
<i>Mediomastus</i> spp.	1	<i>Edwardsia juliae</i>	6
<i>Microphthalmus</i> spp. complex	1	<i>Scoletoma luti</i>	6
<i>Modiolus capax</i>	1	<i>Tenonia priops</i>	6
Nassariidae	1	<i>Aphelochaeta petersenae</i>	5
<i>Ostracoda</i> sp. SF2	1	<i>Carinoma mutabilis</i>	5
<i>Paranemertes californica</i>	1	<i>Crangon nigromaculata</i>	5
<i>Phyllodoce hartmanae</i>	1	Nassariidae	5
<i>Spiochaetopterus costarum</i>	1	<i>Nephtys caecoides</i>	5
<i>Spiophanes berkeleyorum</i>	1	<i>Odostomia</i> spp.	5
<i>Stylatula</i> spp.	1	<i>Pectinaria californiensis</i>	4
<b>STATION 68</b>		<i>Ampelisca agassizi</i>	3
<i>Spiophanes norrisi</i>	578	<i>Caesia rhinetes</i>	3
<i>Rhepoxynius fatigans</i>	134	<i>Euphilomedes carcharodonta</i>	3
<i>Cheirimeideia zotea</i>	121	Isaeidae	3
<i>Protomeideia penates</i>	67	Lineidae	3
<i>Glycinde</i> spp.	54	<i>Turbonilla</i> spp.	3
<i>Mediomastus</i> spp.	48	<i>Amphiteis scaphobranchiata</i>	2
Nematoda	45	Anthozoa	2
<i>Callianax pycna</i>	36	<i>Aricidea (Aricidea)</i> sp. SF2	2
<i>Ampelisca cristata</i>	35	<i>Cylichna</i> spp.	2
<i>Kurtiella tumida</i>	34	Dendrochirotida	2
<i>Ampelisca careyi</i>	25	<i>Eteone</i> sp. SF3	2
<i>Onuphis</i> sp. A	23	Euclymeninae sp. SF1	2
<i>Tellina modesta</i>	19	<i>Glossaluax reclusiana</i>	2
<i>Glycinde picta</i>	17	<i>Lamprops tomalesi</i>	2
<i>Magelona sacculata</i>	17	<i>Mesochaetopterus</i> sp. SF1	2
<i>Photis</i> spp.	14	<i>Micrura</i> spp. (?)	2
<i>Astyris gausapata</i>	12	Nemertea	2
<i>Amphiodia</i> spp.	10	<i>Neotrypaea</i> spp.	2
<i>Hemilamprops californicus</i>	10	<i>Pachynus barnardi</i>	2
<i>Micronephtys cornuta</i>	10	<i>Phyllodoce longipes</i>	2
<i>Magelona hartmanae</i>	9	<i>Podarkeopsis glabrus</i>	2
<i>Ostracoda</i> sp. SF2	9	<i>Scolelepis</i> sp. SF2	2
<i>Pacificolodes barnardi</i>	9	<i>Tritella pilimana</i>	2
<i>Pista wui</i>	9	Ampharetidae	1
<i>Diastylopsis dawsoni</i>	8	<i>Amphiodia digitata</i>	1
<i>Americhelidium shoemakeri</i>	7	<i>Anonyx adoxus</i>	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Aricidea (Acmira) catherinae</i>	1	<i>Ampelisca careyi</i>	21
<i>Armandia brevis</i>	1	<i>Amphiodia</i> spp.	20
<i>Axinopsida serricata</i>	1	<i>Callianax pycna</i>	18
Cardiidae	1	<i>Onuphis</i> spp.	15
Chaetodermatida	1	<i>Photis</i> spp.	15
<i>Diastylis santamariensis</i>	1	<i>Lanassa venusta</i>	10
<i>Dyopedos arcticus</i>	1	<i>Pista wui</i>	10
Enteropneusta	1	<i>Edwardsia juliae</i>	8
<i>Gastropteran pacificum</i>	1	Lineidae	8
<i>Goniada maculata</i>	1	<i>Micronephtys cornuta</i>	8
<i>Goniada</i> spp.	1	<i>Nephtys caecoides</i>	8
<i>Halcampa decententaculata</i>	1	<i>Tritella pilimana</i>	8
<i>Lumbrineris californiensis</i>	1	<i>Apoprionospio pygmaea</i>	7
Mangeliidae	1	<i>Diastylopsis dawsoni</i>	6
<i>Neomysis</i> spp.	1	<i>Macoma</i> spp.	6
<i>Nereis neoneanthes</i>	1	<i>Magelona hartmanae</i>	6
<i>Photis macinerneyi</i>	1	Maldanidae	6
<i>Pleurogonium</i> sp. SF1	1	<i>Eumida longicornuta</i>	5
<i>Spiochaetopterus costarum</i>	1	<i>Odostomia</i> spp.	5
<i>Streblosoma</i> spp.	1	<i>Pleurogonium</i> sp. SF1	5
<i>Tubulanus pellucidus</i>	1	<i>Tiron biocellata</i>	5
<i>Typosyllis farallonensis</i>	1	<i>Diastylis santamariensis</i>	4
<b>STATION 69</b>		<i>Leukoma staminea</i>	4
<i>Spiophanes norrisi</i>	2079	<i>Magelona sacculata</i>	4
<i>Protomedeia penates</i>	175	Polynoidae	4
<i>Rhepoxynius fatigans</i>	107	<i>Amphicteis scaphobranchiata</i>	3
<i>Glycinde</i> spp.	80	<i>Bathycopea daltonae</i>	3
<i>Scoletoma luti</i>	80	<i>Gastropteran pacificum</i>	3
<i>Cheirimedeia zotea</i>	62	<i>Kurtziella plumbea</i>	3
<i>Mediomastus</i> spp.	60	<i>Pectinaria californiensis</i>	3
<i>Glycinde picta</i>	59	<i>Phyllodoce hartmanae</i>	3
<i>Astyris gausapata</i>	53	<i>Aphelochaeta petersenae</i>	2
Terebellidae	38	<i>Cylichna attonsa</i>	2
<i>Tellina modesta</i>	33	<i>Euphilomedes carcharodonta</i>	2
<i>Onuphis</i> sp. A	32	<i>Glycera macrobranchia</i>	2
<i>Ampelisca cristata</i>	31	<i>Lumbrineris californiensis</i>	2
<i>Hemilamprops californicus</i>	28	<i>Metacarcinus gracilis</i>	2
Nematoda	27	<i>Pacifoculodes barnardi</i>	2
Dendrochirotida	24	<i>Photis parvidons</i>	2
<i>Kurtziella tumida</i>	24	<i>Streblosoma</i> spp.	2

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Tenonia priops</i>	2	<i>Onuphis</i> spp.	14
<i>Amaeana occidentalis</i>	1	<i>Photis</i> spp.	14
<i>Ampelisca agassizi</i>	1	<i>Mediomastus</i> spp.	13
<i>Ampharete acutifrons</i>	1	Lineidae	12
<i>Ampharete</i> spp.	1	<i>Pectinaria californiensis</i>	12
<i>Amphiodia digitata</i>	1	Onuphidae	11
<i>Argissa hamatipes</i>	1	<i>Pleurogonium</i> sp. SF1	11
<i>Aricidea (Aedicira)</i> sp. A	1	<i>Macoma</i> spp.	10
<i>Armandia brevis</i>	1	<i>Photis macinerneyi</i>	10
<i>Caesia rhinetes</i>	1	<i>Pacifoculodes barnardi</i>	8
<i>Crangon nigromaculata</i>	1	<i>Scoloplos</i> sp. SF1	8
<i>Cylichna</i> spp.	1	<i>Leukoma staminea</i>	7
<i>Eusyllis transecta</i>	1	<i>Glycera macrobranchia</i>	6
Gastropoda	1	Nemertea	5
<i>Glycera americana</i>	1	<i>Nephtys caecoides</i>	5
<i>Glycinde</i> sp. SF1	1	<i>Ampharete acutifrons</i>	4
<i>Goniada maculata</i>	1	Ampharetidae	4
<i>Halcompa decemtentaculata</i>	1	<i>Carinoma mutabilis</i>	4
<i>Lepidasthenia</i> spp.	1	<i>Magelona hartmanae</i>	4
<i>Lissocrangon stylirostris</i>	1	<i>Pandora bilirata</i>	4
<i>Modiolus</i> spp.	1	<i>Americhelidium shoemakeri</i>	3
Onuphidae	1	<i>Amphiodia</i> spp.	3
<i>Pandora bilirata</i>	1	<i>Aricidea (Aricidea)</i> sp. SF2	2
<i>Paranemertes californica</i>	1	<i>Bathycopea daltonae</i>	2
<i>Pherusa neopapillata</i>	1	<i>Glycinde</i> spp.	2
<i>Podarkeopsis glabrus</i>	1	<i>Leitoscoloplos pugettensis</i>	2
<i>Prionospio lighti</i>	1	<i>Mesolamprops dillonensis</i>	2
<i>Scoloplos</i> sp. SF1	1	Nematoda	2
<i>Spiophanes duplex</i>	1	<i>Ampharete labrops</i>	1
<i>Turbonilla</i> spp.	1	<i>Callianax pycna</i>	1
<i>Typosyllis farallonensis</i>	1	<i>Glycera</i> spp.	1
<i>Yoldia cooperii</i>	1	<i>Glycinde</i> sp. SF1	1
<b>STATION 70</b>		<i>Macoma nasuta</i>	1
<i>Spiophanes norrisi</i>	436	<i>Magelona</i> spp.	1
<i>Scoletoma luti</i>	111	Mangeliidae	1
<i>Apoprionospio pygmaea</i>	38	<i>Mediomastus acutus</i>	1
<i>Mactromeris catilliformis</i>	21	<i>Neotrypaea</i> spp.	1
<i>Magelona sacculata</i>	20	<i>Odostomia</i> spp.	1
<i>Tellina modesta</i>	20	<i>Onuphis</i> sp. A	1
<i>Diastylopsis dawsoni</i>	14	Oweniidae	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Paradialychone eiffelturris</i>	1	<i>Eumida longicornuta</i>	5
<i>Phoronis</i> spp.	1	<i>Glycinde picta</i>	5
<i>Scolecopsis</i> sp. SF2	1	<i>Mesolamprops dillonensis</i>	5
<i>Spiochaetopterus costarum</i>	1	<i>Carinoma mutabilis</i>	4
<i>Sthenelais verruculosa</i>	1	<i>Cylichna</i> spp.	4
<i>Synidotea consolidata</i>	1	<i>Diastylis santamariensis</i>	4
<i>Tecticeps convexus</i>	1	<i>Kurtziella plumbea</i>	4
<b>STATION 71</b>		Lumbrineridae	4
<i>Spiophanes norrisi</i>	519	<i>Macoma nasuta</i>	4
<i>Tellina modesta</i>	115	<i>Protomedeia penates</i>	4
<i>Photis</i> spp.	116	<i>Gastropteron pacificum</i>	3
<i>Scoletoma luti</i>	65	Nassariidae	3
<i>Photis macinerneyi</i>	59	Onuphidae	3
<i>Apoprionospio pygmaea</i>	41	<i>Owenia collaris</i>	3
<i>Pacifoculodes barnardi</i>	36	<i>Rhepoxynius fatigans</i>	3
<i>Onuphis</i> spp.	35	<i>Spiophanes berkeleyorum</i>	3
<i>Diastylopsis dawsoni</i>	31	<i>Ampelisca careyi</i>	2
<i>Macoma</i> spp.	29	<i>Amphiodia</i> spp.	2
<i>Hemilamprops californicus</i>	23	<i>Aricidea (Aricidea)</i> sp. SF3	2
<i>Magelona sacculata</i>	20	<i>Caesia rhinetes</i>	2
<i>Callianax pycna</i>	18	Enteropneusta	2
<i>Pectinaria californiensis</i>	18	<i>Glycera americana</i>	2
<i>Mactromeris catilliformis</i>	17	<i>Glycinde</i> sp. SF1	2
<i>Magelona hartmanae</i>	16	<i>Leitoscoloplos pugettensis</i>	2
<i>Micronephtys cornuta</i>	16	Maldanidae	2
<i>Leukoma staminea</i>	14	<i>Odostomia</i> spp.	2
<i>Pleurogonium</i> sp. SF1	14	<i>Paradialychone eiffelturris</i>	2
<i>Ischyrocerus pelagops</i>	13	<i>Siliqua lucida</i>	2
<i>Astyris gausapata</i>	12	<i>Tritella pilimana</i>	2
<i>Mediomastus</i> spp.	12	<i>Americhelidium shoemakeri</i>	1
Lineidae	10	<i>Ampelisca cristata</i>	1
<i>Pandora bilirata</i>	10	<i>Ampharete</i> spp.	1
<i>Glycinde</i> spp.	9	<i>Aricidea (Aedicira)</i> sp. A	1
Nematoda	9	<i>Cheirimedeia zotea</i>	1
<i>Scoloplos</i> sp. SF1	8	Dendrochirotida	1
<i>Glycera macrobranchia</i>	7	<i>Diaphana californica</i>	1
<i>Nephtys caecoides</i>	7	<i>Eteone</i> spp.	1
<i>Onuphis</i> sp. A	7	Euclymeninae sp. SF1	1
<i>Aphelochaeta petersenae</i>	5	<i>Glossaluax reclusiana</i>	1
<i>Clinocardium nuttallii</i>	5	<i>Lumbrineris californiensis</i>	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Magelona</i> spp.	1	<i>Spiophanes berkeleyorum</i>	7
<i>Modiolus capax</i>	1	<i>Stylatula</i> spp.	7
<i>Modiolus rectus</i>	1	<i>Caesia rhinetes</i>	6
Nemertea	1	Lineidae	6
<i>Nereis neoneanthes</i>	1	<i>Pandora bilirata</i>	6
<i>Phylo felix</i>	1	<i>Photis</i> spp.	6
<i>Pinnixa franciscana</i>	1	<i>Dendraster excentricus</i>	5
<i>Scolelepis</i> sp. SF2	1	Anthozoa	4
<i>Scolelepis</i> spp.	1	<i>Aricidea (Aricidea)</i> sp. SF3	4
<i>Streblosoma</i> sp. SF1	1	<i>Edwardsia juliae</i>	4
Terebellidae	1	<i>Glycinde picta</i>	4
<i>Tresus</i> spp.	1	<i>Lanassa venusta</i>	4
<i>Yoldia cooperii</i>	1	<i>Lepidasthenia longicirrata</i>	4
<b>STATION 72</b>		<i>Leukoma staminea</i>	4
<i>Scoletoma luti</i>	205	<i>Macoma nasuta</i>	4
<i>Mactromeris catilliformis</i>	125	Nassariidae	4
<i>Mediomastus</i> spp.	103	<i>Nephtys caecoides</i>	4
<i>Spiophanes norrisi</i>	95	Paraonidae	4
<i>Protomedeia penates</i>	94	<i>Streblosoma</i> spp.	4
<i>Tellina modesta</i>	78	<i>Tubulanus pellucidus</i>	4
<i>Callianax pycna</i>	66	<i>Yoldia cooperii</i>	4
<i>Onuphis</i> sp. A	53	<i>Amaeana occidentalis</i>	3
<i>Diastylopsis dawsoni</i>	44	<i>Ampelisca cristata</i>	3
<i>Pectinaria californiensis</i>	32	Ampharetidae	3
<i>Rhepoxynius fatigans</i>	30	Hydrozoa	3
<i>Glycinde</i> spp.	29	Nematoda	3
<i>Kurtiella tumida</i>	25	<i>Phylo felix</i>	3
Enteropneusta	22	<i>Solen sicarius</i>	3
Nemertea	21	<i>Streblosoma</i> sp. SF1	3
<i>Magelona hartmanae</i>	19	<i>Tenonia priops</i>	3
<i>Magelona sacculata</i>	18	<i>Ampelisca careyi</i>	2
<i>Owenia collaris</i>	17	<i>Cylichna attonsa</i>	2
<i>Aphelochaeta petersenae</i>	16	<i>Glossaluax reclusiana</i>	2
<i>Leitoscoloplos pugettensis</i>	16	<i>Ischyrocerus pelagops</i>	2
<i>Apoprionospio pygmaea</i>	14	<i>Kellia</i> sp. SF1	2
<i>Odostomia</i> spp.	10	<i>Kurtziella plumbea</i>	2
<i>Amphiodia</i> spp.	9	<i>Modiolus capax</i>	2
<i>Glycera macrobranchia</i>	8	<i>Neotrypaea</i> spp.	2
Dendrochirotida	7	Onuphidae	2
<i>Pleurogonium</i> sp. SF1	7	<i>Paraprionospio alata</i>	2



Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Pholoe glabra</i>	2	<i>Scoletoma luti</i>	111
<i>Sigambra</i> sp. SF2	2	<i>Tresus</i> spp.	77
Terebellidae	2	<i>Phyllodoce williamsi</i>	54
Veneridae	2	<i>Tellina modesta</i>	47
<i>Ampharete acutifrons</i>	1	<i>Foxiphalus obtusidens</i>	42
<i>Ampharete</i> spp.	1	<i>Carinoma mutabilis</i>	40
<i>Aoroides inermis</i>	1	<i>Owenia collaris</i>	39
<i>Aricidea (Acmira) horikoshii</i>	1	<i>Clinocardium nuttallii</i>	35
<i>Aricidea (Aedicira) pacifica</i>	1	<i>Pacifoculodes barnardi</i>	35
<i>Aricidea (Aedicira)</i> sp. A	1	<i>Tiron biocellata</i>	33
<i>Astyris gausapata</i>	1	<i>Gnathopleustes pugettensis</i>	20
<i>Axinopsida serricata</i>	1	<i>Pleurogonium</i> sp. SF1	18
Cardiidae	1	<i>Synidotea consolidata</i>	17
<i>Carinoma mutabilis</i>	1	<i>Paranemertes californica</i>	15
<i>Diastylis santamariensis</i>	1	<i>Diastylopsis dawsoni</i>	10
<i>Eteone fauchaldi</i>	1	<i>Mactromeris catilliformis</i>	10
<i>Goniada maculata</i>	1	<i>Odostomia</i> spp.	10
Isaeidae	1	<i>Callianax pycna</i>	9
<i>Lepidasthenia</i> spp.	1	<i>Cylichna</i> spp.	9
<i>Magelona berkeleyi</i>	1	<i>Phyllodoce williamsi</i>	9
Maldanidae	1	<i>Podarkeopsis glabrus</i>	9
<i>Malmgreniella</i> spp.	1	<i>Eumida longicornuta</i>	8
Mangeliidae	1	<i>Kurtiella coani</i>	7
<i>Pacifoculodes barnardi</i>	1	<i>Protomedeia penates</i>	7
<i>Photis macinerneyi</i>	1	<i>Amphiodia</i> spp.	6
<i>Photis parvidons</i>	1	<i>Glycinde</i> sp. SF1	6
<i>Pista wui</i>	1	<i>Leukoma staminea</i>	6
<i>Podarkeopsis glabrus</i>	1	<i>Kurtiella tumida</i>	5
<i>Prionospio lighti</i>	1	<i>Tritella pilimana</i>	5
<i>Rhepoxynius lucubrans</i>	1	<i>Harmothoe imbricata</i> complex	4
<i>Scoloplos</i> sp. SF1	1	<i>Bathycopea daltonae</i>	3
<i>Sthenelais verruculosa</i>	1	<i>Dyopodos arcticus</i>	3
<b>STATION 73</b>		<i>Edotia sublittoralis</i>	3
<i>Spiophanes norrisi</i>	10861	Flabelligeridae	3
<i>Photis</i> spp.	326	<i>Mediomastus acutus</i>	3
<i>Glycinde picta</i>	307	<i>Mediomastus</i> spp.	3
<i>Ischyrocerus pelagops</i>	218	<i>Nephtys caecoides</i>	3
<i>Glycinde</i> spp.	206	<i>Pectinaria californiensis</i>	3
<i>Photis macinerneyi</i>	176	<i>Rhepoxynius abronius</i>	3
<i>Megamoera subtener</i>	136	<i>Americhelidium shoemakeri</i>	2

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Amphiodia digitata</i>	2	<i>Tellina modesta</i>	42
<i>Aphelochaeta petersenae</i>	2	<i>Pacifoculodes barnardi</i>	40
Bivalvia	2	<i>Rhepoxynius abronius</i>	25
<i>Dendraster excentricus</i>	2	<i>Diastylopsis dawsoni</i>	24
Gastropoda	2	<i>Glycinde</i> spp.	23
<i>Gastropteron pacificum</i>	2	<i>Onuphis</i> sp. A	16
<i>Glycera macrobranchia</i>	2	<i>Paranemertes californica</i>	14
<i>Goniada maculata</i>	2	<i>Clinocardium nuttallii</i>	11
<i>Heptacarpus stimpsoni</i>	2	<i>Synidotea consolidata</i>	11
<i>Mesochaetopterus</i> sp. SF1	2	<i>Rhepoxynius lucubrans</i>	10
Nemertea	2	<i>Aoroides inermis</i>	9
<i>Pherusa neopapillata</i>	2	<i>Eobrolgus spinosus</i>	9
<i>Scolecopsis squamata</i>	2	<i>Glycera macrobranchia</i>	9
<i>Streptosyllis</i> sp. SF1	2	<i>Pleurogonium</i> sp. SF1	9
<i>Amaeana occidentalis</i>	1	<i>Kurtiella coani</i>	8
<i>Ampelisca</i> spp.	1	<i>Lissocrangon stylirostris</i>	7
<i>Diastylis santamariensis</i>	1	<i>Odostomia</i> spp.	6
<i>Glossaluax reclusiana</i>	1	<i>Paradialychone eiffelturris</i>	6
Hesionidae	1	<i>Phyllodoce williamsi</i>	6
<i>Leitoscoloplos pugettensis</i>	1	<i>Onuphis</i> spp.	5
<i>Lumbrineris californiensis</i>	1	<i>Americhelidium shoemakeri</i>	4
<i>Magelona sacculata</i>	1	<i>Nephtys caecoides</i>	4
<i>Notomastus lineatus</i>	1	<i>Apoprionospio pygmaea</i>	3
<i>Philine auriformis</i>	1	<i>Bathycopea daltonae</i>	3
<i>Photis parvidons</i>	1	<i>Diastylis santamariensis</i>	3
<i>Phyllodoce hartmanae</i>	1	<i>Eumida longicornuta</i>	3
Polynoidae	1	<i>Phyllodoce hartmanae</i>	3
<i>Scoloplos</i> sp. SF1	1	<i>Siliqua lucida</i>	3
<i>Sthenelais verruculosa</i>	1	<i>Tecticeps convexus</i>	3
<i>Tenonia priops</i>	1	<i>Tresus</i> spp.	3
Turbellaria	1	<i>Tritella pilimana</i>	3
<b>STATION 75</b>		<i>Armandia brevis</i>	2
<i>Spiophanes norrisi</i>	3452	<i>Eteone (Mysta)</i> sp. SF1	2
<i>Photis</i> spp.	836	<i>Gnathopleustes pugettensis</i>	2
<i>Photis macinerneyi</i>	382	Lineidae	2
<i>Ischyrocerus pelagops</i>	168	Nassariidae	2
<i>Glycinde picta</i>	115	<i>Onuphis</i> spp.	2
<i>Scoletoma luti</i>	86	<i>Owenia collaris</i>	2
<i>Carinoma mutabilis</i>	72	<i>Pandora bilirata</i>	2
<i>Callianax pycna</i>	49	<i>Protomedeia penates</i>	2

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Ampharete labrops</i>	1	<i>Podarkeopsis glabrus</i>	20
<i>Amphiodia</i> spp.	1	<i>Protomedeia penates</i>	20
<i>Crangon nigromaculata</i>	1	<i>Rhepoxynius fatigans</i>	20
<i>Cylichna</i> spp.	1	<i>Pectinaria californiensis</i>	17
<i>Edotia sublittoralis</i>	1	<i>Nephtys caecoides</i>	16
<i>Eohaustorius</i> spp.	1	<i>Diastylopsis dawsoni</i>	15
Holothuroidea	1	<i>Magelona sacculata</i>	14
<i>Kurtiella tumida</i>	1	<i>Clinocardium nuttallii</i>	10
<i>Leitoscoloplos pugettensis</i>	1	<i>Glycera macrobranchia</i>	10
<i>Leukoma staminea</i>	1	Lineidae	10
<i>Macoma</i> spp.	1	<i>Micronephtys cornuta</i>	10
<i>Modiolus rectus</i>	1	<i>Owenia collaris</i>	10
<i>Munnogonium tillerae</i>	1	<i>Paradialychone eiffelturris</i>	10
<i>Photis brevipes</i>	1	<i>Tellina modesta</i>	9
<i>Podarkeopsis glabrus</i>	1	<i>Apoprionospio pygmaea</i>	8
<i>Rhepoxynius vigitegus</i>	1	<i>Cylichna</i> spp.	8
<i>Spiochaetopterus costarum</i>	1	Nemertea	8
<i>Sthenelais verruculosa</i>	1	<i>Odostomia</i> spp.	7
<i>Tenonia priops</i>	1	<i>Enteropneusta</i>	6
Turbellaria	1	<i>Macoma acolasta</i>	6
<b>STATION 77</b>		<i>Prionospio lighti</i>	6
<i>Spiophanes norrisi</i>	5033	<i>Siliqua lucida</i>	6
<i>Photis</i> spp.	233	<i>Amphiodia</i> spp.	5
<i>Ischyrocerus pelagops</i>	212	<i>Caesia rhinetes</i>	5
<i>Photis macinerneyi</i>	183	Nematoda	5
<i>Leukoma staminea</i>	155	<i>Diastylis santamariensis</i>	4
<i>Scoletoma luti</i>	119	<i>Gastropteron pacificum</i>	4
<i>Mediomastus</i> spp.	110	<i>Magelona hartmanae</i>	4
<i>Pleurogonium</i> sp. SF1	108	<i>Sthenelais verruculosa</i>	4
<i>Mactromeris catilliformis</i>	85	<i>Tenonia priops</i>	4
<i>Glycinde picta</i>	82	<i>Dendraster excentricus</i>	3
<i>Kurtiella tumida</i>	65	<i>Euphilomedes carcharodonta</i>	3
<i>Glycinde</i> spp.	45	<i>Fabia subquadrata</i>	3
Bivalvia	34	<i>Rhepoxynius lucubrans</i>	3
<i>Callianax pycna</i>	33	<i>Tiron biocellata</i>	3
<i>Pacifoculodes barnardi</i>	33	<i>Americhelidium shoemakeri</i>	2
<i>Onuphis</i> sp. A	32	<i>Aphelochaeta petersenae</i>	2
<i>Macoma</i> spp.	31	<i>Argissa hamatipes</i>	2
<i>Macoma nasuta</i>	30	<i>Carinoma mutabilis</i>	2
<i>Onuphis</i> spp.	23	Dendrochirotida	2

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Eteone</i> sp. SF4	2	<i>Protomedeia penates</i>	141
<i>Eumida longicornuta</i>	2	<i>Scoletoma luti</i>	136
<i>Halcampa decemtentaculata</i>	2	<i>Ischyrocerus pelagops</i>	91
<i>Modiolus capax</i>	2	<i>Glycinde picta</i>	84
Nassariidae	2	<i>Glycinde</i> spp.	47
<i>Nereis neoneanthes</i>	2	<i>Carinoma mutabilis</i>	19
<i>Phyllodoce williamsi</i>	2	<i>Eumida longicornuta</i>	19
<i>Scoloplos</i> sp. SF1	2	<i>Tellina modesta</i>	16
<i>Sigalion spinosus</i>	2	<i>Amphiodia</i> spp.	14
<i>Amaeana occidentalis</i>	1	<i>Pleurogonium</i> sp. SF1	12
<i>Ampharete acutifrons</i>	1	<i>Diastylopsis dawsoni</i>	10
<i>Aricidea (Aricidea)</i> sp. SF3	1	<i>Onuphis</i> sp. A	10
<i>Bathycopea daltonae</i>	1	<i>Owenia collaris</i>	9
Cancridae	1	<i>Photis parvidons</i>	9
<i>Crangon nigromaculata</i>	1	<i>Micronephthys cornuta</i>	8
Crangonidae	1	<i>Diastylis santamariensis</i>	7
<i>Eusyllis transecta</i>	1	<i>Magelona hartmanae</i>	7
<i>Glycera americana</i>	1	<i>Crangon nigromaculata</i>	6
<i>Glycinde</i> sp. SF1	1	<i>Odostomia</i> spp.	6
<i>Kurtiella coani</i>	1	<i>Paranemertes californica</i>	6
<i>Leitoscoloplos pugettensis</i>	1	<i>Pectinaria californiensis</i>	6
<i>Mesolamprops dillonensis</i>	1	<i>Euphilomedes carcharodonta</i>	5
<i>Neotrypaea</i> spp.	1	<i>Leitoscoloplos pugettensis</i>	5
<i>Pandora bilirata</i>	1	<i>Macoma nasuta</i>	5
<i>Photis parvidons</i>	1	<i>Mediomastus</i> spp.	5
<i>Phyllodoce hartmanae</i>	1	<i>Rhepoxynius fatigans</i>	5
<i>Pinnixa franciscana</i>	1	<i>Scoloplos</i> sp. SF1	5
Polynoidae	1	<i>Eteone fauchaldi</i>	4
<i>Sigambra</i> sp. SF2	1	<i>Glycinde</i> sp. SF1	4
<i>Spiochaetopterus costarum</i>	1	<i>Nephtys caecoides</i>	4
<i>Spiophanes berkeleyorum</i>	1	<i>Pacificolodes barnardi</i>	4
<i>Synidotea consolidata</i>	1	<i>Tenonia priops</i>	4
<i>Tecticeps convexus</i>	1	<i>Magelona sacculata</i>	3
Terebellidae	1	<i>Paradialychone eiffelturris</i>	3
<i>Tubulanus pellucidus</i>	1	<i>Argissa hamatipes</i>	2
<b>STATION 78</b>		Callianassidae	2
<i>Spiophanes norrisi</i>	6565	<i>Gastropteron pacificum</i>	2
<i>Photis</i> spp.	813	<i>Amphiodia digitata</i>	1
<i>Photis macinerneyi</i>	497	<i>Apoprionospio pygmaea</i>	1
<i>Callianax pycna</i>	146	<i>Armandia brevis</i>	1

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

Bivalvia	1	<i>Tellina modesta</i>	69
Corophoidea	1	<i>Scoletoma luti</i>	64
<i>Cylichna</i> spp.	1	<i>Owenia collaris</i>	57
<i>Diaphana californica</i>	1	Nemertea	54
<i>Dyopedos arcticus</i>	1	<i>Spiophanes norrisi</i>	51
<i>Edwardsia juliae</i>	1	<i>Mediomastus</i> spp.	43
<i>Glycera macrobranchia</i>	1	<i>Onuphis</i> sp. A	35
<i>Goniada maculata</i>	1	<i>Photis</i> spp.	34
<i>Kurtiella tumida</i>	1	<i>Magelona sacculata</i>	33
<i>Kurtzina beta</i>	1	<i>Ischyrocerus pelagops</i>	32
<i>Leukoma staminea</i>	1	<i>Pectinaria californiensis</i>	27
<i>Lumbrineris californiensis</i>	1	<i>Nephtys caecoides</i>	26
<i>Mesochaetopterus</i> spp.	1	<i>Macoma nasuta</i>	25
Nematoda	1	<i>Glycinde picta</i>	22
Nemertea	1	<i>Kurtiella tumida</i>	20
<i>Nereis neoneanthes</i>	1	<i>Tecticeps convexus</i>	18
<i>Onuphis</i> spp.	1	<i>Yoldia cooperii</i>	16
Paraonidae	1	<i>Diastylopsis tenuis</i>	15
<i>Paraprionospio alata</i>	1	<i>Photis macinerneyi</i>	14
<i>Pentamera rigida</i>	1	<i>Leitoscoloplos pugettensis</i>	9
<i>Pherusa neopapillata</i>	1	<i>Amphiodia</i> spp.	8
<i>Phyllodoce hartmanae</i>	1	<i>Bathycopea daltonae</i>	8
<i>Phyllodoce longipes</i>	1	<i>Glycinde</i> spp.	8
<i>Pista</i> spp.	1	<i>Diastylis santamariensis</i>	7
<i>Podarkeopsis glabrus</i>	1	<i>Pleurogonium</i> sp. SF1	7
<i>Saccella taphria</i>	1	<i>Americhelidium shoemakeri</i>	5
<i>Scoloplos armiger</i>	1	<i>Magelona hartmanae</i>	5
<i>Spiophanes berkeleyorum</i>	1	Onuphidae	5
<i>Synidotea consolidata</i>	1	<i>Podarkeopsis glabrus</i>	4
Synopiidae	1	<i>Edotia sublittoralis</i>	3
<i>Tresus</i> spp.	1	<i>Modiolus capax</i>	3
Tubulanidae sp. B	1	<i>Tiron biocellata</i>	3
<i>Typosyllis farallonensis</i>	1	<i>Aoroides inermis</i>	2
<i>Yoldia cooperii</i>	1	Bivalvia	2
<b>STATION 79</b>		<i>Eteone ?californica</i>	2
<i>Spiophanes norrisi</i>	1135	<i>Glycinde</i> sp. SF1	2
<i>Callianax pycna</i>	129	<i>Leukoma staminea</i>	2
<i>Diastylopsis dawsoni</i>	124	<i>Micronephtys cornuta</i>	2
<i>Apoprionospio pygmaea</i>	100	<i>Nereis neoneanthes</i>	2
<i>Pacifoculodes barnardi</i>	80	<i>Odostomia</i> spp.	2

Appendix E-2 (cont.)  
Benthic infauna collected in 2012

<i>Pandora bilirata</i>	2	<i>Pacifocolodes barnardi</i>	5
<i>Paradialychone eiffelturris</i>	2	<i>Photis</i> spp.	5
<i>Ampelisca careyi</i>	1	<i>Diastylopsis dawsoni</i>	4
<i>Argissa hamatipes</i>	1	<i>Photis macinerneyi</i>	4
<i>Aricidea (Aricidea)</i> sp. SF1	1	<i>Mediomastus</i> spp.	3
<i>Aricidea (Aricidea)</i> sp. SF3	1	<i>Nephtys caecoides</i>	3
<i>Caesia rhinetes</i>	1	<i>Tellina modesta</i>	3
<i>Crangon nigromaculata</i>	1	<i>Americhelidium shoemakeri</i>	2
<i>Cylichna</i> spp.	1	<i>Apoprionospio pygmaea</i>	2
<i>Diastylopsis</i> spp.	1	Lineidae	2
<i>Eteone</i> spp.	1	<i>Amphiodia</i> spp.	1
<i>Glycera americana</i>	1	Anomura	1
<i>Glycera macrobranchia</i>	1	<i>Bathycopea daltonae</i>	1
<i>Halosydna brevisetosa</i>	1	<i>Caesia rhinetes</i>	1
Holothuroidea	1	<i>Euphilomedes carcharodonta</i>	1
<i>Mactromeris catilliformis</i>	1	<i>Glycinde</i> spp.	1
<i>Modiolus rectus</i>	1	<i>Halosydna brevisetosa</i>	1
<i>Neomysis</i> spp.	1	<i>Ischyrocerus anguipes</i>	1
<i>Neotrypaea</i> spp.	1	<i>Kurtiella coani</i>	1
<i>Pista</i> sp. SF1	1	<i>Lumbrineris californiensis</i>	1
<i>Protomedeia penates</i>	1	<i>Magelona sacculata</i>	1
<i>Scoloplos</i> sp. SF1	1	Nemertea	1
<i>Sigambra</i> sp. SF2	1	<i>Neomysis kadiakensis</i>	1
<i>Spiochaetopterus costarum</i>	1	<i>Nereis neoneanthes</i>	1
<i>Sthenelais verruculosa</i>	1	<i>Pagurus</i> spp.	1
<i>Tenonia priops</i>	1	<i>Scolecopsis</i> sp. SF2	1
<b>STATION 80</b>		<i>Sthenelais verruculosa</i>	1
<i>Eobrolgus spinosus</i>	63		
<i>Rhepoxynius lucubrans</i>	61		
<i>Callianax pycna</i>	44		
<i>Spiophanes norrisi</i>	44		
<i>Scoloplos armiger</i>	14		
<i>Aoroides inermis</i>	11		
<i>Chaetozone bansei</i>	11		
<i>Protomedeia penates</i>	7		
<i>Carinoma mutabilis</i>	6		
<i>Eohaustorius</i> spp.	6		
<i>Rhepoxynius fatigans</i>	6		
<i>Rhepoxynius vigitegus</i>	6		
<i>Scoletoma luti</i>	6		











Appendix E-4  
Benthic infauna collected in 2010 Stations 78 and 79

**STATION 78**

<i>Spiophanes norrisi</i>	1704	<i>Lumbrineris californiensis</i>	1
<i>Scoletoma luti</i>	65	<i>Mactromeris catilliformis</i>	1
<i>Photis</i> spp.	46	Maldanidae	1
<i>Callianax pycna</i>	36	Nassariidae	1
<i>Photis macinerneyi</i>	28	Nematoda	1
<i>Magelona sacculata</i>	21	<i>Odostomia</i> spp.	1
<i>Glycinde</i> spp.	20	Onuphidae	1
<i>Glycinde picta</i>	17	Ostracoda sp. SF2	1
<i>Ischyrocerus pelagops</i>	13	<i>Paranemertes californica</i>	1
<i>Diastylopsis dawsoni</i>	12	<i>Pectinaria californiensis</i>	1
<i>Onuphis</i> sp. A	12	Phyllodoceidae	1
<i>Mediomastus</i> spp.	11	<i>Prionospio lighti</i>	1
<i>Pacifoculodes barnardi</i>	9	<i>Rictaxis punctocaelatus</i>	1
<i>Amphiodia</i> spp.	8	<i>Scoloplos</i> sp. SF1	1
<i>Magelona hartmanae</i>	8	<i>Sthenelais verruculosa</i>	1
<i>Apoprionospio pygmaea</i>	7	Terebellidae	1
<i>Diastylis santamariensis</i>	7	<i>Travisia gigas</i>	1
<i>Tellina modesta</i>	7	Turridae	1
<i>Tenonia priops</i>	6		
<i>Argissa hamatipes</i>	5		
<i>Nephtys caecoides</i>	5		
<i>Euphilomedes carcharodonta</i>	4		
<i>Modiolus capax</i>	4		
<i>Phyllodoce hartmanae</i>	4		
<i>Protomedeia penates</i>	4		
<i>Saccella taphria</i>	4		
<i>Caesia rhinetes</i>	3		
<i>Clinocardium nuttallii</i>	3		
<i>Onuphis</i> spp.	3		
<i>Pleurogonium</i> sp. SF1	3		
<i>Rhepoxynius fatigans</i>	3		
<i>Eteone</i> ? <i>californica</i>	2		
<i>Glycinde</i> sp. SF1	2		
<i>Leukoma staminea</i>	2		
<i>Macoma nasuta</i>	2		
<i>Photis parvidons</i>	2		
<i>Poecilochaetus johnsoni</i>	2		
<i>Scoloplos armiger</i>	2		
<i>Ampelisca milleri</i>	1		
<i>Ampharete acutifrons</i>	1		
<i>Ampharete labrops</i>	1		
Amphipoda	1		
<i>Axinopsida serricata</i>	1		
Dendrochirotida	1		
<i>Glycera macrobranchia</i>	1		
<i>Kurtiella tumida</i>	1		
<i>Leitoscoloplos pugettensis</i>	1		

Appendix E-4 (cont.)  
*Benthic infauna collected in 2010 Stations 78 and 79*

<b>STATION 79</b>		Nassariidae	2
<i>Spiophanes norrisi</i>	5733	<i>Paranemertes californica</i>	2
<i>Photis</i> spp.	186	<i>Scoloplos</i> sp. SF1	2
<i>Scoletoma luti</i>	162	<i>Siliqua lucida</i>	2
<i>Mediomastus</i> spp.	121	<i>Stylatula</i> spp.	2
<i>Photis macinerneyi</i>	102	<i>Synidotea consolidata</i>	2
<i>Pacifoculodes barnardi</i>	35	<i>Tubulanus pellucidus</i>	2
<i>Glycinde picta</i>	34	<i>Amaeana occidentalis</i>	1
<i>Tellina modesta</i>	34	<i>Aricidea (Aedicira) pacifica</i>	1
<i>Onuphis</i> spp.	33	<i>Aricidea (Aedicira) sp. A</i>	1
<i>Paradialychone eiffelturris</i>	33	<i>Aricidea (Aricidea) sp. SF3</i>	1
<i>Apoprionospio pygmaea</i>	32	<i>Axinopsida serricata</i>	1
<i>Onuphis</i> sp. A	22	<i>Caesia rhinetes</i>	1
<i>Glycinde</i> spp.	20	<i>Cylichna attonsa</i>	1
<i>Callianax pycna</i>	19	<i>Dendrochirotida</i>	1
<i>Leukoma staminea</i>	18	<i>Diastylis santamariensis</i>	1
<i>Pleurogonium</i> sp. SF1	13	<i>Edotia sublittoralis</i>	1
<i>Kurtiella tumida</i>	12	Gastropoda	1
<i>Amphiodia</i> spp.	11	<i>Glycinde</i> sp. SF1	1
Lineidae	11	<i>Kurtziella plumbea</i>	1
<i>Leitoscoloplos pugettensis</i>	10	<i>Mediomastus acutus</i>	1
<i>Sthenelais verruculosa</i>	9	<i>Modiolus rectus</i>	1
<i>Cylichna</i> spp.	7	<i>Phoronis</i> spp.	1
<i>Macoma nasuta</i>	7	<i>Photis parvidons</i>	1
<i>Macoma</i> spp.	7	<i>Phyllodoce hartmanae</i>	1
<i>Eteone</i> sp. SF4	6	<i>Rhepoxynius abronius</i>	1
<i>Odostomia</i> spp.	6	Sabellidae	1
<i>Pandora bilirata</i>	6	<i>Spiophanes berkeleyorum</i>	1
<i>Pectinaria californiensis</i>	6	<i>Streptosyllis</i> sp. SF1	1
Bivalvia	5	<i>Tecticeps convexus</i>	1
<i>Diastylopsis dawsoni</i>	5	<i>Typosyllis farallonensis</i>	1
<i>Nephtys caecoides</i>	5		
<i>Tenonia priops</i>	5		
<i>Ampharete acutifrons</i>	4		
Cardiidae	4		
<i>Clinocardium nuttallii</i>	4		
<i>Mactromeris catilliformis</i>	4		
<i>Magelona hartmanae</i>	4		
Ampharetidae	3		
<i>Glycera macrobranchia</i>	3		
<i>Magelona sacculata</i>	3		
Opheliidae	3		
<i>Podarkeopsis glabrus</i>	3		
<i>Ampharete labrops</i>	2		
<i>Bathycopea daltonae</i>	2		
<i>Eumida longicornuta</i>	2		
<i>Ischyrocerus pelagops</i>	2		



**APPENDIX F**

**DEMERSAL FISH AND  
EPIBENTHIC INVERTEBRATES**

APPENDIX F  
DEMERSAL FISH AND EPIBENTHIC INVERTEBRATES,  
1997 to 2008

Appendix		Page
F-1	SWOO Regional Monitoring Program community trawls 1982-2008 by station and year with the number of times sampled per year and (the number of trawls per station)	F-2
F-2	SWOO Regional Monitoring Program community trawls 1997 - 1998 by station and year with the number of times sampled per year and (the number of trawls per station)	F-3

Appendix F-1  
*SWOO Regional Monitoring Program community trawls 1982 - 2008 by station and year  
with the number of times sampled per year and (the number of trawls per station)  
and additional trawls required for bioaccumulation 1999 - 2008*

Station/Year	1982	1983	1984	1985	1986	1987	1988	1989	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005	2006	2007	2008		
01	1(2)	2(2)	2(2)			3(2)	3(2)	3(2)	2(2)	2(2)	2(2)	1(2)	2(2)	2(2)	2(2)	See Appendix F-2	See Appendix F-2	1(1)	1(1)	1(1)	1(1)	1(1)	1(1)	1(1)	1(1)	1(1)	1(1)	1(1)	
02	1(2)	2(2)	2(2)				3(2)													1(1)	1(1)	1(1)	1(1)						
03	1(2)	2(2)	2(2)						2(2)	2(2)																			
04	1(2)	2(2)	2(2)			3(2)		3(2)			2(2)	1(2)	2(2)	2(2)	2(2)					1(1)	1(1)	1(1)	1(1)						
05	1(2)	2(2)	2(2)			3(2)	3(2)	3(2)	2(2)	2(2)																			
06	1(2)	2(2)	2(2)			3(2)	3(2)	3(2)			2(2)	1(2)	2(2)	2(2)	2(2)					1(1)	1(1)	1(1)	1(1)	1(1)	1(1)	1(1)	1(1)	1(1)	1(1)
07						3(2)	3(2)	3(2)																					
28																				1(1)	1(1)	1(1)	1(1)						
29						3(2)	3(2)	3(2)																					
30						3(2)	3(2)	3(2)	2(2)	2(2)																			
31											2(2)	1(2)	2(2)	2(2)	2(2)			1(1)	1(1)	1(1)	1(1)								
32																		1(1)	1(1)	1(1)	1(1)								
34																													
35																		1(1)	1(1)	1(1)	1(1)								
36																		1(1)	1(1)	1(1)	1(1)								
37																		1(1)	1(1)	1(1)	1(1)								
38																		1(1)	1(1)	1(1)	1(1)								
39																		1(1)	1(1)	1(1)	1(1)								
42																		1(1)	1(1)	1(1)	1(1)								
50																		1(1)	1(1)	1(1)	1(1)								
53																		1(1)	1(1)	1(1)	1(1)								
57																		1(1)	1(1)	1(1)	1(1)								
62																		1(1)	1(1)	1(1)	1(1)								
63																													
65																		1(1)	1(1)	1(1)	1(1)								
66																		1(1)	1(1)	1(1)	1(1)								
70																		1(1)	1(1)	1(1)	1(1)								
74																					1(1)								
Number of additional trawls required to obtain sufficient English sole for bioaccumulation analyses:																		12	0	3	12	11	10	21	6	8	11		

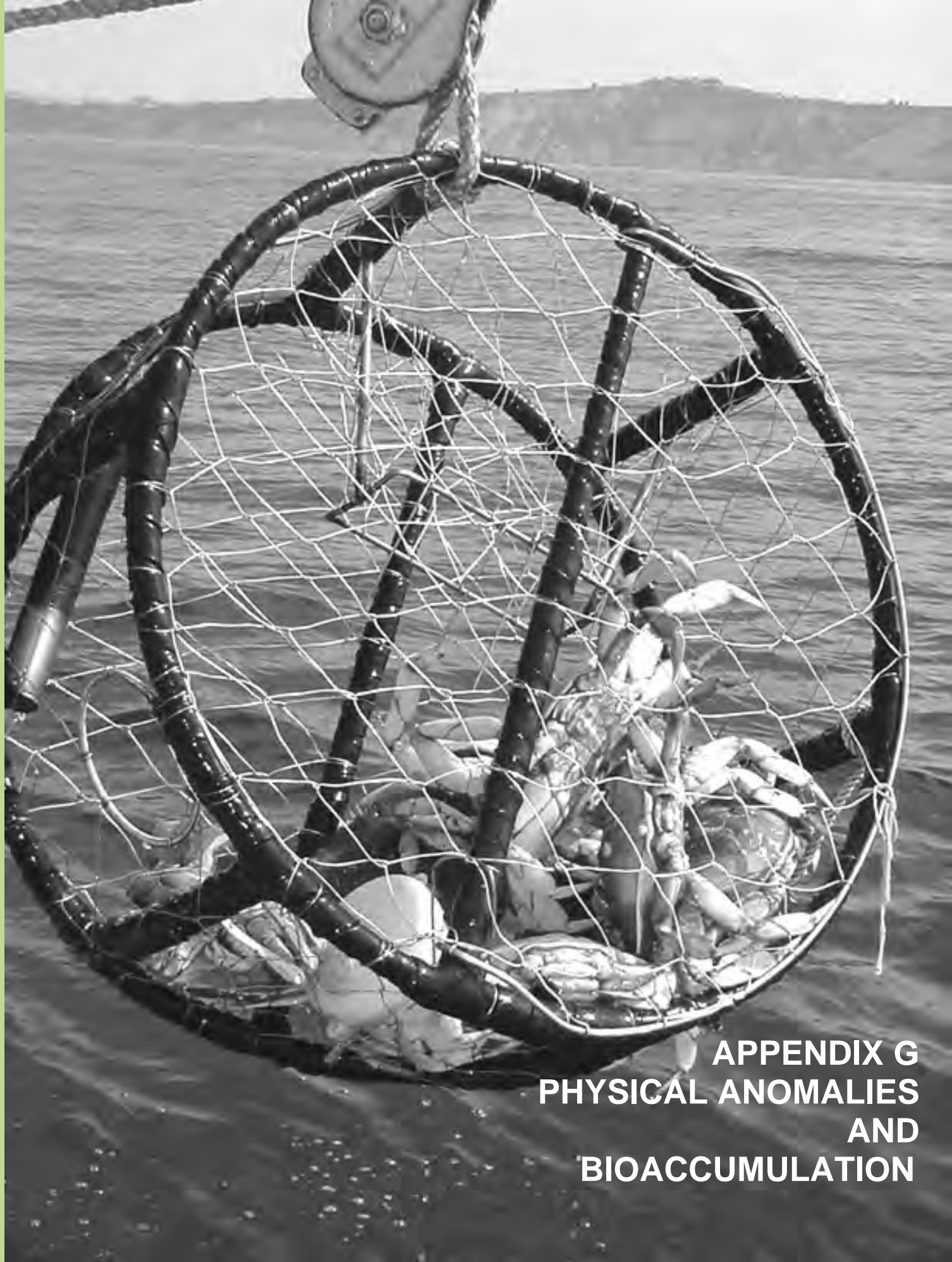
Appendix F-2

*SWOO Regional Monitoring Program community trawls 1997 - 1998 by station and year with the number of times sampled per year and (the number of trawls per station)*

*Stations sampled within strata were not at the same location year to year*

<b>Station</b>	<b>1997</b>	<b>1998</b>
A-1	1(1)	
A-2	1(1)	
A-3	1(1)	
A-4	1(1)	
B-1	1(1)	
B-2	1(1)	
B-3	1(1)	
B-4	1(1)	
C-1	1(1)	
C-2	1(1)	
C-3	1(1)	
C-4	1(1)	
D-1	1(1)	
D-2	1(1)	
D-3	1(1)	
D-4	1(1)	
E-1	1(1)	
E-2	1(1)	
E-3	1(1)	
E-4	1(1)	
F-1	1(1)	
F-2	1(1)	
F-3	1(1)	
F-4	1(1)	
G-1	1(1)	
G-2	1(1)	
G-3	1(1)	
G-4	1(1)	
H-1	1(1)	
H-2	1(1)	
H-3	1(1)	
H-4	1(1)	
A-1		1(1)
A-2		1(1)
A-3		1(1)
B-1		1(1)
B-2		1(1)
B-3		1(1)
D-1		1(1)
D-2		1(1)
D-3		1(1)
E-1		1(1)
E-2		1(1)
E-3		1(1)
E-4		1(1)
F-1		1(1)
F-2		1(1)
F-3		1(1)
G-1		1(1)
G-2		1(1)
G-3		1(1)
H-1		1(1)
H-2		1(1)





**APPENDIX G  
PHYSICAL ANOMALIES  
AND  
BIOACCUMULATION**

APPENDIX G  
PHYSICAL ANOMALIES AND BIOACCUMULATION,  
1997 to 2012

Appendix		Page
G-1	Characteristics of Dungeness crab used for bioaccumulation analyses, collected from Reference and Outfall areas, 1997 – 2012	G-2
G-2	Reporting and detection limits for organic compounds assessed 1997 – 2012	G-3
G-3	Organic compounds detected in tissues of crab from Reference and Outfall areas, 1997 – 2012	G-4
G-4	Detection limits (DL) for trace metals assessed 1997 – 2012	G-6
G-5	Trace metals detected in tissues of crab tissues of Dungeness crab from Reference and Outfall areas, 1997 – 2000	G-7

Appendix G-1

*Characteristics of Dungeness crab used for bioaccumulation analyses, collected from SWOO Reference and Outfall*

Parameter	Survey year																	
	1997	1998	1999	2000	2001	2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012		
Reference Area	Mean carapace width (mm)	168	ND	166	179	180	185	188	183	179	188	187	145	174	186	186	182	
	Mean weight/crab (g)	663	ND	608	745	734	938	870	750	740	859	759	456	677	801	803	791	
	% Lipid in muscle tissue	0.01	0.07	0.70	0.31	0.20	0.05	0.03	0.04	0.04	0.04	0.04	0.04	0.04	2.13	0.30	0.47	0.67
	% Lipid in hepatopancreas tissue average # organisms/replicate	4	5	27	18	14	14	23	24	20	20	20	20	20	14	9	9	22
Outfall Area	Mean carapace width (mm)	179	ND	176	185	182	196	190	184	186	189	188	139	175	184	190	185	
	Mean weight/crab (g)	768	ND	667	830	740	978	840	760	782	516	718	406	680	795	880	818	
	% Lipid in muscle tissue	0.03	0.08	0.40	0.27	0.25	0.06	0.03	0.05	0.05	0.04	0.05	0.05	2.24	0.55	0.40	0.91	
	% Lipid in hepatopancreas tissue average # organisms/replicate	3	4	26	12	15	12	19	20	18	18	18	18	18	12	5	14	25

ND = no data available

Lipids content analyses performed by one lab for years 1997-1998 and a different lab for subsequent years

## Appendix G-2

Reporting and detection (Det.) limits (ppb) for organic compounds assessed 1997-2012. (NA = not assessed)

	1997	1998	1999	2000	2001	2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012
	Reporting	Reporting	Det.	Det.	Det.	Det.	Det.	Det.	Det.	Det.	Det.	Det.	Det.	Det.	Det.	Det.
	Limit	Limit	Limit	Limit	Limit	Limit	Limit	Limit	Limit	Limit	Limit	Limit	Limit	Limit	Limit	Limit
4,4'-DDE	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
4,4'-DDD	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
4,4'-DDT	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
Naphthalene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Acenaphthylene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Acenaphthene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Fluorene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Phenanthrene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Anthracene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Fluoranthene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Pyrene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Benzo[a]anthracene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Chrysene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Benzo[b]fluoranthene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Benzo[k]fluoranthene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Benzo[e]pyrene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Perylene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Benzo[a]pyrene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Indeno[1,2,3-cd]pyrene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Dibenzo[a,h]anthracene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
Benzo[ghi]perylene	10	10	10	2	2	2	2	2	2	2	2	2	2	2	2	2
2,4'-Dichlorobiphenyl PCB 008	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',5'-Trichlorobiphenyl PCB 018	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,4,4'-Trichlorobiphenyl PCB 028	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',5,5'-Tetrachlorobiphenyl PCB 052	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,5'-Tetrachlorobiphenyl PCB 044	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,3',4,4'-Tetrachlorobiphenyl PCB 066	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',4,5,5'-Pentachlorobiphenyl PCB 101	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
3,4,4',5'-Tetrachlorobiphenyl PCB 081	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	2	2	2	2	2	2
3,3',4,4'-Tetrachlorobiphenyl PCB 077	NA	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2',3,4,4',5'-Pentachlorobiphenyl PCB 123	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	2	2	2	2	2	2
2,3',4,4',5'-Pentachlorobiphenyl PCB 118	10	10	NA	2	NA	NA	NA	NA	NA	2	2	2	2	2	2	2
2,2',4,4',5,5'-Hexachlorobiphenyl PCB 153	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
3,3',4,5,5'-Pentachlorobiphenyl PCB 127	NA	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,3,3',4,4'-Pentachlorobiphenyl PCB 105	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,4,4',5'-Hexachlorobiphenyl PCB 137	NA	NA	4	2	2	2	2	2	2	2	2	2	2	2	2	2
3,3',4,4',5'-Pentachlorobiphenyl PCB 126	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,4',5,5',6-Heptachlorobiphenyl PCB 187	NA	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,3',4,4'-Hexachlorobiphenyl PCB 128	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,3,3',4,4',5'-Hexachlorobiphenyl PCB 157	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	2	2	2	2	2	2
2,2',3,4,4',5,5'-Heptachlorobiphenyl PCB 180	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,3',4,4',5-Heptachlorobiphenyl PCB 170	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,3,3',4,4',5,5'-Heptachlorobiphenyl PCB 189	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	2	2	2	2	2	2
2,2',3,3',4,4',5,6-Octachlorobiphenyl PCB 195	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,3',4,4',5,5',6-Nonachlorobiphenyl PCB 206	NA	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
Decachlorobiphenyl PCB 209	NA	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,4',5-Trichlorobiphenyl PCB 31	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2',3,4-Trichlorobiphenyl PCB 33	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',4',5-Tetrachlorobiphenyl PCB 49	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,4,4',5-Tetrachlorobiphenyl PCB 74	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,3',4',5-Tetrachlorobiphenyl PCB 70	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,5',6-Pentachlorobiphenyl PCB 95	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,3,3',4'-Tetrachlorobiphenyl PCB 56	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,3,4,4'-Tetrachlorobiphenyl PCB 60	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',4,4',5-Pentachlorobiphenyl PCB 99	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3',4,5-Pentachlorobiphenyl PCB 97	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,4,5-Pentachlorobiphenyl PCB 87	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,3,3',4',6-Pentachlorobiphenyl PCB 110	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,3,4,4',5-Pentachlorobiphenyl PCB 114	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	2	2	2	2	2	2
2,2',3,5,5',6-Hexachlorobiphenyl PCB 151	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,4,5',6-Hexachlorobiphenyl PCB 149	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,3',4,6'-Hexachlorobiphenyl PCB 132	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,4,5,5'-Hexachlorobiphenyl PCB 141	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,3,3',4,4',6-Hexachlorobiphenyl PCB 158	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,4,4',5',6-Heptachlorobiphenyl PCB 183	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,3',4,4',5,5'-Hexachlorobiphenyl PCB 167	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	2	2	2	2	2
2,2',3,3',4,5,6-Heptachlorobiphenyl PCB 174	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,3',4',5,6-Heptachlorobiphenyl PCB 177	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,3,3',4,4',5'-Hexachlorobiphenyl PCB 156	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
3,3',4,4',5,5'-Hexachlorobiphenyl PCB 169	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	2	2	2	2	2	2
2,2',3,3',4,4',5,5',6-Octachlorobiphenyl PCB 201	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,4,4',5,5',6-Octachlorobiphenyl PCB 203	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,3',4,4',5,5'-Octachlorobiphenyl PCB 194	10	10	4	2	2	2	2	2	2	2	2	2	2	2	2	2
2,2',3,4,4',5-Hexachlorobiphenyl PCB 138	10	10	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA





Appendix G-4

*Detection limits (DL) for trace metals in tissues, 1997-2012*

	1997 (µg/g)	1998 (µg/g)	1999 (µg/g)	2000 (µg/g)	2001 (µg/g)	2002 (µg/g)	2003 (µg/g)	2004 (µg/g)	2005 (µg/g)
Silver (Ag)	0.081	0.025	0.03	0.025	0.02	0.015	0.030	0.030	0.010
Aluminum (Al)	1.04	0.5	0.50	0.500	0.3	0.26	0.300	0.300	0.200
Arsenic (As)	0.49	1	1.00	1.00	1	0.5	0.500	0.500	0.100
Cadmium (Cd)	0.037	0.01	0.10	0.100	0.1	0.05	0.100	0.100	0.010
Chromium (Cr)	0.22	0.1	0.10	0.100	0.1	0.05	0.100	0.100	0.100
Copper (Cu)	0.088	0.1	0.10	0.100	0.3	0.05	0.100	0.100	0.100
Iron (Fe)	0.43	0.5	0.50	0.500	0.2	0.26	0.500	0.500	0.200
Mercury (Hg)	0.020	0.02	0.02	0.0200	0.005	0.010	0.010	0.010	0.020
Manganese (Mn)	0.016	0.1	0.10	0.100	0.4	0.05	0.100	0.100	0.050
Nickel (Ni)	0.031	0.035	0.04	0.040	0.03	0.10	0.100	0.100	0.200
Lead (Pb)	0.23	0.09	0.100	0.100	0.3	0.10	0.200	0.200	0.100
Selenium (Se)	0.92	1	1.00	1.000	0.03	0.51	0.500	0.500	2.000
Zinc (Zn)	0.31	0.1	0.10	0.100	0.1	0.05	0.100	0.100	0.100

	2006 (mg/kg)	2007 (mg/Kg )	2008 (mg/Kg)	2009 (mg/Kg)	2010 (mg/Kg)	2011 (mg/Kg)	2012 (mg/Kg)
Silver (Ag)	0.008	0.001	0.002	0.001	0.007	0.006	0.003
Aluminum (Al)	0.15	0.060	0.76	3.9	2.5	2.12	16
Arsenic (As)	5.52	0.03	0.10	0.05	0.17	0.17	0.08
Cadmium (Cd)	0.070	0.005	0.007	0.01	0.015	0.09	0.04
Chromium (Cr)	0.58	0.003	0.08	0.02	0.01	0.13	0.06
Copper (Cu)	0.110	0.051	0.030	0.06	0.07	0.43	0.5
Iron (Fe)	0.15	0.60	0.76	0.325	2.5	2.13	2
Mercury (Hg)	0.016	0.003	0.018	0.015	0.02	0.013	0.006
Manganese (Mn)	0.150	0.009	0.023	0.16	0.07	0.06	0.5
Nickel (Ni)	0.480	0.003	0.005	0.03	0.05	0.17	0.08
Lead (Pb)	0.14	0.01	0.02	0.02	0.05	0.11	0.05
Selenium (Se)	3.72	0.02	0.05	0.01	0.07	0.06	0.03
Zinc (Zn)	0.34	0.08	0.76	0.6	0.51	1.28	0.6

Appendix G-5

Mean concentrations (ppm, dry weight) of trace metals detected in tissues of Dungeness

	Crab Muscle							
	Reference area				Outfall area			
	1997	1998	1999	2000	1997	1998	1999	2000
Silver	NS	0.76	0.76	0.51	NS	3.30	1.07	0.58
Aluminum	NS	36.1	43.8	20.2	NS	7.6	13.8	6.7
Arsenic	NS	92.0	84.6	76.4	NS	91.5	91.0	94.4
Cadmium	NS	0.03	0.26	0.14	NS	0.06	< 0.10	< 0.10
Chromium	NS	0.10	1.38	0.30	NS	0.10	1.12	0.28
Copper	NS	43.0	46.3	29.3	NS	41.5	43.7	<b>37.1</b>
Iron	NS	17.3	52.7	24.3	NS	19.3	32.9	21.7
Mercury	NS	0.21	0.63	0.50	NS	0.25	0.67	0.51
Manganese	NS	0.7	1.4	0.8	NS	0.7	1.1	0.8
Nickel	NS	0.09	0.18	0.10	NS	0.19	0.15	0.07
Lead	NS	0.19	0.20	< 0.10	NS	0.11	0.19	0.12
Selenium	NS	3.4	4.7	3.2	NS	3.3	4.4	4.0
Zinc	NS	184	200	163	NS	176	184	168

	Crab Hepatopancreas							
	Reference area				Outfall area			
	1997	1998	1999	2000	1997	1998	1999	2000
Silver	NS	4.02	10.70	9.45	NS	6.09	11.82	6.40
Aluminum	NS	11.9	23.0	22.9	NS	2.5	23.2	10.2
Arsenic	NS	86.9	115.2	136.1	NS	112.1	83.8	121.3
Cadmium	NS	82.50	86.48	103.20	NS	79.30	75.75	101.33
Chromium	NS	0.15	0.29	0.29	NS	0.11	0.34	0.33
Copper	NS	170.2	315.8	395.6	NS	227.5	412.6	291.0
Iron	NS	255.9	262.2	321.9	NS	247.3	296.4	395.1
Mercury	NS	0.26	0.43	0.54	NS	0.28	0.73	0.58
Manganese	NS	6.0	8.9	7.4	NS	7.0	7.8	9.0
Nickel	NS	4.03	2.33	3.06	NS	3.08	3.78	4.30
Lead	NS	0.18	0.20	0.25	NS	0.15	0.32	0.29
Selenium	NS	11.8	13.9	17.1	NS	13.3	14.5	20.1
Zinc	NS	159	128	198	NS	151	152	320

NS = No sample data

**Bold font** indicates statistically significantly higher than corresponding tissue at other area; one-tailed Student's T-test, unequal variance,  $\alpha = 0.05$ .